

Article

Internal Combustion Engines as the Main Source of Ultrafine Particles in Residential Neighborhoods: Field Measurements in the Czech Republic

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Abstract: Ultrafine particles (UFP, diameter < 100 nm) exposure has already been associated with adverse effects on human health. Spatial distribution of UFP is non-uniform; they concentrate in the vicinity of the source, e.g. traffic, because of their short lifespan. This work investigates spatial distribution of UFP in three areas in the Czech Republic with different traffic load: High traffic (Prague neighborhood—Sporilov), commuter road vicinity (Libeznice), and a small city with only local traffic (Celakovice). Size-resolved measurements of particles in the 5–500 nm range were taken with a particle

classifier mounted, along with batteries, GPS and other accessories, on a handcart and pushed around the areas, making one-minute or longer stops at places of interest. Concentrations along main roads were elevated in comparison with places farther from the road; this pattern was observed in all sites, while particle number distributions both close and away from main roads had similar patterns. The absence of larger particles, the relative absence of higher concentrations of particles away from the main roads, and similar number distributions suggest that high particle number concentrations cannot be readily attributed to sources other than internal combustion engines in vehicles and mobile machinery (*i.e.*, mowers and construction machines).

Keywords: ultrafine particles; nanoparticles; UFP; air pollution; traffic load; spatial distribution; UFP source; mobile measurement; field study

1. Introduction

Ultra fine particles (UFP, <100 nm in diameter) are the smallest fraction of particles. They are formed directly within combustion processes such as within internal combustion engines and/or they arise in the air from precursors. Traffic is currently the biggest source of UFP in big cities [1,2]. UFP represent a small fraction of the total particle mass due to their small size; however, they dominate the total particle counts or number concentrations. Their lifespan ranges from seconds to a few hours because of condensation and coagulation into larger particles. Thus, they are concentrated near their sources such as roadways, while the number concentration of UFP rapidly decays (>50% at 150 m) along with the distance from the busy road [3–5], there are even differences between particle concentrations measured kerbside and building side [6].

UFP exposure has already been associated with some adverse health effects involving respiratory and cardiovascular systems in humans [7–10]. There is also evidence that urban UFP, in comparison with larger particles, induced greater systemic oxidative stress in rats, which was enhanced by an inhibition of the anti-inflammatory capacity and subsequently led to larger atherosclerotic lesions [11]. Epidemiological studies have found association between residence location near the busy road and pulmonary impairments [12,13] as well as increase in the risk of cardiovascular disease [14].

With vehicle fleet turnaround along with the tightening of emissions limits for new on-road vehicles sold in the European Union, the relative contribution of other sources, namely home heating appliances, to the total particle mass in the Czech Republic is increasing [15], and often, improperly designed, maintained or operated stoves are the dominant sources of visible smoke. Such appliances, however, produce larger particles, typically hundreds of nanometers in diameter [16–18], and their exhaust is released from chimneys, while engines produce rather small particles, mostly on the order of 10 nm to tens of nm diameter [19,20], which are released directly in the streets. In addition, real driving emissions measurements show that engine UFP emissions are not evenly distributed among vehicles or along the path of travel, but rather, a small number of vehicles is responsible for a large fraction of the total emissions from the vehicle fleet [21,22], and a large fraction of UFP emissions from a given vehicle is attributable to short episodes with high emissions [23].

In this study, the UFP concentrations at the breathing levels on the streets were investigated in multiple residential neighborhoods in the Czech Republic with a hand-pushed instrument cart, with the goal of using particle size distributions and geographic distributions to investigate the spatial distribution of UFP and to provide a rough source apportionment between road traffic and local heating.

All measurements were conducted during the heating season, and it was assumed that home heating devices as well as their operation and emissions are distributed randomly (and on a large scale, evenly) throughout the neighborhood, and that high emitters will be spotted by visible smoke.

2. Experimental Section

The fast mobility particle sizer was used for measurements because it is, in contrast with scanning mobility particle sizer (SMPS), able to record rapid changes in particle number distribution caused by traffic, which SMPS could miss [24]. Therefore two fast mobility particle sizers, EEPS (Engine Exhaust Particle Sizer, model 3090, TSI, measurement of 32 channels distributed over 5–560 nm range, 1 Hz sampling frequency), and one condensation UFP counter (UF-CPC, model 200, Palas, 50% sensitivity at 5 nm, using butanol as working fluid), all already modified and mounted for in-vehicle operation, were mounted on 60 cm wide hand-operated carts, along with batteries (12 V 90 Ah LiFePo), inverters, GPS (eTrex 10, Garmin, 1 Hz sampling rate), and notebook computers (Figure 1).



Figure 1. Installed mobile measurement apparatus; EEPS—Engine Exhaust Particle, CPC—condensation particle counter.

Prior to the study, both EEPS instruments were co-located with a reference SMPS (SMPS 3034, TSI) for multiple days to assess their ambient air monitoring capabilities. The reference SMPS is owned by the Institute of Chemical Process Fundamentals, Academy of Sciences of the Czech Republic v.v.i. and was operated within the UFIREG Quality Assurance Program [25]. The SMPS measured particles in range of 10–500 nm and its scanning interval was set to 5 min.

The carts with instruments were pushed on different days through a residential neighborhood with local traffic (a small city of Celakovice), a neighborhood intersected by an arterial road frequented by commuter traffic to and from Prague (village of Libeznice), and an urban neighborhood surrounded by high-traffic motorways (Sporilov neighborhood in the city of Prague). Site details are given in Figure 2 and below. The carts were pushed through the neighborhoods to places of interest identified by representatives of local citizens and by the authors. Logging intervals of all instruments (EEPS, CPC and GPS) were set to 1 s. As the EEPS has shown increased noise during transportation due to vibrations (see Section 3.1), only data measured during frequent stops were used. A typical stop took one minute; this interval was extended in the case of unstable concentrations or anomalies due to the passage of individual vehicles.

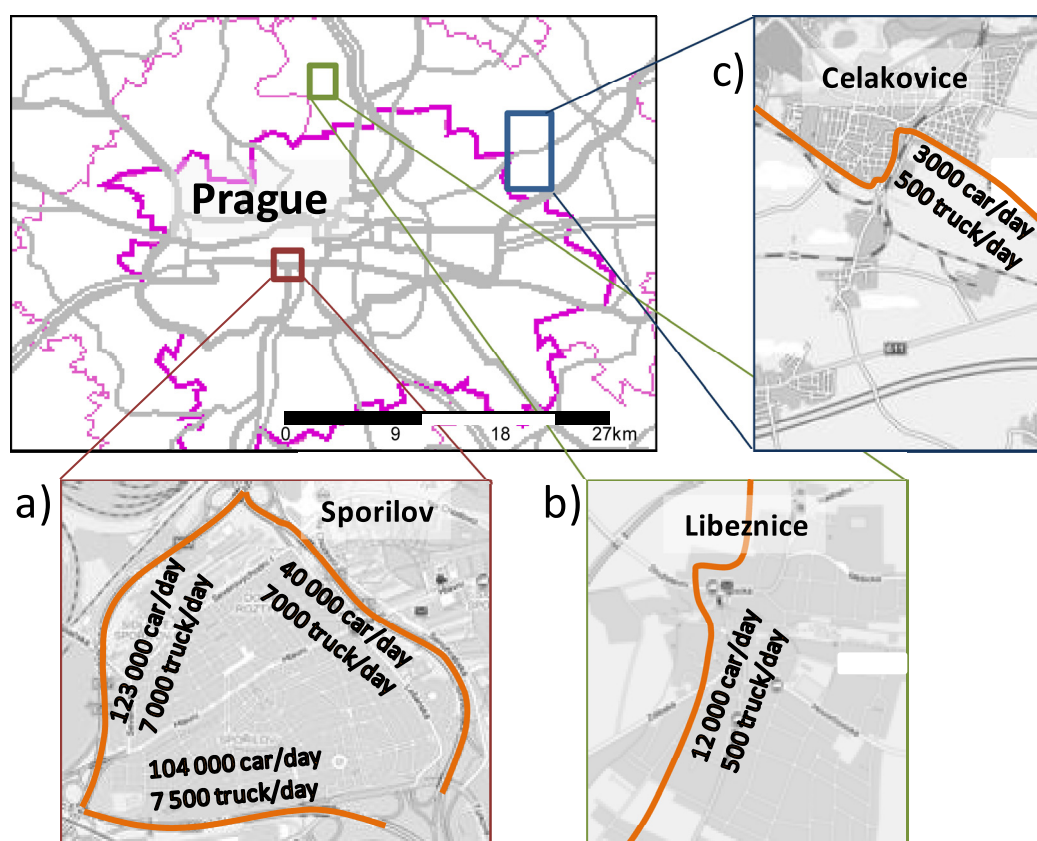


Figure 2. Location of measurement sites with given traffic density: (a) Spořilov, (b) Libeznice, and (c) Celakovice. Map sources: Geofond, Czech Geological Survey; Mapy.cz, Seznam.cz a.s.

Meteorological data from nearby observatories were used to supplement the data about temperature, humidity, and wind speed and direction.

High traffic location: Sporilov is a Prague residential neighborhood with 14,000 inhabitants completely surrounded by three freeways that are major heavy truck thoroughfares. Traffic intensity on two roads exceeded 100,000 cars per day in 2014 and each of the main roads is travelled by 7000 trucks per day [26] (Figure 2a). Measurements in this area were performed within weekdays on 20 February, and 6, 12 and 26 March 2014. Meteorological conditions were acquired of meteorological station Sporilov operated by Institute of Atmospheric Physics, Academy of Sciences of the Czech Republic.

Commuter traffic location: Libeznice is a village with 2000 inhabitants situated nearby Prague border. A road that goes through Libeznice is congested during morning and afternoon traffic peaks because the road is a feeder for many drivers coming into and out of Prague. There is a bypass away from Libeznice but that way is longer, therefore many drivers still go through the village. The daily traffic load is 12,000 cars and 500 trucks [27] (Figure 2b). Meteorological conditions were obtained from meteorological station Kojetice. The station is 5 km North from Libeznice.

Local traffic location: Celakovice is a small city situated 15 km northeast of Prague's built up area, and 3 km from the closest highway. Inhabitants commute daily to work to Prague by public transportation and many of them by car. This situation is typical for the most small cities and satellites around larger cities. Traffic intensity is relatively low, 3000 cars and around 500 trucks per day [28] (Figure 2c). Meteorological conditions were obtained from a station operated by Celakovice city.

3. Results

3.1. Effects of Vibrations

The online particle classifiers EEPS and reference SMPS size distributions and total particle count were in general agreement during a co-location exercise during the UFIREG Quality Assurance Program [25]. The EEPS, normally used for engine exhaust measurement, uses, however, electrometers as opposed to condensation particle counter, and exhibits a higher detection limit as well as zero drift of units of thousands of particles per cm^3 , which may be significant at ambient levels. This drift was determined by periodically placing a HEPA filter on the EEPS inlet and recording the "zero" size distribution, which was then subtracted from all measured size distributions (Figure 3). In addition, during the operation of a handcart, the EEPS was exposed to different vibrations than those encountered in the vehicle, resulting in increased noise. While this noise could be suppressed by improved suspension system (designed for the mass of the handcart as opposed to the mass of a vehicle), the effects of vibrations on EEPS measurement artifacts are not well quantified, and a decision was made to treat the data while moving and for about 10 s afterwards as indicative only. For this reason, EEPS data were considered to be quantitative only at standstill, and were frequently compared against the UF-CPC data. The effects of instrument travel on the measurement noise are shown in Figure 4.

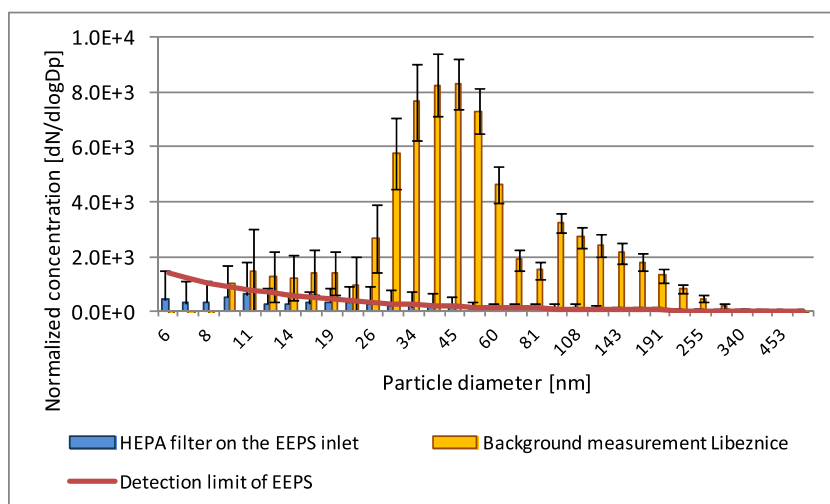


Figure 3. Comparison of manufacturer-supplied detection limit of EEPS, instrument electrometer noise (High-efficiency particulate arrestance (HEPA) filter on EEPS inlet) and the lowest measured number concentrations (Libeznice background) with standard deviations. The data representing the lowest measured concentration among campaigns ($4.2 \times 10^3 \text{ \#}/\text{cm}^3$, Libeznice background—best dispersion conditions) show that particles with diameter $>25 \text{ nm}$ are well above the detection limit. Concentrations measured on the other sites were several times/orders of magnitude higher. Measured concentrations were above detection limit of the instrument.

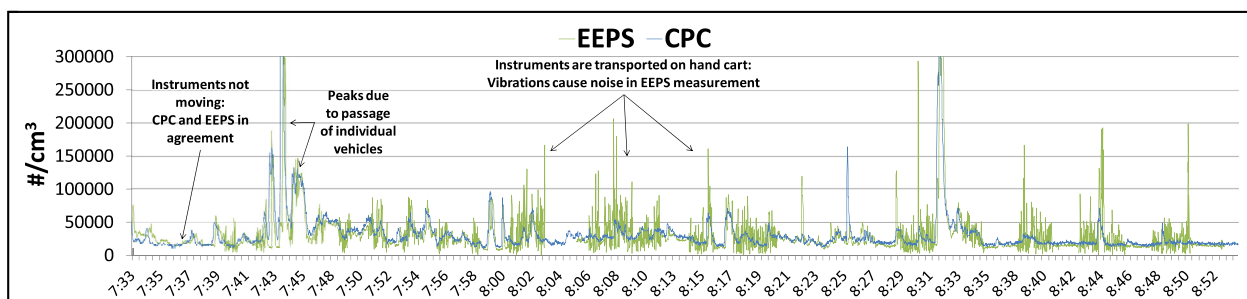


Figure 4. Comparison of particle number concentrations measured by CPC and EEPS; EEPS noise due to handcart vibration during its movement to another measurement stop and measurement stops are indicated on 30 January, EEPS and CPC are in agreement in total particle counts.

Measurements were carried out with one or two mobile apparatuses, which were pulled around the site making one-minute or longer stops at various places of interest. Measurements always took several hours, involving both peak and off-peak traffic, so that the intensity of traffic emissions was changing over the course of measurement. Presented concentrations include particles with electric mobility diameter between 5–100 nm. Distributions include all measured sizes and demonstrate a very small contribution of particles larger than 100 nm to the total particle counts.

3.2. Sporilov—Effects of Distance from Major Roads

Measurements in Sporilov were performed within weekdays on 20 February, and 6, 12 and 26 March 2014 with one mobile apparatus. Sporilov is a part of Prague with very high traffic load.

Morning and afternoon traffic peaks were recorded on 6 March (7:15–11:15 a.m.) and 20 February (1:15–6:00 p.m.), respectively. During both days the wind was stable and moderate (Figure 5). Near the road, *circa* 50 m away from the road, averaged concentrations were 3.4×10^4 #/cm³ (median 2.3×10^4 #/cm³) and 3.1×10^4 #/cm³ (median 3.1×10^4 #/cm³) for 20 February and 6 March, respectively. Further from the main roads the concentration decreased to average values 1.2×10^4 #/cm³ (median 9×10^3 #/cm³) on 20 February and 1.5×10^4 #/cm³ (median 1.3×10^4 #/cm³) on 6 March. Anomalous value 3.3×10^5 #/cm³ in the middle of building area was caused by a passing scooter. This value was not involved in average and median calculation (Figure 5). Distributions for both days are shown in Figure 6. There are strong peaks in particle concentration for 10 nm particles and less visible for particles with 40 nm diameter.

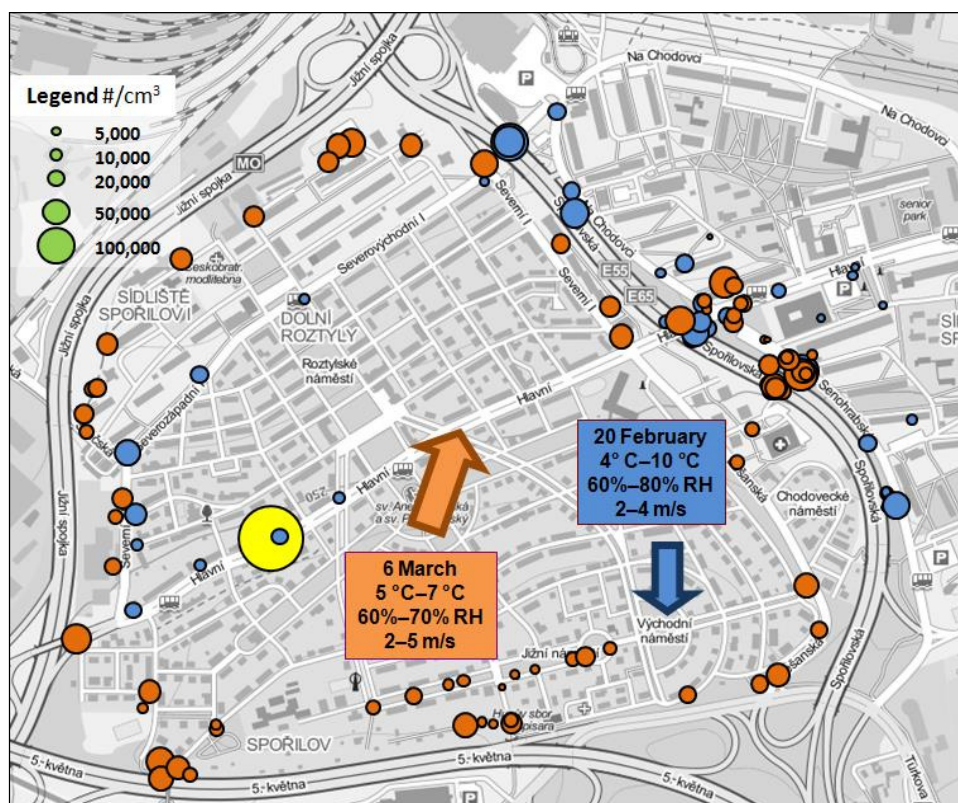


Figure 5. Measured average UFP concentrations and meteorological conditions in Sporilov on 20 February—blue, 6 March—orange. There can be seen substantial difference between higher concentrations measured closed to main roadways and lower concentrations measured deeper in a housing estate. Anomalous concentration due to scooter passing was indicated as a yellow dot; Map sources: Mapy.cz, Seznam.cz a.s.

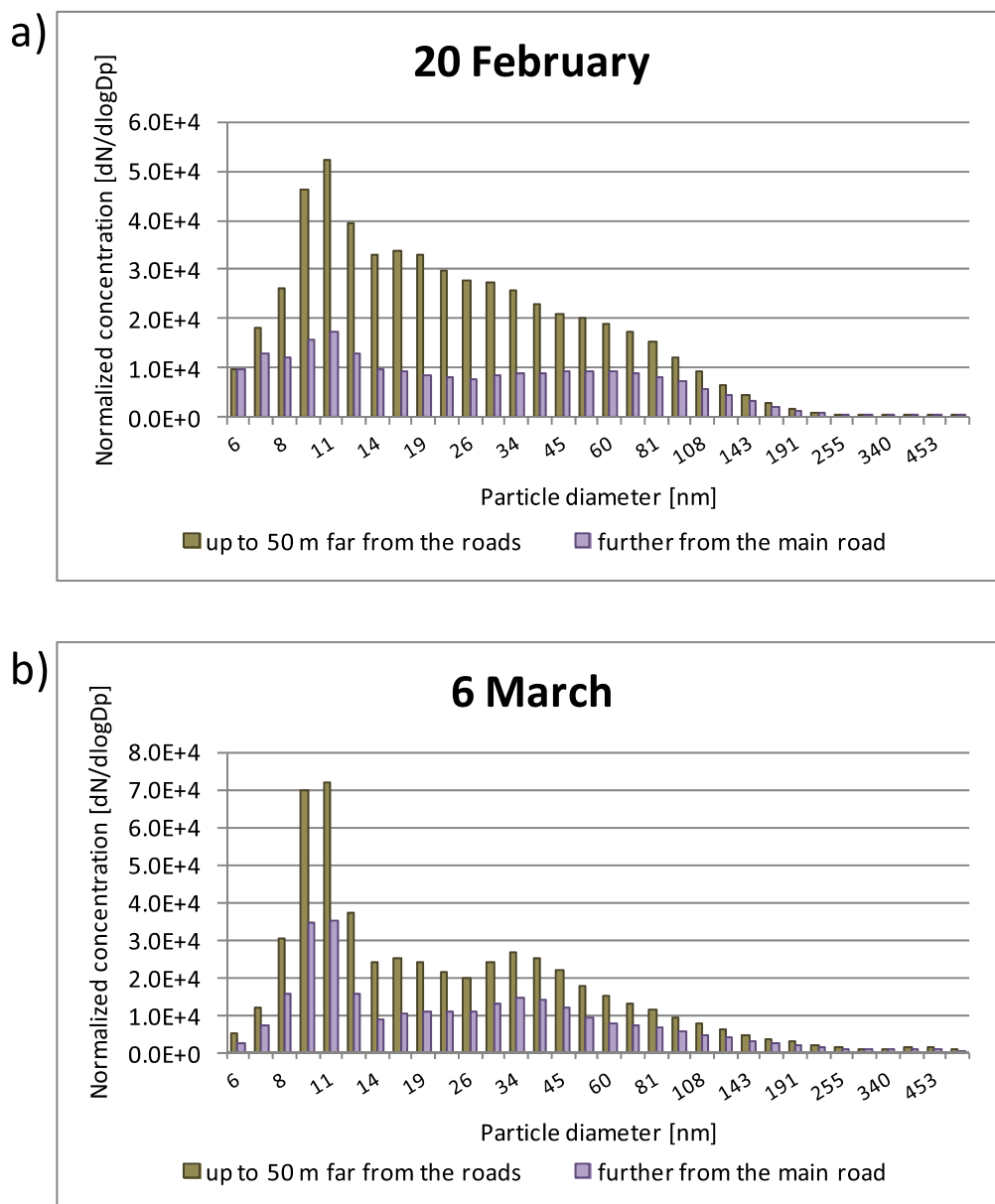


Figure 6. Particle number distribution measured in Sporilov on 20 February (a) and 6 March (b) according to distance from the main road. The highest peaks for both measurements were observed in 10 nm diameter particles, the second smaller peaks were formed in 20 and 40 nm particles for 20 February and 6 March, respectively. Similarity in distributions near and away from the main roadways and relative shift of smaller particles to larger ones in places further from the main roadways in comparison with places near the roadways indicated that particles measured deeper in a housing estate had a traffic origin.

3.3. Sporilov—Effects of Wind Direction

Measurement on 12 March was performed from 7:20 a.m. to 11:30 a.m. During this campaign the temperature and wind velocity gradually increased. Along with better dispersion condition, measured concentrations were decreasing. Meteorological conditions were recognized as the main determinant of number concentrations. Data were divided into three datasets: measurement 7:20–8:30 a.m. windless

condition, 8:30–10:00 a.m. moderate wind along a main road, and 10:00–11:30 a.m. measurements upwind from a major road where the emissions were blown away (Figure 7). Windless condition, while traffic culminated, led to averaged concentration $8.9 \times 10^4 \text{ \#/cm}^3$ (median $8.6 \times 10^4 \text{ \#/cm}^3$), distance from a main road was between 10–200 m. Moderate wind along a main road resulted in averaged concentration $3.8 \times 10^4 \text{ \#/cm}^3$ (median $3.5 \times 10^4 \text{ \#/cm}^3$), distance from a main road was between 50–170 m. Situation when emissions were blown away occurred after 10:00 a.m., under this condition, concentrations were the lowest with average $1.3 \times 10^4 \text{ \#/cm}^3$ (median $1.0 \times 10^4 \text{ \#/cm}^3$), however, still above the Prague urban background of $7.3 \times 10^3 \text{ \#/cm}^3$ [29]. Particle number distributions have been changing with dispersion conditions. The highest peak seen under windless condition with maximum around 11 nm particles decreased and an almost imperceptible peak in bigger sizes (*ca.* 100 nm) increased in comparison with the peak in smaller sizes (Figure 8).

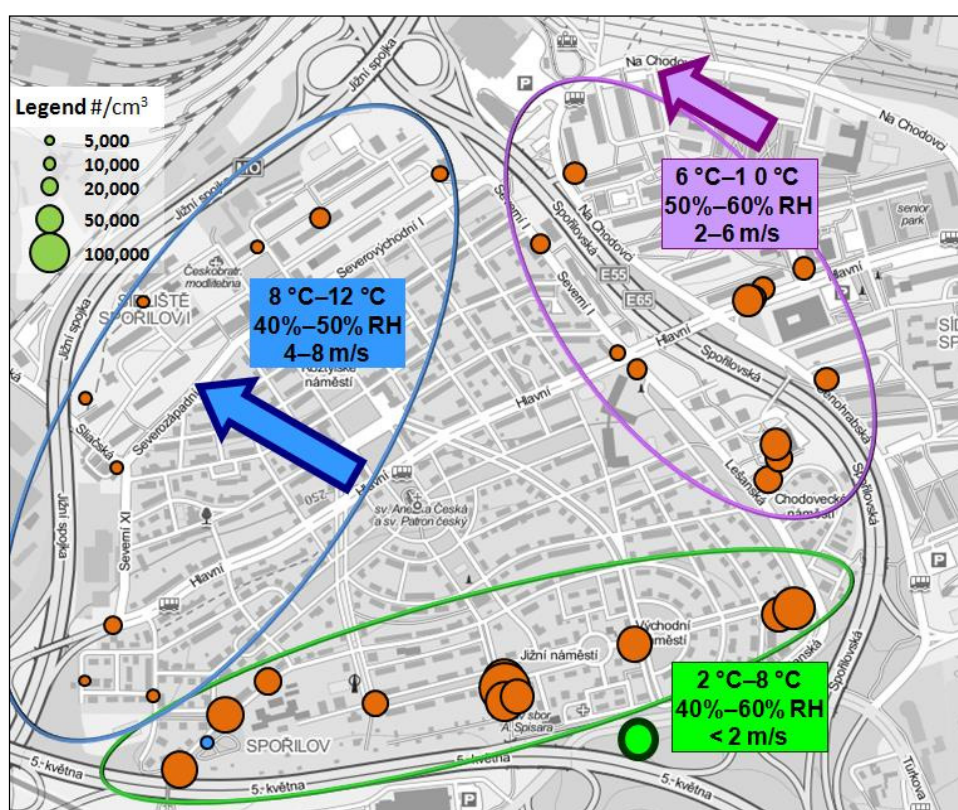


Figure 7. Measured average UFP concentrations within changing meteorological conditions in Sporilov on 12 March; during windless conditions, the highest averaged concentrations were measured in a building estate. Along with increasing wind speed measured particle number concentrations were gradually decreasing. Even within good dispersion conditions the particle number concentrations exceeded Prague background. Dots in a green ellipse represent measurement between 7:20–8:30 a.m.; purple, 8:30–10:00 a.m.; and blue, and 10:00–11:30 a.m.; Map sources: Mapy.cz, Seznam.cz a.s.

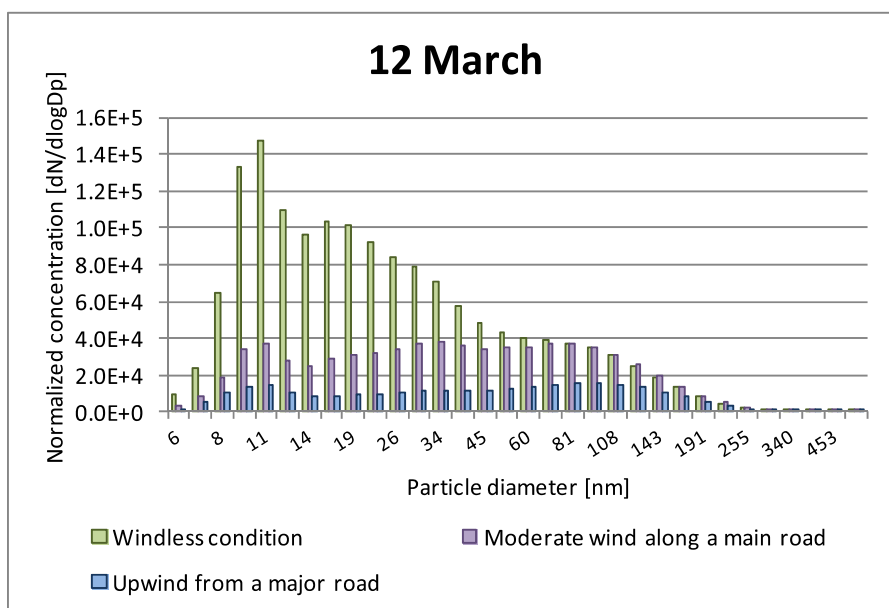


Figure 8. Particle number distribution within changing meteorological conditions in Sporilov on 12 March. Within windless condition a particle number distribution showed similar pattern as measured on other days (20 February and 26 March; Figure 3) close to main roadways although this measurements involving places even further from the main roadway. Along with increasing wind speed relative proportion of smallest particles decreased while larger particles increased relatively.

3.4. Sporilov—Effects of Truck Acceleration Hotspot

Additional measurement was performed on 26 March from 10:45 a.m. to 1:20 p.m. after dilution of morning inversion. The highest concentrations, with a peak of 1.0×10^6 #/cm³, were measured on the footbridge over a freeway. The eastbound section of the freeway is traveled by trucks, which have made a transit through Prague, often through extended periods of heavy congestion, entered the freeway on a ramp located just west of the footbridge, and are accelerating from the ramp while negotiating an incline. Earlier simulation of such conditions in an engine laboratory by this group [23] has demonstrated that extended low load leads to the cooling of the combustion chamber and of the exhaust system, which in turn results in worsened combustion, formation of deposits within the engine and the exhaust system, and loss of efficiency of catalytic devices due to low exhaust temperature. When such engine is subsequently operated at higher load, the deposits are driven off, and the emissions of particulate matter and temporarily increased; the emissions of carcinogenic polyaromatic hydrocarbons are increased an order of magnitude during such conditions over a stabilized high load operation.

The concentrations on the footbridge, nominally in hundreds of thousands of particles per cubic cm, were fluctuating, with peaks attributed to the passage of individual vehicles (Figure 9). The concentrations decreased markedly with the distance from the freeway to around 5.0×10^4 #/cm³ 50 meters away from the main road. Except for this acceleration site and related main road, the concentrations were relatively low thanks to good dispersion condition, average 7.0×10^3 #/cm³, median 6.0×10^3 #/cm³ (Figure 10). Distributions are shown in Figure 11. A bimodal distribution (10 nm and 20 nm peaks) was observed for site on/around the footbridge, with concentrations

decreasing with the distance from the road. Farther places had low particle concentration, even though a peak in 10 nm particles was still noticeable. These results contain only particles in range 5–100 nm; instrument channels counting larger particles were affected by a measurement artifact (a dog hair lodged on an EEPS electrometer).

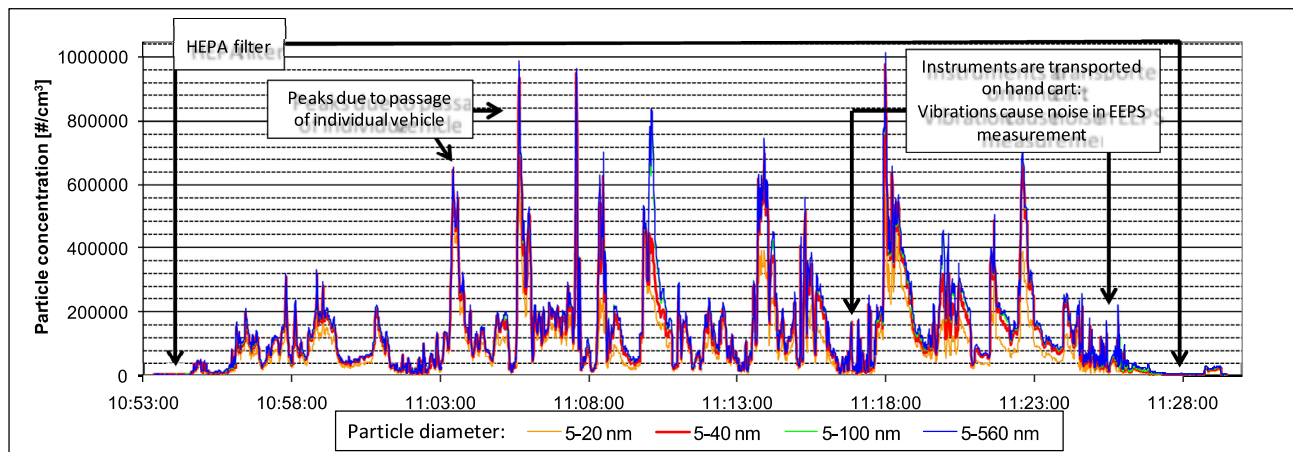


Figure 9. Number particle concentration time course within hotspot measurement in Sporilov on 26 March. Truck acceleration is peculiar for this measurement site. Additionally, before the trucks reach this place, they often spend some time in congestion. Under congestion condition extended low load leads to the cooling of the combustion chamber and of the exhaust system, which in turn results in worsened combustion, formation of deposits within the engine and the exhaust system, and loss of efficiency of catalytic devices due to low exhaust temperature. When such engine is subsequently operated at higher load, the deposits are driven off, and the emissions of particulate matter and temporarily increased.

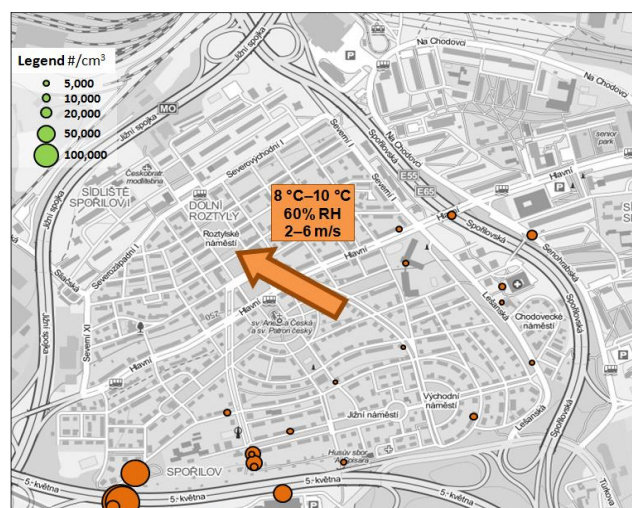


Figure 10. Measured average UFP concentrations near a hotspot and around in Sporilov on 26 March; High concentrations measured on a footbridge over a freeway were a result of emissions from truck accelerations after extended periods of heavy congestion. Impact of this hotspot can be seen in a near buildup area. Due to good dispersion conditions majority locations of housing estate showed low concentrations. Map sources: Mapy.cz, Seznam.cz a.s.

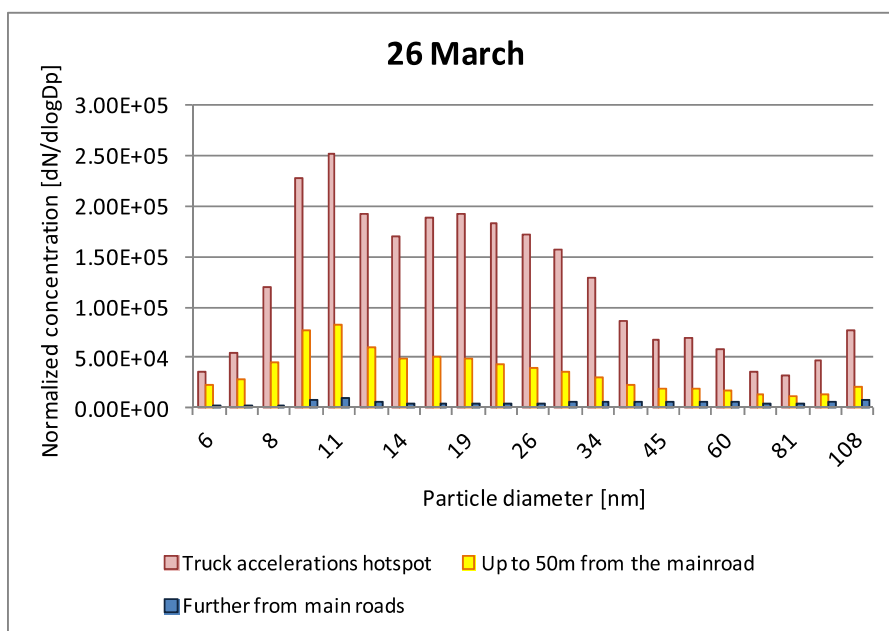


Figure 11. Particle number distributions near a hotspot and around according to distance from the hotspot and a main road measured in Sporilov on 26 March; Truck accelerations hotspot showed bimodal distribution with peaks at 10 and 20 nm. There should probably be another peak in larger particles; however, those data are not available because of the contamination of the corresponding instrument channels. Particles concentrations decreased with the distance from the road (up to 50 m), while the size distribution pattern remained similar. Farther places had low particle concentrations, even though a peak in 10 nm particles is still noticeable.

3.5. Libeznice

Measurement was performed during morning traffic peak on 15 May 2014 (6:20–8:15). Particles with diameter <10 nm were excluded because of periodically recurring artifact affecting the smallest two channels, resembling a nucleation event, which was considered improbable. Actual concentrations varied from tens of thousands to hundreds of thousands with peaks exceeding a million particles in cubic centimeter after passing of a high emitting vehicle. Average concentrations are plotted in Figure 12 and were $2.7 \times 10^4 \text{ \#/cm}^3$ (median $2.5 \times 10^4 \text{ \#/cm}^3$) next to the main road. Further from the main road, the number concentrations decreased rapidly to $7.5 \times 10^3 \text{ \#/cm}^3$ at 100 m and $4.5 \times 10^3 \text{ \#/cm}^3$ at 150 m.

A local and temporary hotspot was created by a small riding lawnmower. Lawnmowers are powered by gasoline engines with no electronic controls and no exhaust after treatment, and are, up to 19 kW power, not subject to any limit on particle emissions. As such, they can be a considerable source of particle emissions in residential neighborhoods. Further, the particles are emitted in the immediate vicinity of the operator, and therefore could pose significant health risk to the workers using such equipment on a regular basis.

The concentrations of particles were in the order of hundreds of thousands of particles per cm^3 circa ten meters away from the mower (Figure 12), a higher than expected value considering favorable

dispersion conditions. Background concentration was measured in the outskirts between rapeseed fields, concentration were around $3.8 \times 10^3 \text{ \#/cm}^3$.

Number distributions revealed difference between two sources of excess emissions: smoking vehicles and a lawn mower (Figure 13). Smoking vehicles (automobiles, buses and trucks typically in unsatisfactory technical condition) emitted particles with larger diameter, which is consistent with the visual observation of smoke. The lawn mower was responsible for similar amount of particles by count, but those particles were much smaller than the wavelength of the visible light and therefore practically invisible.

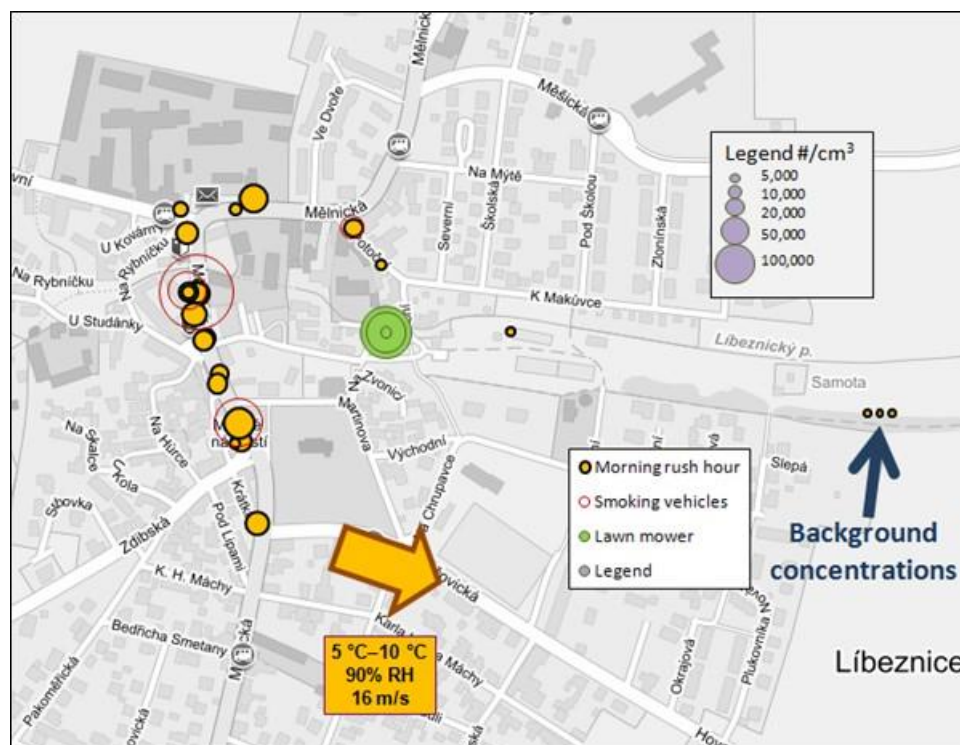


Figure 12. Measured average UFP concentrations and meteorological conditions on 15 May in Libeznice; UFP number concentrations closed to a main road were higher than concentrations measured further from the main road. Attention should be paid to concentrations measured near riding lawn mower (green dots), which are comparable with concentrations measured within passing smoking vehicle (red rings). Operating lawn mower created temporary local UFP hotspot. Lawnmowers are powered by gasoline engines with no electronic controls and no exhaust after-treatment, which are, up to 19 kW power, not subject to any limit on particle emissions. Particles with diameter $<10 \text{ nm}$ were excluded because of periodically recurring artifact affecting the smallest two channels, resembling a nucleation event, which was considered improbable; Map sources: Mapy.cz, Seznam.cz a.s.

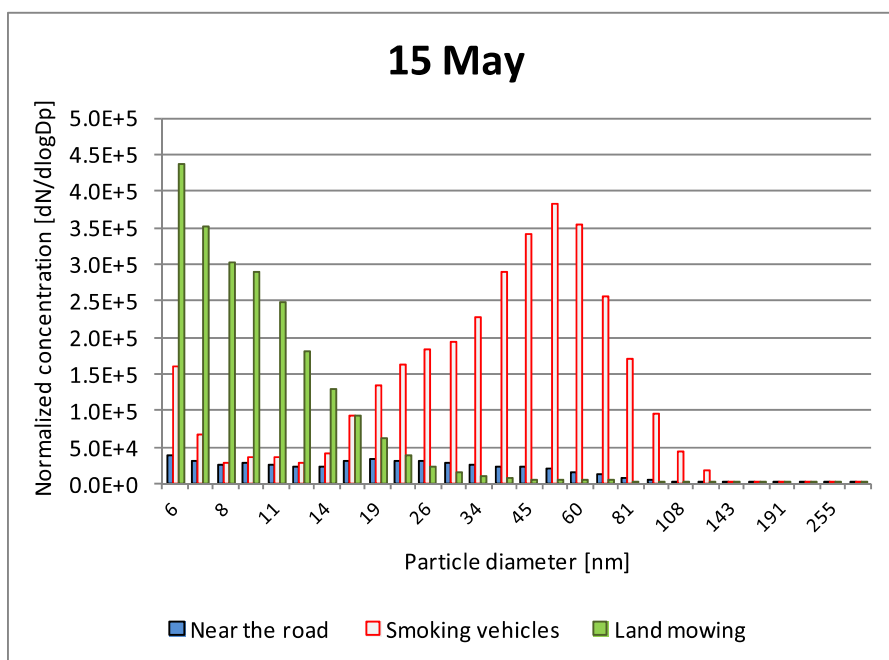


Figure 13. Particle number distribution measured in Libeznice on 15 May according to source of particles (road traffic excluding high emitters, smoking vehicles, lawn mower). Smoking vehicles emitted particles with larger diameter, which is consistent with the visual observation of smoke. The lawn mower is responsible for similar amount of particles by count, but those particles are much smaller than the wavelength of the visible light and therefore practically invisible. Particles with diameter <10 nm were excluded because of periodically recurring measurement artifacts affecting the smallest two channels.

3.6. Celakovice

Measurement in Celakovice city was carried out with two mobile apparatuses measuring in parallel; the first apparatus equipped with EEPS and CPC measured concentrations next to the main road, the second measured further from the road (*ca.* 100–150 m). Measurements were performed on two days, 30 January (morning and afternoon measurement) and 3 February (all day measurement). Data presented herein comprise all measured particles (5–500 nm) due to inconsistency results obtained from CPC and EEPS. On 3 February, particle number concentrations from the EEPS were consistently lower by about 6500 \#/cm^3 compared to the CPC during campaign on 3 February, an offset attributed to an erroneous zeroing of the EEPS. As CPC can be considered the ultimate source of total particle number concentrations, total counts calculated from EEPS size distributions were corrected for the zero drift to match the CPC. The UF-CPC measures particles with diameter between 4 nm–10 μm , while the EEPS range is 5–500 nm; difference in instrument ranges on the upper end was considered to be insignificant because vast majority of the urban particles, from particle number concentration point of view, fell within the EEPS range. However, we were not able to reconstructed particle number distribution due to the shifted zero offset on 3 February, therefore distribution is presented only for 30 January measurement.

During traffic peaks concentrations next to the road varied around $2 \times 10^4 \text{ \#/cm}^3$ with typical peaks after passing older vehicles around $4 \times 10^4 \text{ \#/cm}^3$; together average particle number concentrations for

morning traffic peaks were $3.6 \times 10^4 \text{ \#/cm}^3$ (median $3.0 \times 10^4 \text{ \#/cm}^3$) on 30 January and $3.2 \times 10^4 \text{ \#/cm}^3$ (median $2.0 \times 10^4 \text{ \#/cm}^3$) on 3 February. Maximal measured concentration, more than $5 \times 10^5 \text{ \#/cm}^3$, occurred after the passages of high emitters (a tractor and several automobiles). The influence of the main road was not noticeable at 150 m from the road, where the concentrations were mainly around $1.5 \times 10^4 \text{ \#/cm}^3$ or lower (Figure 14). Particle number distributions measured next to the road on 30 January and 150 m away from the road are shown in (Figure 15) and are similar for both places. It suggests that traffic, even in small city with low traffic load, is a key contributor to the ambient UFP concentrations.

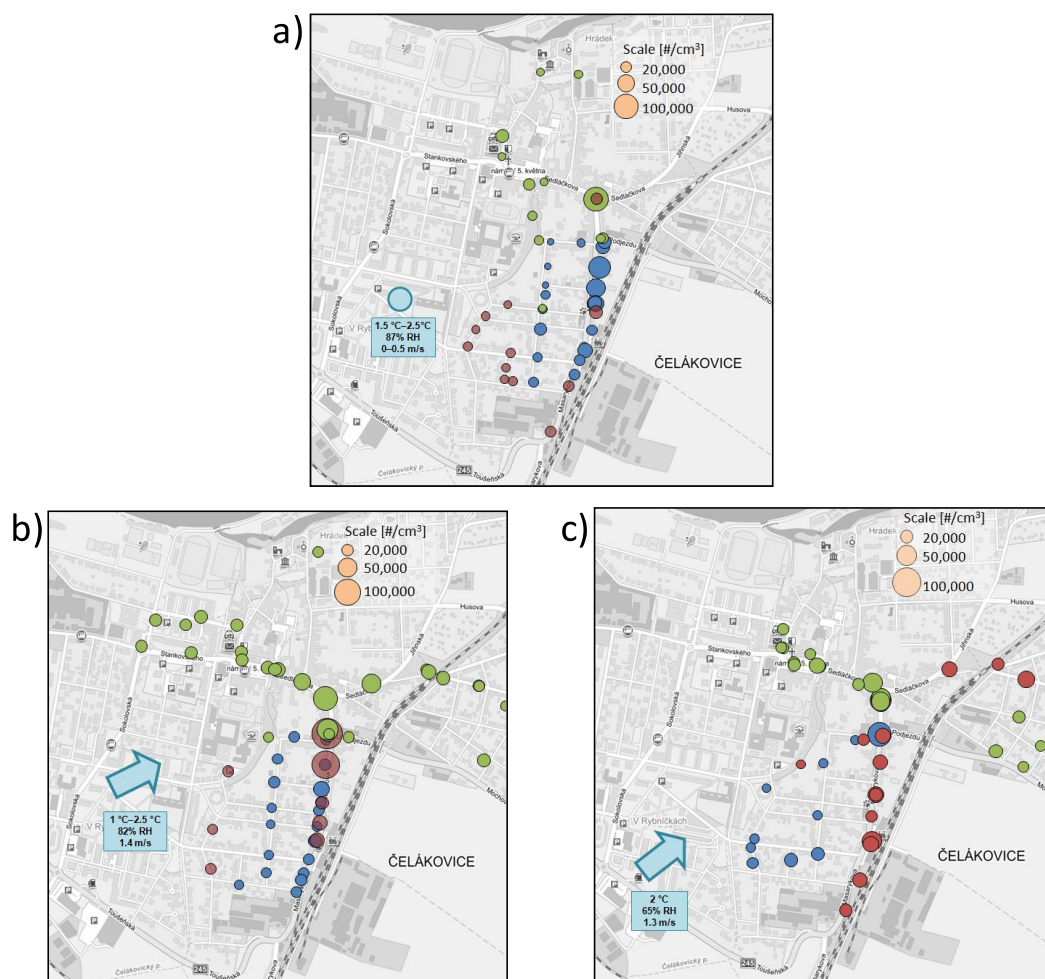


Figure 14. Simultaneous particles number concentration measurements with two mobile apparatuses. Dot colors indicate time intervals. The first apparatus measured closed to a main road while the second measured parallel away from the main road. Concentrations along main roads were elevated in comparison with places further from the road. (a) Particle number concentrations 30 January; simultaneous measurement at 7:30–8:00 a.m. blue points, 8:00–8:20 a.m. red points, and 8:20–9:00 a.m. green points. (b) Particle number concentrations on 3 February; simultaneous measurement at 6:45–7:20 a.m. blue points, 7:20–7:35 a.m. red points, and 7:35–9:00 a.m. green points. (c) Particle number concentrations on 3 February; simultaneous measurement at 5:10–5:15 p.m. blue points, 5:15–5:50 p.m. red points, 5:50–6:30 p.m. green points; Map sources: Mapy.cz, Seznam.cz a.s.

Afternoon traffic peak was measured on 3 February, while one apparatus stayed on an intersection and the second operated away from the main road. The intersection represents a hotspot of the Celakovice city, it is the part of the main road and many cars have to slow down or stop and then accelerate to common city speed. Average particle number concentration was $6.1 \times 10^4 \text{ \#/cm}^3$ (median $5.7 \times 10^4 \text{ \#/cm}^3$) next to the intersection. The apparatus measured away from the crossroad (ca. 150–250 m) measured average concentration $1.4 \times 10^4 \text{ \#/cm}^3$ (median $1.2 \times 10^4 \text{ \#/cm}^3$) during the traffic peak.

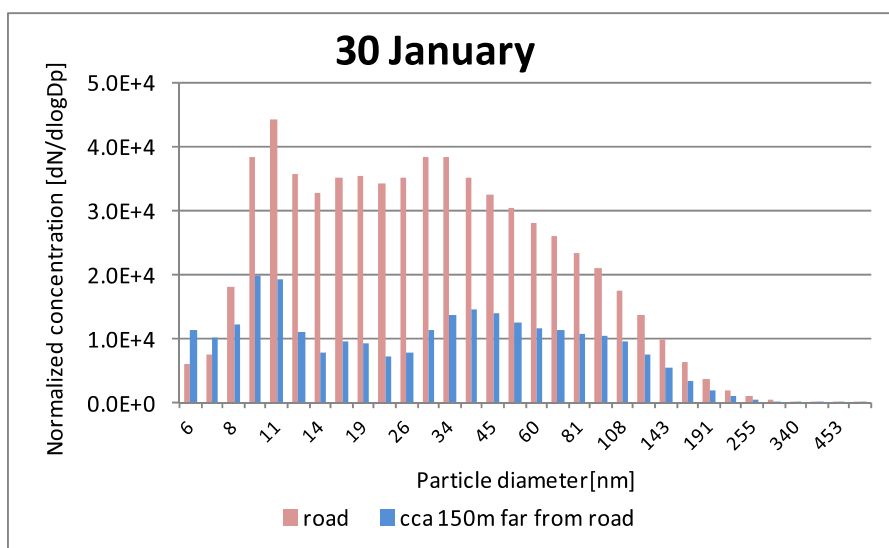


Figure 15. Particle number distributions during simultaneous measurement by two instruments sets, one near the main road, the second ca. 150 m away from the road on 30 January, 7:30–8:00 a.m., corresponding to blue dots in Figure 14a. Concentrations measured away from a main road in Celakovice had similar distribution as particles measured next to the main road.

Concentrations after morning traffic peaks were stable around $1.5 \times 10^4 \text{ \#/cm}^3$ up to 10:00 a.m., then gradually decreased to concentrations $1.0 \times 10^4 \text{ \#/cm}^3$ (3 February) and $7.0 \times 10^3 \text{ \#/cm}^3$ (30 January) at midday. Lower concentrations on 30 January were not likely to have been caused by better dispersion conditions because wind speed were around 0.5–2 m/s, while on 3 February the wind speed varied between 0.5–3 m/s (Figure 16). We expected lower concentrations in comparison with other studies, so there should be other sources of ultrafine particles, but no such sources were identified during the measurements. A heavily traveled freeway to Prague is 3 km from the measurement site. UFP generated by vehicles going on the highway should have already coagulated into large particles before they reached Celakovice; however, at this time, the measurement was downwind from the freeway.

After the afternoon traffic peak, 3 February 6:00–6:30 p.m., concentrations were stabilized around $2.0 \times 10^4 \text{ \#/cm}^3$ with some exceeding values, while the source was always vehicle e.g., idling diesel vehicle in front of a house or a high emitting vehicle passage. Together, average particle number concentration was $2.2 \times 10^4 \text{ \#/cm}^3$ (median $1.8 \times 10^4 \text{ \#/cm}^3$).

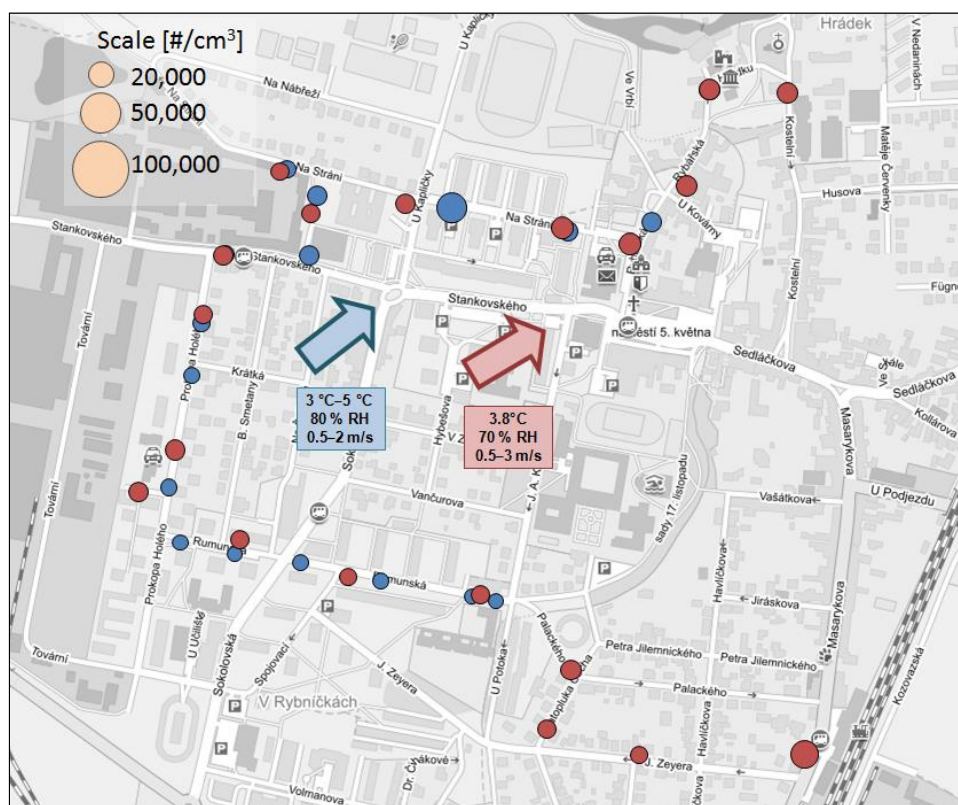


Figure 16. Particle number concentrations measured after traffic peaks on 30 January—blue, and 3 February—red; Vehicles on the main road going through the city were not proved as the main source due to very low traffic load after morning peak. Other source of UFP emissions is a motorway, about 3 km and upwind from the measurement site. Map sources: Mapy.cz, Seznam.cz a.s.

4. Discussion

4.1. Instrumentation

The results confirmed that fast particle classifiers are useful for mobile measurements [24], as the size information is helpful in the explanation of the origin of the particles, and the time resolution allows for detection of temporary or localized sources such as passing high emitting vehicles. The particle classifiers closely matched both SMPS data during a stationary co-location exercise, and condensation particle data during field measurements when not moving. The classifiers were subject to increased noise in data when traveling on a small cart, and to a non-negligible zero drift, suggesting that they should be complemented by a CPC as a data validation tool. Mounting the particles on a handcart has allowed much more detailed measurements that could have been accomplished by instrumented motor vehicles.

4.2. Effects of Traffic

In all locations, elevated UFP concentrations associated with the proximity of internal combustion engines (mostly vehicular traffic, but also a lawn mower) were found. The size distribution was generally bi-modal, with nucleation peak around 10 nm, and a second accumulation mode peak in the

tens of nm range, corresponding to, for example, thermodenuded diesel exhaust [20]. The particles were generally smaller compared to the size distributions measured at background stations throughout Europe [30] as well as at the Czech background station in Košetice [31], but well in agreement in terms of size and concentration with observations in the center of Milano [32]. The nucleation peak was more pronounced, and the accumulation mode peak shifted towards smaller particles, in the proximity of the source, except for high emitting vehicles, where the accumulation mode has dominated the size distribution. Farther from the source, the concentration diminished, in agreement with [5], the nucleation peak was smaller or not present, and the accumulation mode peak was shifted to larger particles. The contribution of high emitting vehicles was readily noticeable. Highest concentrations were observed in an area where truck engines were operated at high load which was likely to have immediately followed extended periods of low load, which is in agreement with earlier results of engine tests showing detrimental effects of extended low load [23].

It is the opinion of the authors that avoiding overloading of the roads (seen in Sporilov and Libeznice), during which both the engine combustion efficiency and the catalyst efficiency can decrease, improved vehicle emissions inspection program eliminating high emitting vehicles (smoking vehicles were observed in all places), support of public transit and non-motorized passenger transport, shifting transit road transport to more ecological railways, and other means are some relatively low-cost measures that would bring a relatively fast reduction in ambient UFP concentrations.

4.3. Effects of Local Heating Appliances

Virtually all measurements were done at temperatures when regular ordinary use of home heating appliances was expected. In the Czech Republic, large apartment buildings often use central heating or have their own natural gas furnaces, while family houses are heated by a mix of natural gas, wood and coal. The mix is dependent on the economic situation of the families; in economically depressed areas, the use of poor-quality coal, waste products, and other inferior fuels is not uncommon.

In the Sporilov neighborhood, particles originating from residential houses in the central and southern parts of the triangle would be expected to be present (a) in the central part of the triangle formed by the freeways in Figure 5; (b) downwind of the central part of the triangle along its northwest border in Figure 7, and (c) throughout the southern part of the triangle in Figure 9. Further, such particles would be expected to be observed at least at some points in Libeznice and Celakovice. No substantially elevated concentrations and no hotspots (it was recognized that hotspots from heating appliances would be broader than from engines), which were not attributed to internal combustion engines, were, however, found at any of the places examined. The domination of traffic sources is consistent with ambient measurement in schools in Australia [33] and ambient measurements in Pittsburgh [34]. The large contrast between busy streets and urban background was also observed in a Copenhagen study, which concluded that a “significant underestimation of disease burden may occur when relying too much on the regulated components” [35].

However, when emissions limits of heating appliances and engines are compared either on a PM mass per kg of fuel or PM mass per cubic meter of exhaust, the currently allowed maximum emissions of heating appliances exceed heavy duty on-road engines by orders of magnitude. It therefore appears that while local heating appliances might be a significant source of particulate matter in general, and

notably total particle mass, they were not found to be a noticeable contributor to UFP at street level, possibly as they produce larger particles which are also less available at the street level due to smokestacks.

5. Conclusions

This work investigated the spatial distribution of UFP in three areas in the Czech Republic with different traffic load by fast mobility particle sizers (EEPS) mounted on a handcart and pushed around the areas. Utilization of a fast mobility particle sizer was appropriate for UFP measurement influenced by traffic, because it was able to record rapid changes in particle number distribution.

In all locations, the highest concentrations were observed in the vicinity of vehicular traffic (or operation of non-road engines), and observed particle number distributions both close to and away from main roads had similar patterns consistent with size distributions observed in engine exhaust. Further, the observed patterns corresponded to findings from vehicle emissions studies, attributing high emissions to extended idling, transient operation, and high emitting vehicles. On the contrary, the particle number concentrations in other parts of the neighborhoods were relatively small, and the share of particles larger than 100 nm was relatively low in all number size distributions. Therefore, it appears that the number counts of UFP at breathing level are dominated by UFP originating from vehicular traffic. Vehicles, while contributing only tens of percent to the total emissions of total particulate mass, could therefore bear much larger share in the total health effects of particulate matter, due to a combination of high share of UFP and proximity of sources to citizens. The study reiterates the need to further reduce vehicle emissions through reduction in vehicular traffic, improvements in vehicle maintenance, cleaner fuels and technologies, and other measures.

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Author Contributions

All authors participated in transport and operation of the instrumentation during the measurement campaigns. Lubos Dittrich and Martin Pechout designed and fabricated the carts and instrument

mounting. Jakub Ondracek ran the reference SMPS and took part in the collocation exercise. Jitka Stolcpartova and Michal Vojtisek-Lom analyzed and interpreted the data. Michal Vojtisek-Lom was the visionary behind the measurement setup and methods and leader of the project.

Conflicts of Interest

The authors declare no conflict of interest.

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