

Article

Vacant Bivalve Boreholes Increase Invertebrate Species Richness in a Physically Harsh, Low Intertidal Platform

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Abstract: Ecosystem engineers can modulate harsh abiotic conditions, thus creating habitat for species that cannot withstand the local environment. In this study, we investigated if vacant boreholes created by the rock-boring bivalve *Petricola dactylus* increase species richness in the low intertidal zone of a Patagonian rocky shore characterized by intense hydrodynamic forcing and sediment scour. Invertebrate species richness was three times higher in engineered than unengineered habitats (i.e., with and without *Petricola* boreholes, respectively) and the increase in species richness was area-independent. The most prevalent species in unengineered areas showed strong adhesion mechanisms, whereas infaunal and vagile species were mostly restricted to boreholes. The positive influence of engineered microhabitats on species richness can largely be attributed to amelioration of physical conditions, particularly a reduction in hydrodynamic forces and sediment trapping/stabilization within boreholes. We conclude that vacant boreholes are essential microhabitats for the maintenance of biodiversity within the otherwise inhospitable low intertidal zone.

Keywords: ecosystem engineer; rock-boring; boreholes; harsh environment; low intertidal; rocky shores; increase species richness; *Petricola dactylus*

1. Introduction

Local abiotic conditions are often suitable to a fraction of the species in the broader species pool (e.g., [1,2]). In these cases, biodiversity is frequently enhanced by species that structurally modify the environment and, in so doing, create microhabitats where the impact of abiotic conditions becomes reduced (e.g., [3–5]). These physical ecosystem engineers (sensu [6,7]) can create favorable microhabitats for other organisms either via their own structures (e.g., cushion plants that retain moisture and modulate extreme soil temperatures to other plants in otherwise unsuitable alpine environments [3]) or their activities (e.g., rabbits that excavate burrows, and thus create thermoregulatory retreats for lizards in thermally variable and otherwise inhospitable Mediterranean

pastures [8]). Oftentimes, many species in a locale occur in strict association with the microhabitats engineered by one or a few species (e.g., [3–5]).

Physically engineered microhabitats are known to be critical for biodiversity in rocky intertidal shores (e.g., [4,5,9–11]), where wave exposure, desiccation, and temperature often exceed species tolerance limits [12–14]. Therefore, many of the species inhabiting this environment rely on the creation of protected microhabitat by a few tolerant and often spatially dominant sessile species (e.g., mussels, tunicates, seaweeds [4,5,10,11,15]). Typically, the facilitative effects of physically engineered microhabitats on other species have been reported to occur at high intertidal elevations, where the impacts of desiccation and temperature variations are expected to be higher (e.g., [9,16]). In contrast, the lower intertidal zone has been generally considered as a physically “benign” habitat where biodiversity is usually controlled by biotic factors (predation and space competition [17–19]) and physical habitat amelioration by ecosystem engineers becomes less important [9,16,20].

Here we evaluate if the microhabitats created by a physical ecosystem engineer—the rock-boring bivalve, *Petricola dactylus*—positively contribute to species richness in the physically harsh, low intertidal zone of a southwestern Atlantic rocky shore. *Petricola dactylus* is a mechanical borer that inhabits soft rock in low intertidal and subtidal habitats of the southwestern Atlantic and southeastern Pacific, from Uruguay to Southern Chile [21–23]. The boreholes left vacant by dead *P. dactylus* reach high densities at our study site (often more than 600 boreholes m^{-2}) and are prominent structural elements in otherwise barren rock surfaces (Figure 1b). Since the microhabitat within boreholes is protected from extreme hydrodynamic forces and scour and, as a consequence, retains sediments (i.e., a critical habitat element for the infauna, Figure 1c), we predicted that it will positively contribute to the overall invertebrate species richness of the low intertidal zone and, to an extent, an area larger than what it adds to the seabed. To test this prediction, we compared invertebrate species richness in patches with and without *P. dactylus* boreholes, and evaluated if differences in species richness are proportional to the amount of surface area sampled in each patch type. Additionally, we tested if the species occurring in strict association to boreholes are a subset of those occurring in mussel beds at midshore elevations.

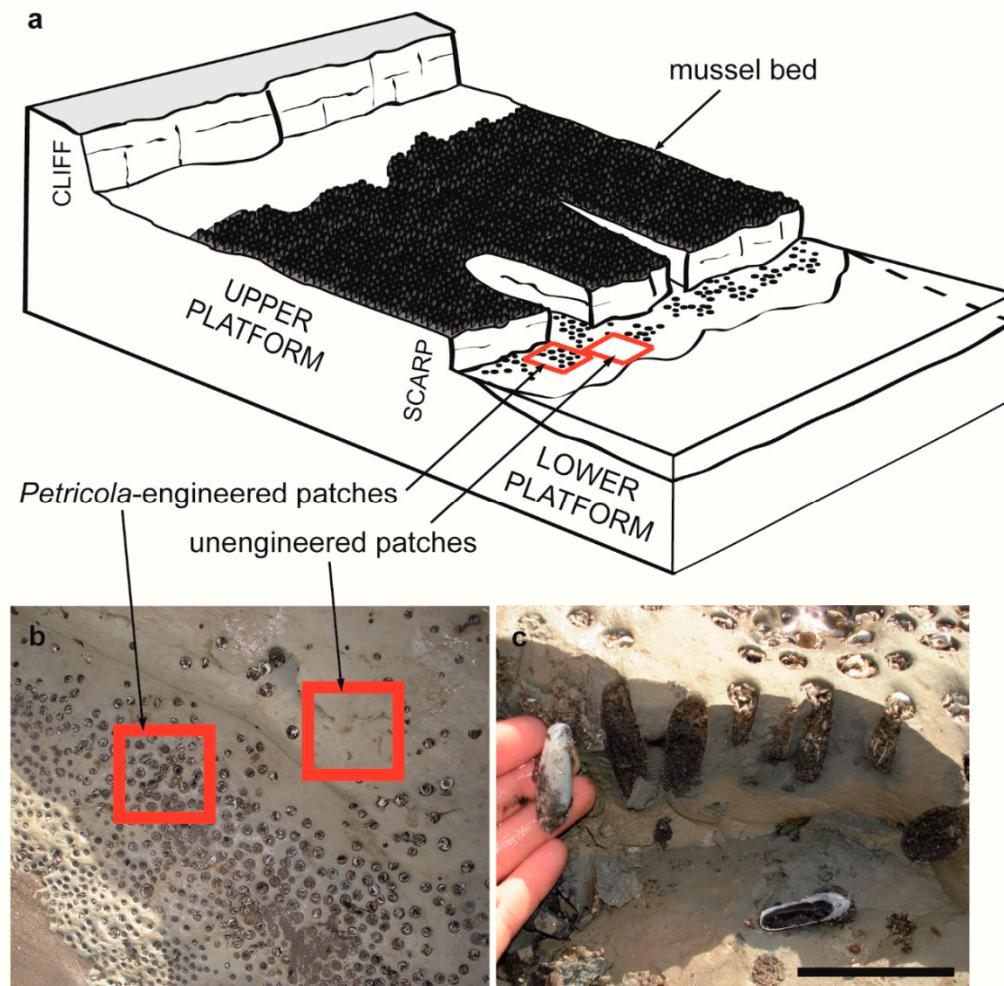


Figure 1. (a) Scheme of the horizontal shore platform at El Espigón, showing upper and lower platforms separated by a scarp. Arrows show the position of the mussel beds, *Petricola*-engineered patches (with boreholes) and unengineered patches (without burrows). (b) Close view of *Petricola*-engineered and unengineered patches. The red squares are scaled, 10 cm side quadrats as those used for sampling. (c) Lateral view of boreholes filled with sediment and empty *Petricola dactylus* shells (the rock was cut perpendicular to the surface). Scale is 5 cm.

2. Materials and Methods

2.1. Study Site and Organisms

Samplings were conducted in a rocky platform at El Espigón, Río Negro Province, Argentina ($41^{\circ}07'$ S, $63^{\circ}00'$ W), between February 2010 and May 2018. This is an extensive (600 m width) horizontal or Type B platform (see [24]) showing a vertical seaward scarp (ca. 2 m high) that sets the limit between the upper and lower intertidal levels (hereafter upper and lower platforms, Figure 1a). This platform lies at the base of a high, active cliff (up to 65 m height [25]) and both the cliff and the platform are composed of fine- to midgrained sandstones and limestone, corresponding to the geological unit locally known as Río Negro Formation [26]. Tides are semidiurnal and macrotidal (4.33 m max. amplitude; Servicio de Hidrografía Naval, Argentina, www.hidro.gov.ar).

The upper platform is characterized by extensive mussel beds (mean densities vary between 36,000 and 55,000 ind. m^{-2} , [27]). The dominant mussel species in this region is *Brachidontes rodriguezii*, though scattered individuals of a morphologically similar species, *Perumytilus purpuratus* (also referred to in the literature as *Brachidontes purpuratus*), are often found within *B. rodriguezii* beds (both species coexist

between 41° and 43° S; see [28–30]). These mussels are important ecosystem engineers in southwestern Atlantic rocky intertidal shores, as their beds provide critical habitats to other organisms [4,5,11,28].

The lower platform, on the other hand, is essentially flat and barren (Figure S1). Apart from scattered rock fractures, the only structural elements at this shore level are the perforations produced by *Petricola dactylus*. Most of these boreholes are structural legacies resulting from *P. dactylus* death (just 2% of boreholes are inhabited by living specimens). They occur aggregated into dense patches ($630 \pm 329 \text{ ind.m}^{-2}$; Figure 1b) and are often occupied by individual mussels (although mussels are generally smaller than those in the upper platform, pers. obs.) and other invertebrates as well. This zone is characterized by intense wave action (the scarp induces local wave energy dissipation and reflection; see [31,32]) and scour due to sand and gravel transport (Figure S1). This shore level becomes exposed only during very low (<0.60 m) tides (mean low tide level is 0.70 m; Servicio de Hidrografía Naval) and can be temporarily covered by sand and/or gravel (Figure S1, pers. obs.). Hydrodynamic forcing and sediment scour are the only logical alternatives to explain the predominance of bare rock surfaces in the low platform (see also Figure S1). Indeed, desiccation is expected to be insignificant at this shore level since emersion takes place 8–21 days per month (mean 15.25 ± 4.11) but for short periods (less than 2 h) that do not necessarily coincide with desiccative conditions (e.g., daytime, high irradiation, low ambient moisture). Grazers and predators are also unexpected to produce bare rock surfaces in this area as the removal of dominant sessile species by these consumers should still allow colonization by competitively inferior, consumer-resistant, and/or nonfood species (e.g., [33,34]).

2.2. Sampling

To evaluate if boreholes created by *P. dactylus* can increase species richness, we measured macroinvertebrate species presence in engineered (i.e., with boreholes) and unengineered (i.e., without boreholes) patches at the lower platform. Engineered and unengineered patches were randomly sampled during 2010 (February, May, July, September, and November), 2011 (March and June), and 2018 (May) using $10 \times 10 \text{ cm}$ quadrats ($n = 65$, between 5 and 25 samples per patch type per date). The unengineered patches sampled here were located amongst boreholes (see Figure 1b) and samples from each patch type came from the same tidal level and position on the shore. The number of samples taken differed between sampling dates because the time available for sampling varied with the duration of the tidal emersion period. Nonetheless, the number of samples taken from engineered and unengineered patches was the same in each sampling date. Samples from engineered patches included invertebrates inside boreholes and those occurring on the rock surface between neighboring boreholes. A laboratory spatula was used to remove the sediments and invertebrates from inside the boreholes. Invertebrates visible on the rock surface or inside the boreholes were either identified in situ or collected and preserved in 96% ethanol for subsequent identification in the laboratory. The same methodology was applied for the invertebrates on the rock surface in the unengineered habitat. The number of boreholes and their diameter were registered in all the samples taken from engineered patches.

Additionally, we sampled the mussel beds at the upper platform to compare their invertebrate species composition with that of *Petricola*-engineered areas. Mussels and their associated sediments were sampled in 2009 (September), 2010 (February, May, July, and November), and 2011 (March, June, and August) using cylindrical cores (10 cm diameter, 8 replicates per date, $n = 64$). These samples were preserved in 96% ethanol, transported to the laboratory, and sieved (500 μm mesh) to separate and identify the macrofauna.

2.3. Data Analysis

Species richness (S_{obs}) in *Petricola*-engineered and unengineered habitat at the lower platform was quantified with sample-based species accumulation curves constructed by means of rarefaction techniques [35]. Species accumulation was first evaluated as a function of the number of samples collected in each habitat type ($n = 65$). Then, species accumulation was evaluated as a function of

the rock surface area of boreholes in *Petricola*-engineered patches and compared with areal species accumulation in unengineered rock patches. Invertebrate samples were standardized to borehole surface area as follows:

(a) The rock surface area added by individual boreholes was estimated as the surface area of a cylinder, based on field measurements of borehole diameter (see above) and assuming a uniform borehole depth of 4 cm. Such a depth is about the maximum depth observed for boreholes at our study site (see also Figure 1b), and thus provides an estimate of the maximum rock surface area potentially added by boreholes (i.e., most likely an overestimate of the actual surface area).

(b) Data from individual boreholes (3 to 11) were grouped adding to an estimated maximum surface area of 100 cm² per group (76 groups, mean = 99.8 cm², SD = 7.8), thus matching the surface area of unengineered rock samples (i.e., 10 cm side quadrats). Invertebrate species composition was established at the group level. Groups were used as grain (areal units of observation; see [36]) when constructing accumulation curves for the species sampled from boreholes. As the surface area of boreholes was likely overestimated here (see (a)), our estimates of species accumulation per surface area unit can be considered conservative.

Confidence intervals (95%) for Sobs were calculated based on the unconditional variance estimate developed by Colwell et al. [37]. Nonoverlapping confidence intervals were considered to indicate significant differences between pairs of Sobs values. This approach is normally used as a simple but conservative criterion of statistical difference in the absence of any computationally practical standard test to compare Sobs values [37,38]. Moreover, a *t*-test for unequal variances (or “Welch’s approximate *t*”, [39]) was used to compare the mean number of species per sample in engineered and unengineered patches and in boreholes versus unengineered rock surface area. To evaluate whether our sampling effort was enough to obtain reliable estimates of species richness in each habitat type (*Petricola*-engineered, unengineered, and mussel beds), we compared Sobs values with the Chao 2 estimate of the total number of species likely to be observed in the community [3,40]. EstimateS 9.1.0 [38] was used to construct rarefied species accumulation curves and to estimate Chao 2.

Differences in the identity of species associated with the two engineers (*Petricola dactylus* at the lower platform and mussels at the upper platform) were quantified as the paired overlap (PO [41], see also [5]). PO was calculated for the original data matrix and for 1000 randomized matrices of the same size. The species in habitat j (*Petricola*-engineered) were considered a significant subset of those in habitat i (mussel-engineered) if the observed PO_{ij} lay within the upper 95% percentile of the frequency distribution of PO_{ij} values of the randomized matrices. Randomized matrices and their associated PO values were obtained using the software ANINHADO [42].

3. Results

3.1. Invertebrate Species Composition and Richness in SAMPLEs from *Petricola*-Engineered and Unengineered Patches

Twenty-one species were found at the lower platform (Table 1). Fourteen species from eight different taxonomic groups were exclusively found in *Petricola*-engineered patches. In contrast, only one species, the chiton *Plaxiphora aurata*, was exclusively found in unengineered patches. Six species were shared between these two habitat types (Table 1).

Table 1. Species found at the lower platform in unengineered and *Petricola*-engineered patches and at the upper platform in mussel-engineered patches (i.e., mussel beds).

Species	Taxa ¹	Adhesion ²	Unengineered	<i>Petricola</i> -Engineered	Mussel-Engineered
<i>Parabunodactys imperfecta</i>	CnA	YES	X	X	
Acotylea indet.	P	NO		X	
<i>Lumbrineris sp.</i>	AP	NO		X	X
<i>Perinereis sp.</i>	AP	NO		X	X
<i>Syllis gracilis</i>	AP	NO		X	X
<i>Syllis prolixa</i>	AP	NO		X	X
Cirratulidae	AP	NO		X	
<i>Protoarcinella uncinata</i>	AP	NO		X	X
<i>Platynereis sp.</i>	AP	NO			X
<i>Boccardia polybranchia</i>	AP	NO			X
<i>Capitella capitata?</i>	AP	NO			X
Nereididae (<i>Allita?</i>)	AP	NO			X
<i>Brachidontes rodriguezii</i>	MB	YES	X	X	X
<i>Petricola lapicida</i>	MB	NO		X	
<i>Mytilus edulis</i>	MB	YES			X
<i>Plaxiphora aurata</i>	MP	YES	X		
<i>Siphonaria lessoni</i>	MG	YES			X
<i>Themiste alutacea</i>	S	NO		X	
Nemertea indet.	N	NO		X	X
Chironomidae indet.	ID	YES	X	X	X
<i>Balanus glandula</i>	CrCi	YES	X	X	X
Copepoda indet.	CrCo	NO		X	
<i>Hyale grandicornis</i>	CrA	NO			X
Melitidae indet.	CrA	NO			X
Ampithoidae indet.	CrA	NO			X
Corophiidae indet.	CrA	NO	X	X	X
<i>Exosphaeroma sp.</i>	CrI	NO	X	X	X
<i>Tanais sp.</i>	CrT	NO		X	X
<i>Cyrtograpsus altimanus</i>	CrD	NO		X	X
<i>Platyxanthus crenulatus</i>	CrD	NO		X	
Total			7	20	22

¹ Taxa—AP: Annelida Polychaeta, CnA: Cnidaria Anthozoa, CrA: Crustacea Amphipoda, CrCi: Crustacea Cirripedia, CrCo: Crustacea Copepoda, CrD: Crustacea Decapoda, CrI: Crustacea Isopoda, CrT: Crustacea Tanaidacea, ID: Insecta Diptera, MB: Mollusca Bivalvia, MG: Mollusca Gastropoda, MP: Mollusca Polyplacophora, N: Nemertea, P: Platyhelminthes, S: Sipunculida. ² Adhesion—YES: Species that permanently or temporarily attach to the rock by means of secretions, NO: species showing no attachment.

The species accumulation curve obtained for the unengineered habitat approached the asymptote and observed species richness values (*Sobs*) were close to the Chao 2 richness estimator, which suggests that this habitat was sampled thoroughly enough to characterize its species richness (Figure 2a). On the contrary, the species accumulation curve corresponding to *Petricola*-engineered patches did not reach the asymptote and *Sobs* values were lower than the Chao 2 estimate (Figure 2a), which suggests additional species should be found with further sampling. *Sobs* was significantly higher in the *Petricola*-engineered than unengineered patches (no overlap between confidence intervals; see Figure 2a). The mean number of species per sample was higher in engineered patches (mean 1.92 ± 2.38) than in unengineered ones (0.38 ± 0.49) (Welch's *t*-test: $t = 7.31$, $df = 89$, $P < 0.001$).

Fifty-six out of 65 sampling quadrats in the *Petricola*-engineered patches (86%) showed at least one invertebrate species. In contrast, invertebrates were found just in 17 out of the 65 quadrats from unengineered patches (26%). The most prevalent species in the unengineered habitat were the mussel *Brachidontes rodriguezii*, the barnacle *Balanus glandula*, and the chiton *Plaxiphora aurata* (15, 9, and 5% of samples, respectively; Figure 3). *B. rodriguezii* was also the most prevalent species in *Petricola*-engineered patches (78% of samples), followed by the polychaetes *Lumbrineris sp.*, *Syllis gracilis*, and *S. prolixa* (26, 23, and 14% of samples, respectively; Figure 3). Nine of the species sampled in *Petricola*-engineered patches were found only one time (in just one quadrat) during the sampled period.

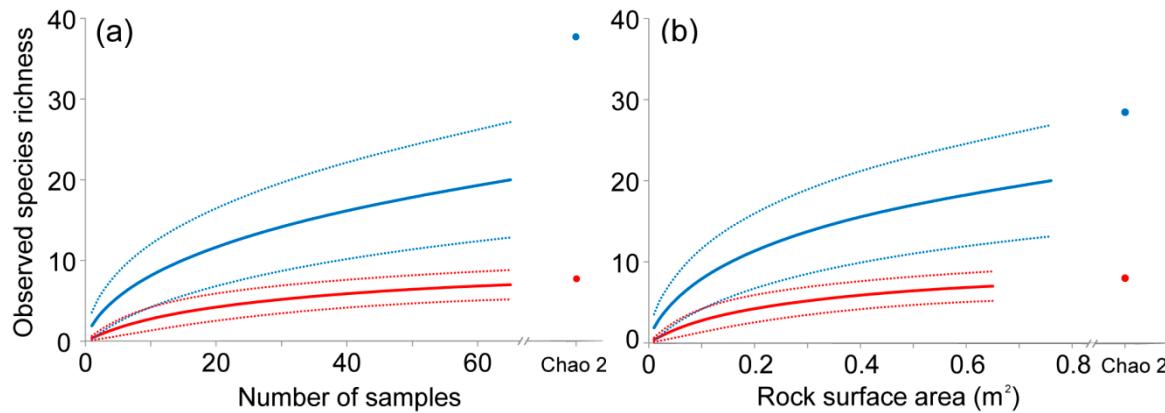


Figure 2. (a) Species accumulation curves (continuous lines) for *Petricola*-engineered (blue) and unengineered (red) patches at the lower platform with their respective 95% confidence intervals (dashed lines). The observed species richness (*S*_{obs}) for 65 samples is shown for both habitats together with the values of the Chao 2 estimate of total species richness (blue and red circles for *Petricola*-engineered and unengineered patches, respectively). (b) Species accumulation curves and their confidence intervals for species present inside *Petricola* boreholes in engineered patches (blue) and on the rock surface in unengineered patches (red) in relation to the rock surface area. The curve for boreholes extends further than that of unengineered rock surfaces because of differences in surface area sampled (boreholes = 0.76 m⁻², unengineered rock surface = 0.65 m⁻²).

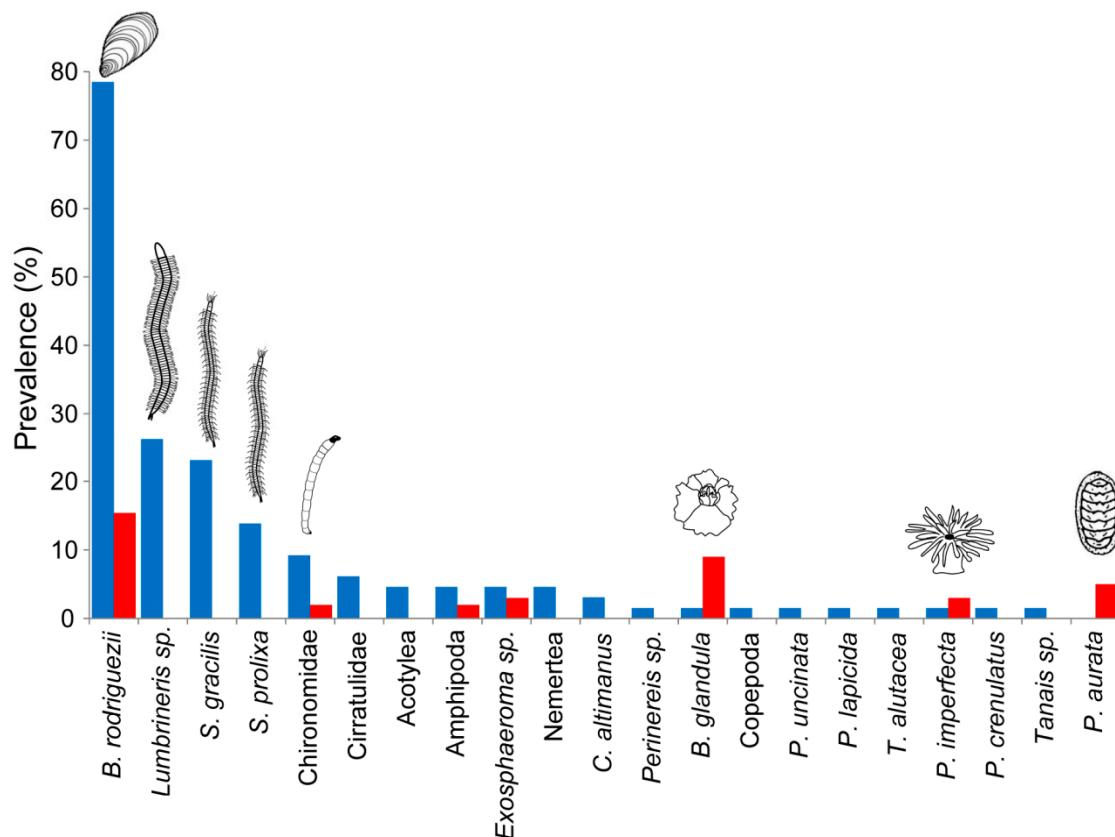


Figure 3. Prevalence (% of samples with species present) of invertebrate species in *Petricola*-engineered (blue bars) and unengineered (red bars) patches. The most frequent species in the two different habitats are represented by an icon (see text in Results).

3.2. Invertebrate Species Richness as a Function of Sampled Area

We found 416 boreholes in the 65 sample quadrats from *Petricola*-engineered patches (mean = 6.38 ind. 100 cm^{-2} , SD = 3.29). Their aggregate surface area was estimated to be 0.76 m^2 (i.e., assuming a constant borehole depth of 4 cm; see Methods). The rock surface area between boreholes (i.e., total quadrat area minus the area covered by borehole openings) was estimated to be 0.6 m^2 . Based on these estimates, the total rock surface area sampled in *Petricola*-engineered patches was 1.36 m^2 ($0.6 + 0.76$). In the absence of boreholes, the total area sampled would be 0.65 m^2 (i.e., 65 quadrats, 10 cm side each). This indicates that *Petricola* boreholes cause a 117% increase in rock surface area.

The accumulation curves relating observed species richness to rock surface area indicate that, for an equal surface area of 0.65 m^2 (i.e., the total area sampled in unengineered patches), species richness is higher inside boreholes than on the rock surface of unengineered areas (Figure 2b, no overlap between confidence intervals at 0.65 m^2 rock surface).

The mean number of species per 100 cm^2 surface area was higher inside boreholes (mean 1.96 ± 1.79) than in the unengineered rock surface (0.38 ± 0.49) (Welch's *t*-test: $t = 8.92$, $df = 117$, $P < 0.001$).

3.3. Invertebrate Species Composition in *Petricola*-Engineered Patches and Mussel Beds

Twenty-two invertebrate species were found in the mussel beds samples from the upper platform (Table 1; the species accumulation curve for this habitat was asymptotic, suggesting sufficient sampling, see Supporting Materials, Figure S2: Species accumulation curve in mussel beds). Thirteen of the species in the mussel beds were also found in the *Petricola*-engineered habitat while the remaining nine species were exclusive of the mussel habitat (Table 1). Partial overlap (PO_{ij}) in the observed data was thus 65. This value lay below the upper 5% percentile of PO_{ij} values of the randomized matrices (i.e., 85), which indicates that the species associated to the *Petricola*-engineered habitat are not a significant subset of those occurring in the mussel beds ($P = 0.96$).

4. Discussion

Our results show a threefold increase in the species richness of the low intertidal zone associated with the presence of boreholes made by *Petricola dactylus*. Yet, this is likely an underestimate provided that our species accumulation curves suggest that further sampling would lead to additional species records in the *Petricola*-engineered habitat but not in the unengineered one (see Figure 2a). This indicates that vacant *P. dactylus* boreholes are essential microhabitats for the maintenance of biodiversity within the otherwise inhospitable low platform.

The increased species richness of *Petricola*-engineered patches cannot simply be explained as an effect of the surface area added by boreholes. Indeed, species richness was also higher in engineered patches than unengineered ones when analyzed as a function of the surface area sampled (see Figure 2b). This indicates that the species richness of engineered patches is controlled by favorable environmental conditions within boreholes rather than increased rock surface area.

Vacant bivalve boreholes in intertidal platforms can protect other organisms from the impacts of desiccation, hydrodynamic forces, and predators, and also retain sediments that may be critical for the infauna [5,43,44]. Here, we anticipated that hydrodynamic forces and sediment scour should limit invertebrate colonization in the low platform (cf., unlikely importance of desiccation and predators; see Study Site and Organisms). In addition, we postulated that the retention of sediment within boreholes should favor infaunal colonization (see Introduction). In agreement, our results suggest that the increased richness of *Petricola*-engineered patches results from reduced water flow and sediment trapping/stabilization within boreholes. This is supported by the fact that only species with strong adhesion mechanisms (barnacles, chitons, sea anemones) were frequent on unengineered rock surfaces while vagile species poor in adhesion mechanisms and, therefore, prone to being washed out by waves (e.g., polychaetes, nemerteans, small crustaceans) were restricted to boreholes in most cases (see Table 1

and Figure 3). Furthermore, the species restricted to boreholes included an array of infaunal and/or deposit-feeding taxa (e.g., polychaetes, tanaidaceans, chironomid larvae; see Table 1), which rely on stable sediments as habitat and/or food. In the absence of boreholes, infaunal and/or weakly adhesive vagile species would not withstand the harsh hydrodynamic regime and the concomitant sediment instability and scour that characterize the low platform.

The species in *Petricola*-engineered habitat are not a significant subset of those inhabiting mussel beds in the upper platform. Although both habitats share some infaunal (e.g., *Lumbrineris* sp., *Syllis* spp., *Tanais* sp., nemerteans) or interstitial species (e.g., *Cyrtograpsus altimanus*), several other species were exclusively found in *Petricola*-engineered patches (see Table 1). Therefore, *Petricola*-engineered patches extend the distribution of upper platform species into the lower platform but also promote the presence of species that do not occur in the former. These species are probably not adapted to the longer emersion periods that characterize the upper platform and are therefore limited to lower intertidal levels. By facilitating these species, *Petricola dactylus* could be increasing the overall species richness of the intertidal zone at this site.

By creating protected microhabitats within a harsh physical setting, physical ecosystem engineers can increase the extent of suitable habitat for other species (e.g., [2,16]). In the particular case of rocky shores, the high intertidal zone has been typically considered a harsh environment where air exposure and desiccation are maximum [9,17,18] and where ecosystem engineers that buffer these physical conditions can be critical for other species (e.g., [9,20]). The low intertidal zone, on the other hand, is generally considered a physically benign habitat where ecosystem engineers facilitate the occurrence of other species via the provision of refugia from predators or competitors (e.g., [43,44]). However, this may not be the case of our low platform. Here, physical factors (i.e., hydrodynamics and sediment transport/scour) apparently override any biotic influence on species colonization (see Study Site and Organisms). Hence, the positive influence of *Petricola*-engineered microhabitats on species richness could largely be attributed to amelioration of physical conditions rather than the provision of enemy-free space. Experiments would ultimately be necessary to evaluate the above prediction.

Finally, our findings add to a few recent studies illustrating positive contributions of boring organisms to coastal species richness via the creation of boreholes [5,44,45]. Boring organisms have traditionally been viewed as bioeroders that can compromise the stability and persistence of habitats such as coastal rock, corals, and wood, with potential to reduce biodiversity [46–48]. Yet, bioerosion rates have rarely been compared to the rates of basal habitat formation (e.g., coral or wood production) and/or destruction due to other factors (e.g., physical rock erosion, coral reef degradation due to disease or acidification). If bioerosion is not a significant source of basal habitat loss (e.g., [22]), then boring organisms should better be construed as sources of microhabitat diversity and, thus, as potential benefactors of biodiversity. Further surveys and experiments will certainly be necessary to evaluate if the effects of boring organisms are also positive at other habitat types (e.g., coral reefs) and/or when considering boring species that substantially differ in borehole traits (e.g., size, shape, persistence).

Supplementary Materials: The following are available online at <http://www.mdpi.com/1424-2818/11/3/39/s1>, Figure S1: Barren rock surface in the lower platform, without (a) and with (b) accumulated sand. Figure S2: Species accumulation curve in mussel beds at the upper platform.

Author Contributions: M.B., J.L.G., and M.G.P. conceived and designed the study. M.B., L.P.A., and J.L.G. carried out the field research and sample analysis. M.B. analyzed the data and wrote manuscript drafts with input from J.L.G. L.P.A. and M.G.P. contributed to interpretation of findings and commented on manuscript drafts. All authors read and approved the final manuscript.

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