



Article

Characterization of Particulate Matter ($PM_{2.5}$ and PM_{10}) Relating to a Coal Power Plant in the Boroughs of Springdale and Cheswick, PA

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Abstract: Ambient concentrations of both fine particulate matter (PM_{2.5}) and particulate matter with an aerodynamic diameter less than 10 micron (PM₁₀) were measured from 10 June 2015 to 13 July 2015 at three locations surrounding the Cheswick Power Plant, which is located between the boroughs of Springdale and Cheswick, Pennsylvania. The average concentrations of PM₁₀ observed during the periods were 20.5 \pm 10.2 μg m $^{-3}$ (Station 1), 16.1 \pm 4.9 μg m $^{-3}$ (Station 2) and 16.5 \pm 7.1 μg m $^{-3}$ (Station 3). The average concentrations of PM_{2.5} observed at the stations were 9.1 \pm 5.1 μg m $^{-3}$ (Station 1), 0.2 \pm 0.4 μg m $^{-3}$ (Station 2) and 11.6 \pm 4.8 μg m $^{-3}$ (Station 3). In addition, concentrations of PM_{2.5} measured by four Pennsylvania Department of Environmental Protection air quality monitors (all within a radius of 40 miles) were also analyzed. The observed average concentrations at these sites were 12.7 \pm 6.9 μg m $^{-3}$ (Beaver Falls), 11.2 \pm 4.7 μg m $^{-3}$ (Florence), 12.2 \pm 5.3 μg m $^{-3}$ (Greensburg) and 12.2 \pm 5.5 μg m $^{-3}$ (Washington). Elemental analysis for samples (blank – corrected) revealed the presence of metals that are present in coal (i.e., antimony, arsenic, beryllium, cadmium, chromium, cobalt, lead, manganese, mercury, nickel and selenium).

Keywords: PM_{2.5}; PM₁₀; coal-fired power plant; particulate matter emissions

1. Introduction

With over 600 active coal-fired power plants in the US, 39% of the country's total electricity generation is attributed to coal (Available online: http://www.eia.gov). While these plants may bring jobs and prosperity to their surrounding regions, they also emit dangerous pollutants into the atmosphere. According to the US Environmental Protection Agency (EPA) National Emissions Inventory, coal-fired power plants emit sulfur dioxide, oxides of nitrogen and particulate matter into the atmosphere. They also emit 84 of the 187 Hazardous Air Pollutants (HAP) regulated by the US EPA [1]. In addition to the stack emissions from the coal-fired power plant, coal handling may also emit pollutants into the atmosphere and thus degrade the air quality in the vicinity near the power plant [2]. Several studies have been conducted on the relationship between particulate matter and emissions from coal-fired power plants [2–14]. Particulate matter measured near a coal-fired power plant is known to contain a number of harmful chemicals that are present in coal and are also known to have carcinogenic properties, such as antimony, arsenic, beryllium, cadmium, chromium, cobalt, lead, manganese, mercury, nickel, selenium and polycyclic organic matter (POM) [2-4]. All of these chemicals are known to be hazardous and are thus regulated by the US EPA as HAPs under the Clean Air Act, as amended in 1990 [15]. The emission of these pollutants into the atmosphere is known to be dangerous to both human health and welfare. Exposure to high concentrations of HAPs can lead to a number of adverse health effects such as damage to the eyes, skin, lungs, kidneys and the nervous system, and can even cause cancer, pulmonary disease and cardiovascular disease [14,16]. Particulate

matter ($PM_{2.5}$), which is particulate matter with an aerodynamic diameter of 2.5 microns or less, is also a dangerous atmospheric pollutant due to its small size, which can travel deep into people's lungs and lead to a number of severe health effects. Elevated concentrations of $PM_{2.5}$ are known to be associated with cardiovascular issues (heart disease, heart attacks, etc.) as well as respiratory issues, reproductive issues and even cancer [16,17]. In addition to harming human health, coal-fired power plants can also lead to a number of environmental impacts as well, such as acidification of the environment, bioaccumulation of toxic metals, the contamination of water sources, reduced visibility due to haze as well as degradation of buildings and monuments.

The Cheswick Power Plant is located in the southwestern part of Pennsylvania along the Allegheny River right between the boroughs of Springdale (40.5414° N, 79.7821° W) and Cheswick (40.5416° N, 79.8002° W) (Figure 1). While the Cheswick Power Plant has brought economic benefits to the small boroughs of Springdale and Cheswick, PA, the plant also brought with it a multitude of harmful effects on both human health and welfare. According to the Allegheny County Health Department's Point Source Emissions Inventory Report, the Cheswick Power Station is the largest point source emitter of both criteria pollutants and hazardous air pollutants (HAP) in Allegheny County, PA [18]. In 2006, the plant was listed as the 17th Dirtiest Power Plant for sulfur dioxide [1]. In 2010, the plant once again made headlines, ranking 41st in the Top Power Plant Hydrochloric Acid Emitters and 91st in the Top US Power Plant Lead Emitters [19]. Since then, significant efforts have been made to reduce emissions [18]. Overall, emissions of carbon monoxide and sulfur dioxide did decrease in 2011, while emissions of nitrogen oxides, particulate matter, NO_x, and volatile organic compounds (VOC) increased [18]. However, the plant remains Allegheny County's largest point source emitter of both criteria pollutants and hazardous air pollutants. The objective of this study was to measure the air quality at three different locations within the two boroughs surrounding the power plant in order to determine concentrations of both PM_{2.5} and PM₁₀ due to emissions from the power plant. Particulate matter measurements measured by the PA Department of Environmental Protection (DEP) were also analyzed for the study period and compared against the measurements taken for this study.

2. Experiments

2.1. NCSU Monitoring Stations

Three experimental sites were set up within Springdale and Cheswick, PA, each within a mile of the Cheswick power plant (Figure 1). There were two sampling periods in this study. Two of the three sites took samples from 10 June 2015 to 27 June 2015 and then from 30 June 2015 to 13 July 2015. Station 1 was located at 244 Center St., Springdale, PA (40.5402° N, 79.7884° W), which is less than half of a mile from the power plant. Samples were taken at this location during both sampling periods. Station 2 was located at 200 Hill Avenue, Cheswick, PA (40.542° N, 79.80° W), which is approximately one mile from the power plant. The sampling period for this study was from 10 June 2015 to 27 June 2015. The third station was located at 1212 Fairmont St, Cheswick, PA (40.5465° N, 79.8031° W). The sampling period for this site was from 30 June 2015 to 13 July 2013. All three of these sites are in residential areas, very close to the power plant. The monitoring sites were chosen considering the prevailing wind direction (west southwesterly) such that during both sampling periods, one site was upwind of the power plant (those sites in Cheswick) and one site was downwind of the power plant (the site in Springdale). However, it must be noted that the wind during the sampling period was highly variable.

Each experimental site was equipped to measure $PM_{2.5}$ and PM_{10} using Reference Ambient Air Sampler (RAAS) high volume air samplers by Anderson Instruments as well as meteorological data measured by Met One Model Automet portable weather stations equipped with an onboard data logger. The sampling period for each instrument was set to 24 h, and the air volumetric flow on both types of instruments was set at the standard flow rate of 16.67 Liters/minute. The sampling filters used to collect both $PM_{2.5}$ and PM_{10} were 47 mm diameter Teflon filters. For $PM_{2.5}$ sampling, there was an additional smaller filter used (~39 mm diameter) in the WINS impactor, which was wetted

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with impactor oil, in order to restrict particles larger than 2.5 micron from reaching the 47 mm Teflon filter. All the filters were pre-tared and numbered.



Figure 1. Location of monitoring stations in relation to the Cheswick Power Plant and the two boroughs of Springdale and Cheswick.

High Volume (Hi Vol) 47 mm MTL Teflon filters were used to analyze mass concentration of PM_{10} while Hi Vol 39 mm Teflon filters were used to analyze mass concentration of $PM_{2.5}$. Filter weighing measurements took place inside a temperature/relative humidity-controlled ISO Class 6 (<1000 $PM_{0.5}$ /cf) clean room employing a draft-shielded microbalance with anti-static wand. Temperature and relative humidity were controlled to 21 °C and 35%, respectively. The filter mass was determined through five weightings of the filter, with each weight bracketed by a reading of the internal zero of the balance. A buoyancy correction was applied to the mass and the average of the zeros bracketing the mass was then subtracted from the result. This was repeated four additional times, with the average of these five results being reported as the final mass. The gravitational analysis was performed by the Research Triangle Park's (RTP) office of Applied Research Associates, Inc. (ARA).

Ten samples were chosen from this study to undergo an inorganic analysis, i.e., five for $PM_{2.5}$ and five for PM_{10} . In addition, two sample blanks were analyzed for calibration. Samples were analyzed from all three monitoring locations. The samples were selected to represent a mix of particulate matter concentrations collected during times when the plant was running. The inorganic analysis was then performed by X-Ray Fluorescence (XRF) and then Ion Chromatography (IC) analysis.

2.2. PA PM_{2.5} Monitoring Sites

In addition to the measurement sites used in the study, data from four monitoring stations run by the state of PA (PA Department of Environmental Protection) was also used to compare against measurements taken in the field campaign. The four monitoring sites used in this study (Figure 2) were Beaver Falls (40.7478° N, 80.3157° W), Florence (40.4454° N, 80.4212° W), Greensburg (40.3043° N, 79.5060° W) and Washington (40.1706° N, 80.2617° W). These four sites contain continuous Met One BAM 1020 PM_{2.5} monitors [20]. It is important to note that while the NCSU monitoring sites were located fairly close to the Cheswick Power Plant (i.e., less than 5 miles), the PA DEP monitoring sites were located 20–40 miles from the power plant.

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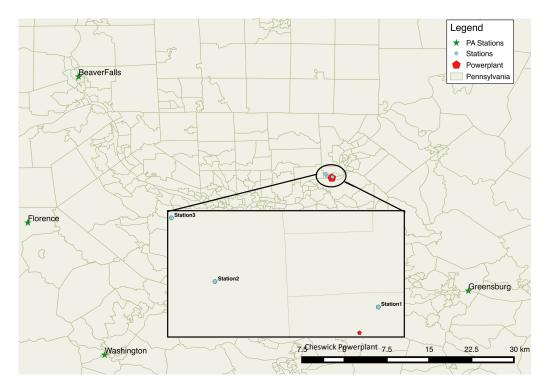


Figure 2. Location of the PA DEP monitoring sites in relation to the NCSU monitoring stations. The PA DEP monitoring sites are represented by the green stars while the NCSU monitoring sites are represented by the blue dots.

3. Results and Discussion

The observed average 24-hour concentrations of both $PM_{2.5}$ (0.02–22.76 μg m⁻³) and PM_{10} (0.9–44.15 μg m⁻³) from the NCSU monitoring sites were found to be lower than the US EPA National Ambient Air Quality Standard of 35 μg m⁻³ and 150 μg m⁻³, respectively (Figure 3).

The average concentrations of PM₁₀ observed during the periods were 20.5 \pm 10.2 μg m⁻³ (Station 1), $16.1 \pm 4.9 ~\mu g~m^{-3}$ (Station 2) and $16.5 \pm 7.1 ~\mu g~m^{-3}$ (Station 3). The average concentrations of PM_{2.5} observed at the stations were 9.1 \pm 5.1 μg m⁻³ (Station 1), 0.2 \pm 0.4 μg m⁻³ (Station 2) and 11.6 \pm 4.8 μg m⁻³ (Station 3). Station 1 observed the maximum concentrations of both PM_{2.5} (22.76 $\mu g m^{-3}$) and PM₁₀ (44.15 $\mu g m^{-3} \mu g m^{-3}$). While Station 3 also observed higher concentrations of particulate matter (both $PM_{2.5}$ and PM_{10}), the results of a t test (at $\alpha = 0.05$) indicated that the concentrations at Station 1 were indeed higher than those observed at Stations 2 and 3. There are several plausible causes for this: the location of Station 1 versus Stations 2 and 3 in relation to the meteorological conditions as well as additional potential sources of particulate matter into the atmosphere (i.e., traffic emissions, roadway dust, transport). Concentrations of particulate matter tended to be higher when conditions were warm and sunny. Conversely, the lowest concentrations of particulate matter were primarily observed during rainy conditions. This was expected due to the removal of the particulates from the atmosphere through wet deposition. However, there were a few days where elevated concentrations were observed when the plant was running during rainy conditions. This may have resulted from the stability of the atmosphere in these conditions, which allows for the atmospheric accumulation of particulate matter. Nevertheless, comparisons of particulate matter concentrations with meteorological conditions (Figures 4 and 5) failed to show any strong correlations.

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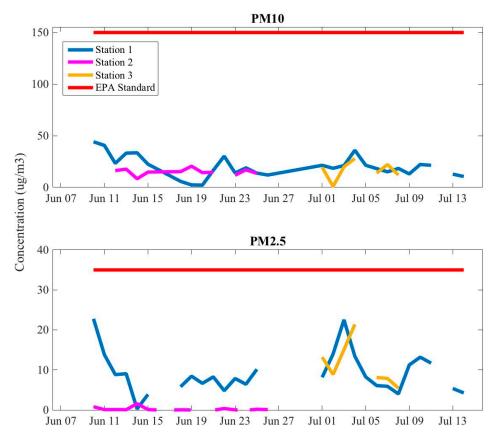


Figure 3. The daily 24-hour concentration of PM_{10} and $PM_{2.5}$ at Station 1 (blue), Station 2 (magenta) and Station 3 (gold) plotted against the US EPA National Ambient Air Quality Standards value for 24-hour average PM_{10} concentration (red).

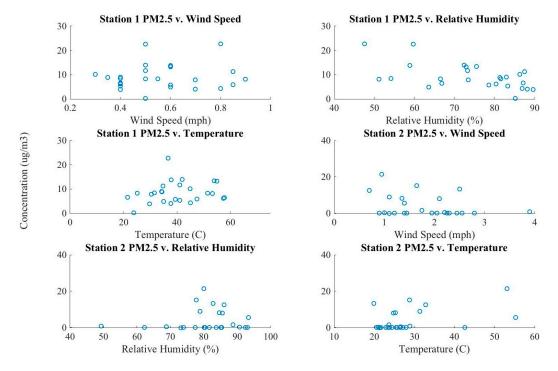


Figure 4. Comparing concentrations of PM_{2.5} with meteorological variables for both Station 1 and Stations 2 and 3, denoted by Station 2 in this figure.

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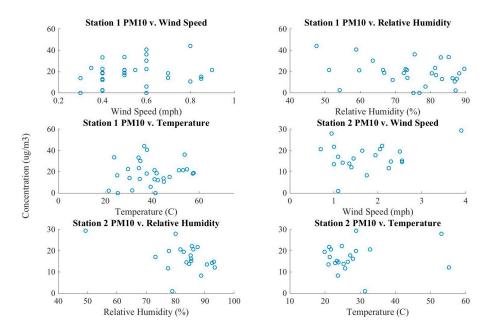


Figure 5. Comparing concentrations of PM_{10} with meteorological variables for both Station 1 and Stations 2 and 3, denoted by Station 2 in this figure.

When comparing the median wind speed versus the concentration of PM_{10} at Stations 1 and 2, positive correlations of 0.2 and 0.27, respectively, were observed. When comparing the median relative humidity with the concentration of PM_{10} , negative correlations of 0.3 and 0.44 were observed, respectively, while minor positive correlations of 0.08 and 0.12 were observed for temperature and concentration at Stations 1 and 2, respectively. When comparing meteorological conditions with concentrations of $PM_{2.5}$ at Stations 1 and 2, a negative correlation of 0.32 between wind speed and concentration of $PM_{2.5}$ was observed at Station 2 while Station 1 observed a positive correlation of 0.12. The strongest meteorological correlation was a negative correlation of -0.56 between relative humidity and the concentration of $PM_{2.5}$ at Stations 1, which is nominally downwind from the plant. In contrast, the correlation between relative humidity and $PM_{2.5}$ was weakly positive, at 0.09, for Station 2. Positive correlations of 0.54 and 0.46 were observed when comparing temperature versus concentrations of $PM_{2.5}$ at Stations 1 and 2, respectively. However, when considering these correlations, it is important to acknowledge the short sample period and thus small sample.

The median wind direction at upwind versus downwind stations was compared with the average 24-hour concentrations of particulate matter, where the categorizations of upwind and downwind were denoted based on the daily median wind direction. The results of a t test provide p=0.5 and p=0.4, for concentrations of PM $_{10}$ and PM $_{2.5}$, respectively, which suggests that the relationship is not statistically significant.

Despite the fact that the median wind direction was not significantly correlated with the concentrations of particulate matter, the local hourly wind direction likely did impact concentrations of particulate matter. Furthermore, it is likely that some of the relatively high concentrations of both particulate matter observed at Station 1 were found to have come from the direction of the coal site on the power plant's property (Figures 6 and 7). In addition to observing elevated concentrations coming directly from the plant, elevated concentrations of both $PM_{2.5}$ and PM_{10} were also observed coming from other directions. This can likely be attributed to a number of sources, such as emissions from vehicles, road dust. In addition, the elevated concentrations can also be attributed to transport due to local and regional meteorological phenomena. Furthermore, fairly elevated concentrations of PM_{10} are observed at faster wind speeds. While wind speeds would generally reduce ambient concentrations of particulate matter in the atmosphere, the observed elevated concentrations could simply be due to an emission source located extremely close to the monitoring site (i.e., a stalled truck on the side of the road).

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Figure 6. Wind rose created for $PM_{2.5}$ concentrations measured at Station 1. Green colors represent the highest concentrations of particulate matter while the orange and red colors represent lower concentrations of particulate matter, as shown in the legend. Units of particulate matter concentration are $\mu g \ m^{-3}$. The upper side of the wind rose represents North. The length of the wind rose vectors represents average wind speed, where the longer vectors represent a higher average wind speed.

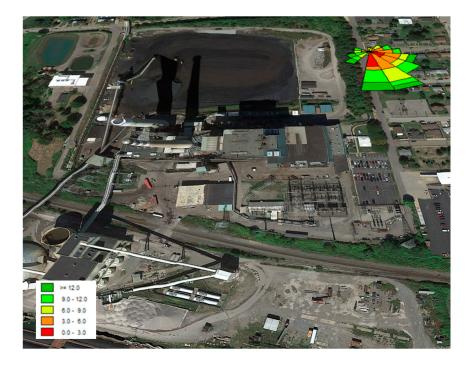


Figure 7. Wind rose created for PM_{10} concentrations measured at Station 1. Green colors represent the highest concentrations of particulate matter while the orange and red colors represent lower concentrations of particulate matter, as shown in the legend. Units of particulate matter concentration are $\mu g \, m^{-3}$. The upper side of the wind rose represents North. The length of the wind rose vectors represents average wind speed, where the longer vectors represent a higher average wind speed.

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In contrast to this, neither Stations 2 nor 3 (not pictured) showed extremely high concentrations coming from the power plant. However, this is not entirely surprising for two reasons: the first reason is that the winds were primarily such that those stations were upwind of the power plant and that both of these stations were twice as far as Station 1 for the plant, thus allowing for some dispersion of pollutants emitted from the plant and the coal site to occur.

In addition to this, concentrations of both $PM_{2.5}$ and PM_{10} were compared at calm wind conditions (median wind speeds less than 1 kt) and non-calm wind conditions (Table 1). When comparing average concentrations of $PM_{2.5}$ during calm winds ($9.29 \pm 5.59 \, \mu g \, m^{-3}$) with average concentrations during not calm winds ($3.53 \pm 5.18 \, \mu g \, m^{-3}$), it is clear that the average concentrations during calm winds are higher than the average concentrations during not calm winds. Similarly, average concentrations of PM_{10} observed during calm winds ($19.83 \pm 10.29 \, \mu g \, m^{-3}$) was higher than the average concentrations observed when the winds were not calm ($16.25 \pm 4.87 \, \mu g \, m^{-3}$). Two t tests were also conducted in order to determine whether or not the comparison between concentrations of $PM_{2.5}$ and PM_{10} during calm and not calm winds was statistically significant. The results of the t tests provide p = 0.052 and p = 0.001, for concentrations of $PM_{2.5}$ during calm and not calm winds is statistically significant, the comparison for concentrations of PM_{10} is not quite statistically significant.

Concentrations of particulate matter were also compared with the plant's gross load during the period (power plant gross load obtained via: EPA. Available online: http://www.ampd.epa.gov/ampd). Figure 8 compares the Cheswick Power Plant's daily gross load with the daily 24-hour average concentration of PM_{10} and $PM_{2.5}$. When comparing the daily gross load with the particulate matter concentrations, the correlation coefficients (ranging between -0.23 and 0.15) suggest that the concentrations of particulate matter at each station are not statistically correlated with the plant gross load.

Table 1. Comparison of concentrations of $PM_{2.5}$ and PM_{10} (µg m⁻³) during both calm and not calm wind conditions. Calm wind conditions are considered to be conditions where the median wind speed is less than 1 kt.

	Calm Winds (µg m ⁻³)		Not Calm Winds (μg m ⁻³)	
	PM _{2.5}	PM_{10}	PM _{2.5}	PM_{10}
Maximum	22.76	44.15	15.05	29.31
Minimum	0.16	0.91	0.02	8.30
Mean	9.29	19.83	3.53	16.25
Median	8.26	20.55	0.35	15.03
Standard Deviation	5.59	10.29	5.18	4.87
Number of Samples	31	33	15	16

Five PM_{10} samples and five $PM_{2.5}$ samples were subjected to XRF and IC analysis. The XRF analysis represents the PM samples as ions. When comparing the results of the analyses, it is evident that they are consistent. The results of the XRF analysis (Figure 9) showed that the primary constituents of PM_{10} are sulfur (0.66–1.77 μg m⁻³), silicon (0.14–2.47 μg m⁻³), aluminum (0.03–1.06 μg m⁻³) and iron (0.08–0.95 μg m⁻³). Similarly, the primary constituents of $PM_{2.5}$ were found to be sulfur (0.41–1.17 μg m⁻³), silicon (0.02–0.23 μg m⁻³) and iron (0.03–0.06 μg m⁻³). Particulate matter is typically composed of a complex mixture of chemicals that are strongly dependent on source characteristics. Inorganic analysis of the particulate matter revealed the presence of antimony, arsenic, beryllium, cadmium, chromium, cobalt, lead, manganese, mercury, nickel and selenium (Figure 9). All of these metals present in the PM samples are known to be present in coal [21]. While these elements are all found within coal and coal ash, they are also present in crustal material. Based on the small number of samples analyzed, it was not possible to discern differences between the upwind and downwind samples. The highest concentration of sulfur observed in a sample during the period was observed at Station 1. In this case, it is possible that

this can be attributed to the coal pit, which is located less than half of a mile away from the station. However, it is also important to note that there are other coal-fired power plants located within 50 miles of the Cheswick Power plant, which could also contribute to the elevated sulfur concentrations.

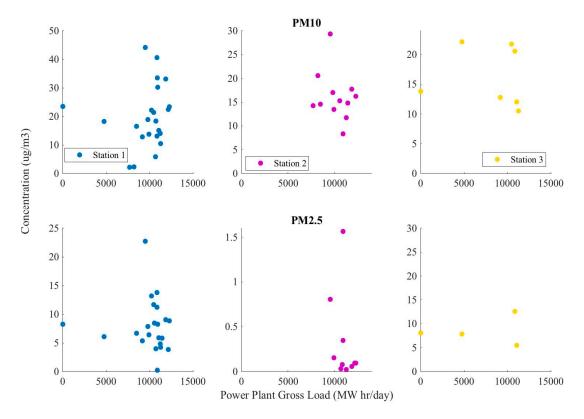


Figure 8. Comparing concentrations of PM_{10} and $PM_{2.5}$ at each station with the power plant gross load. The blue dots represent the concentrations at Station 1, the magenta dots represent the concentrations at Station 2 and the yellow/gold dots represent the concentrations at Station 3.

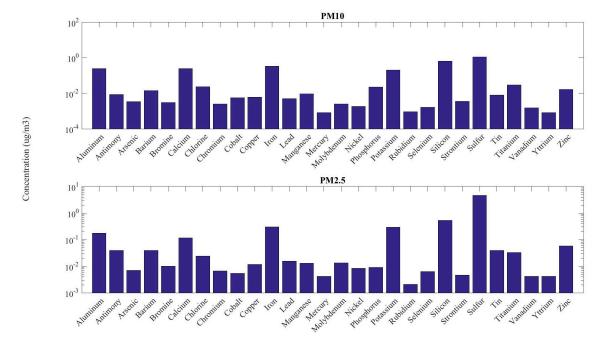


Figure 9. The results of the XRF analysis plotted on a log scale.

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The results of the IC analysis (Figure 10) showed that the dominant species of particulate matter in this region are primarily sulfate, nitrate and ammonium. Within the PM $_{2.5}$ samples analyzed, concentrations of sulfate ranged from 2.31 μg m $^{-3}$ to 3.08 μg m $^{-3}$, nitrate concentrations ranged from 0.03 μg m $^{-3}$ to 0.58 μg m $^{-3}$, and concentrations of ammonium ranged from 0.82 μg m $^{-3}$ to 1.15 μg m $^{-3}$. Within the analyzed PM $_{10}$ samples, concentrations of sulfate ranged from 1.05 μg m $^{-3}$ to 4.73 μg m $^{-3}$, concentrations of nitrate ranged from 0.01 μg m $^{-3}$ to 0.22 μg m $^{-3}$ and concentrations of ammonium ranged from 0.38 μg m $^{-3}$ to 1.69 μg m $^{-3}$. As described above, it was not possible to discern differences between the upwind and downwind samples based on the small sample size. It is also important to note that sulfate, nitrate and ammonium are also common constituents of particulate matter and thus cannot be attributed entirely to the power plant.

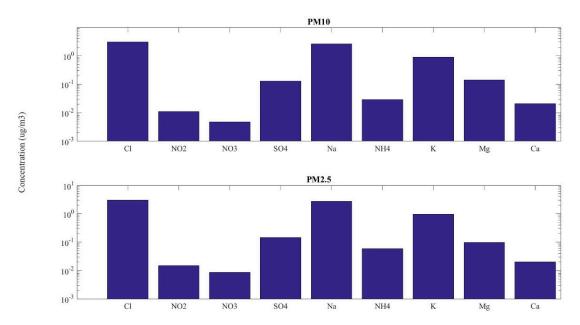


Figure 10. The results of the IC analysis, plotted on a log scale.

The concentrations of $PM_{2.5}$ measured by four PA Department of Environmental Protection air quality monitors were also analyzed in this study. The observed average 24-hour concentrations at these sites were $12.7 \pm 6.9~\mu g~m^{-3}$ (Beaver Falls), $11.2 \pm 4.7~\mu g~m^{-3}$ (Florence), $12.2 \pm 5.3~\mu g~m^{-3}$ (Greensburg) and $12.2 \pm 5.5~\mu g~m^{-3}$ (Washington). Similar to what was observed at the NCSU monitoring sites, the PA DEP monitors also observed 24-hour average concentrations of fine particulate (2–28.2 $\mu g~m^{-3}$) matter below the US EPA 24-hour NAAQS (35 $\mu g~m^{-3}$), with the exception of the Beaver Falls monitoring stations, which observed a concentration of 42.6 $\mu g~m^{-3}$ on 4 July 2015 (Figure 11). However, this elevated concentration can likely be explained by the presence of fireworks and other combustion processes due to the holiday.

When comparing the average PA DEP monitoring station $PM_{2.5}$ concentrations [20] with the average NCSU monitor $PM_{2.5}$ concentrations (Figure 12), it is evident that the monitors used in this study observed lower concentrations than the concentrations observed at the PA DEP monitoring sites. However, the general trend in concentrations for both monitoring networks is the same. There are several possible explanations for this discrepancy. One potential cause for the differences in observations could be due to differences in the instrumentation used to measure the particulate matter. In addition, another potential cause for the observed differences could be due to location. Since the monitors are farther away, it is possible that the pollutants were transported aloft before mixing down to the surface level. Furthermore, it is also possible that there are other potential sources of particulate matter located near the sites.

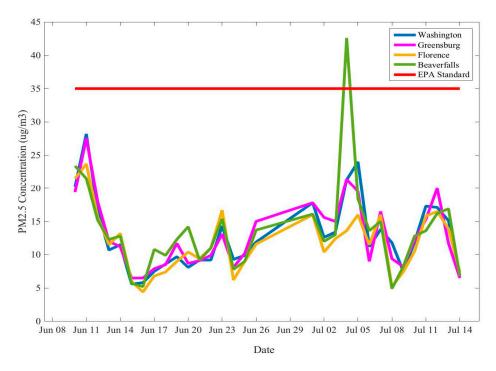


Figure 11. $PM_{2.5}$ measurements taken by the PA DEP monitoring sites during the study period. The red line represents the US EPA 24-hour NAAQS, the blue line represents the Washington monitoring site, the magenta represents the Greensburg monitoring site, the gold line represents the Florence monitoring site and the green line represents the Beaver Falls monitoring site.

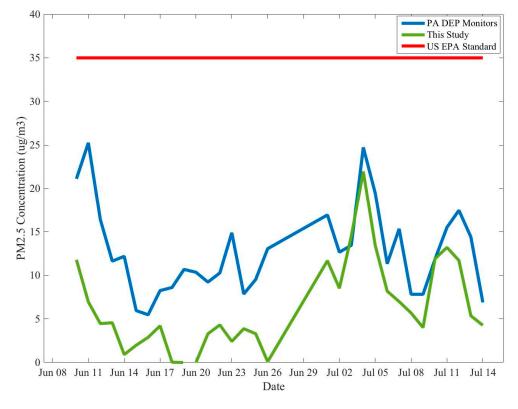


Figure 12. The average concentration measured by the NCSU monitors (this study, green line) compared with the average concentration measured by the PA DEP monitors (blue line line) and the US EPA 24-hour NAAQS (red line) for $PM_{2.5}$.

4. Conclusions

The ambient 24-hour average concentrations of both $PM_{2.5}$ (measured in this study as well as by the PA DEP) and PM_{10} did not reach levels higher than what is permitted by the US EPA 24-hour NAAQS. The average concentrations of PM_{10} observed during the periods were $20.5 \pm 10.2~\mu g~m^{-3}$ for Station 1, $16.1 \pm 4.9~\mu g~m^{-3}$ for Station 2 and $16.5 \pm 7.1~\mu g~m^{-3}$ for Station 3. The average concentrations of $PM_{2.5}$ observed at the stations were $9.1 \pm 5.1~\mu g~m^{-3}$ for Station 1, $0.2 \pm 0.4~\mu g~m^{-3}$ for Station 2 and $11.6 \pm 4.8~\mu g~m^{-3}$ for Station 3. The highest average concentration was observed at Station 1 for PM_{10} and at Station 3 for $PM_{2.5}$. However, Station 1 observed the highest daily average concentration for both $PM_{2.5}$ and PM_{10} . Not only was Station 1 the closest to the power plant, but the wind rose analysis showed that some of the elevated concentrations potentially came directly from the power plant's coal pit, in addition to other PM sources. However, other elevated concentrations were observed coming from other directions, suggesting that there are several sources for particulate matter located in the region.

The IC analysis showed that the dominant species of particulate matter were primarily sulfate, nitrate and ammonium. The results of the XRF analysis showed that the primary constituents of PM_{10} and $PM_{2.5}$ were sulfur, silicon, aluminum and iron. While these are all constituents that are observed from coal combustion emissions, these constituents are also prominent in most particulate matter speciation and therefore cannot be contributed directly to emissions from the Cheswick Power Plant.

The low particulate matter emissions and results of the speciation analyses could be attributed to a number of factors. Because the study period was so short, it is likely that the local scale meteorological conditions led to bias in the results. In addition, it is also possible that there were errors in the instrumentation that led to such low concentrations, particularly for $PM_{2.5}$ concentrations observed at Station 2. Furthermore, the low concentrations of particulate matter observed at the sites could also be attributed to the emission control technology that is added to the Cheswick power plant, which includes wet lime flue gas desulfurization, low NO_x burner technology with separated over fire air selective catalytic reduction and an electrostatic precipitator (Available online: http://www.ampd.epa.gov/ampd), which would explain the reduced concentrations of pollutants observed in this study.

Based on the results of this study, is not possible to determine a concrete conclusion on the role of the Cheswick Power Plant in concentrations of particulate matter in the Cheswick and Springdale boroughs. These inconclusive results can be attributed to a number of factors, including unfavorable meteorological conditions, potential issues with the measurement equipment as well as an extremely short sampling period. It must be emphasized that this study was conducted over a limited time period. Therefore, it is recommended that further work be done on this matter, with longer sampling periods occurring in each season in order to capture a seasonal profile of concentrations of particulate matter in this region.

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Author Contributions: Viney Aneja was the PI for this project. Priya Pillai led the measurement process. Pornpan Uttamang helped take measurements for the project and helped in research for the manuscript. William Battye helped in the data analysis and in the revision and improvement of the manuscript. Casey Bray assisted in the measurement process, analyzed the data and wrote the manuscript.

Conflicts of Interest: The authors report no conflict of interest.

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