

Article



# Future-Year Ozone Isopleths for South Coast, San Joaquin Valley, and Maryland

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**Abstract:** Many areas of the United States are working toward achieving the 2015 ozone National Ambient Air Quality Standard (NAAQS) attainment level. The objective of this study was to develop future-year (2030) volatile organic compounds and nitrogen oxides (VOC-NO<sub>x</sub>) isopleth diagrams of the 4th highest maximum daily 8-h average ozone design value concentrations at monitors of interest in the South Coast Air Basin (SoCAB) and San Joaquin Valley (SJV) in California, and in Maryland. The simulation results showed there would be attainment of the 2015 ozone NAAQS in 2030 without further controls at the selected monitors: 27% in SoCAB, 57% in SJV, and 100% in Maryland. The SoCAB ozone isopleths developed in this study were compared with those reported in the South Coast Air Quality Management District 2016 Air Quality Management Plan. There are several differences between the two modeling studies, the results are qualitatively similar for most of the monitors in the relative amounts of additional emission reductions needed to achieve the ozone NAAQS. The results of this study provide insight into designing potential control strategies for ozone attainment in future years for areas currently in non-attainment. Additional photochemical modeling using these strategies can then provide confirmation of the effectiveness of the controls.

**Keywords:** ozone; attainment; isopleth; empirical kinetic modeling approach; EKMA; South Coast Air Basin; San Joaquin Valley; Maryland

## 1. Introduction

Many areas of the United States (U.S.) are trying to achieve the current National Ambient Air Quality Standards (NAAQS) attainment level for ozone. Ozone formation is linked to emissions of volatile organic compounds (VOCs) and nitrogen oxides (NO<sub>x</sub>) [1]. The chemistry leading to ozone production from emissions of VOC and NO<sub>x</sub> is non-linear [2]. Contour plots of ozone versus NO<sub>x</sub> and VOC emissions (ozone isopleths) are often employed to illustrate the non-linear response of ozone levels to NO<sub>x</sub> and VOC changes, and have been used in the past in the empirical kinetic modeling approach (EKMA) developed by the U.S. Environmental Protection Agency (EPA) to develop control strategies for ozone reduction [3]. Thus, the isopleths of ozone versus VOC and NO<sub>x</sub> emissions are often referred to as EKMA diagrams. The EKMA generates ozone isopleths for specific monitors in a city. Once the maximum measured ozone concentration at a monitor has been identified, the VOC and NO<sub>x</sub> reductions needed to achieve the NAAQS level are determined using the EKMA from the distance along the VOC and NO<sub>x</sub> axes to the isopleth that represents the desired peak ozone concentration mandated by the NAAQS. The abundance of VOCs relative to NO<sub>x</sub> is characterized by the VOC/NO<sub>x</sub> ratio. The VOC/NO<sub>x</sub> affect ozone concentrations [4].

While the EKMA for regulatory planning has been superseded by photochemical grid modeling, the ozone isopleths are useful for understanding how ozone levels at a particular location may respond to reductions in NO<sub>x</sub>, VOC, or both. The objective of this study was to develop future-year (2030) VOC-NO<sub>x</sub> isopleth diagrams of the 4th highest maximum daily 8-h average (H4MDA8) ozone design value concentrations (DVC) at monitors of interest in the South Coast Air Basin (SoCAB) and San Joaquin Valley (SJV) in California, and in Maryland. These areas were selected to illustrate contrasting current ozone classification levels: SoCAB and SJV at extreme or severe [5], and Maryland at moderate or marginal [6]. Regions where ozone production is limited by the presence of nitrogen oxides or hydrocarbons are readily apparent in such a diagram [7]. The Comprehensive Air Quality Model with Extensions (CAMx) version 6.4 [8] is the latest version of the photochemical grid model (PGM) and was used to conduct several brute force scenarios. The CAMx allows for integrated "one-atmosphere" assessments of tropospheric air pollution (ozone, particulates, air toxics) over spatial scales ranging from neighborhoods to continents. It is a "state-of-the-science" open-source system that is computationally efficient, flexible, and publicly available. Meteorological fields are supplied to the CAMx from separate, commonly used weather prediction models such as the Weather Research and Forecasting (WRF) model. All emission inputs are supplied from external pre-processing systems. The CAMx simulates the emission, dispersion, chemical reaction, and removal of pollutants by marching the Eulerian continuity equation forward in time for each chemical species in a system of nested three-dimensional grids. The ozone isopleths developed in this study using the CAMx were compared with the South Coast Air Quality Management District (SCAQMD) 2016 Air Quality Management Plan (AQMP) [9] isopleths, which were generated with the U.S. EPA photochemical grid model, the Community Multiscale Air Quality (CMAQ) [10,11].

#### 2. Method

In previous modeling studies, Collet et al. [12,13] simulated the entire contiguous U.S. (continental U.S. or CONUS) domain (396 vertical by 246 horizontal grid cells) at 12-km resolution with the CAMx 6.4. For this study, a different approach was adopted because the development of ozone isopleths relating ozone responses to  $NO_x$  and VOC emissions changes required several tens of emissions scenario simulations to be conducted. This approach, referred to as brute force sensitivity analyses, is impractical for the large CONUS domain, since it is time- and resource-intensive. Since the objective of this study was to develop ozone isopleth diagrams for specific sub-regions of the country, sub-regional CAMx simulations were conducted to achieve the study objective. Specifically, the emissions scenario simulations were conducted for two subdomains using 12 km by 12 km grid resolution, one in the Western U.S. and the second in the Eastern U.S.

Figure 1 shows the two subdomains within the outer CONUS domain. The western subdomain (66 vertical by 105 horizontal grid cells) covers most of California and includes the SoCAB and SJV, while the eastern subdomain (40 vertical by 29 horizontal grid cells) covers Maryland and portions of the surrounding states.

The existing CAMx modeling databases used by Collet et al. [13] were developed for the CONUS domain. For that study, the 2025 CMAQ-ready emissions provided by the EPA were used for the future-year (2030) emissions. Future-year emissions from the natural source categories were assumed to be the same as the 2011 base-year emissions. The 2030 emissions for all anthropogenic source categories, except on-road mobile sources, were assumed to be the same as the EPA's 2025 emissions case. The on-road mobile emissions for 2030 were developed using MOVES 2014 and the California Air Resources Board's (CARB) EMFAC2014 models. It was necessary to develop new inputs for the two subdomains shown in Figure 1. The inputs for this study consist of the following: meteorology, emissions, land use, ozone column and photolysis rates, and boundary conditions. The gridded subdomain meteorological fields and land use data were developed by extracting ("windowing") the subdomain data from the corresponding CONUS files. Ozone column and photolysis rate data were developed as new inputs for the subdomains. Boundary conditions for the subdomains were

developed by first conducting base-year (2011) and future-year (2030) CONUS domain simulations in which the 3-dimensional (3-D) concentration outputs were saved, and then windowing these outputs for the subdomains.



**Figure 1.** Western (blue outline) and eastern (red outline) subdomains within the 12-km resolution continental U.S. (CONUS) domain.

The 2030 future-year emissions for the CONUS domain from the Collet et al. [13] study were windowed for the two subdomains in the Western and Eastern U.S. In that study, the 2030 future-year emissions were based on the 2025 projected emissions from the EPA 2011v6.2 platform and 2030 on-road mobile emissions generated using the latest on-road emission models. To perform the emission scenario simulations required for generating the ozone isopleths, the emissions within the sub-regions of interest (the two air basins in the western subdomain and the state of Maryland in the eastern subdomain) were adjusted. These adjustments were made using cell masks created by ArcGIS [14] that only mark the cells within the two air basins and the Maryland state boundary. A set of reduction factors (20%, 40%, 60%, and 80%) was applied to emissions of NO<sub>x</sub> (NO, NO<sub>2</sub>, HONO), anthropogenic VOCs (i.e., excluding isoprene, monoterpene, and sesquiterpene emissions), and CO in grid cells identified by the cell masks, resulting in a total of 25 emission scenarios, including the base emission scenario (0% reduction), for each subdomain.

The air quality modeling was conducted using the most recent version (6.4) of the CAMx with the CB6r4 chemical mechanism. The base-year (2011) and future-year (2030) simulations for the CONUS domain were conducted with the CAMx to develop boundary conditions for the 2011 and 2030 subdomain simulations. A 10-day spin-up period was used for the two CONUS domain simulations.

A total of 52 CAMx simulations were conducted for the ozone season (1 May to 30 September) for the two subdomains. These simulations consisted of a base-year simulation and 25 future-year simulations for each of the two subdomains. The base-year results are required to calculate future-year ozone design values for the California air basins and the state of Maryland using EPA's Modeled Attainment Test Software (MATS). The base-year subdomain simulations and the future-year subdomain simulations used a 5-day period for spin-up. The 25 future-year emission scenario calculations correspond to a 5 by 5 matrix of future-year emissions (20%, 40%, 60%, 80%, and 100% NO<sub>x</sub> emissions by 20%, 40%, 60%, 80%, and 100% VOC and CO emissions).

#### 3. Results

Following EPA guidance, the outputs of the base-year and future-year subdomain simulations were processed using EPA's MATS to project future-year design values (DVF) at monitoring locations of interest. Briefly, the modeling results are used in a relative sense to scale observed site-specific 8-h ozone concentrations for current-year design values (DVC) based on the relative changes in the modeled 8-h ozone concentrations between the current and future years. The model-derived scaling factors are called relative response factors (RRFs), and are based on the relative changes in the modeling results between the current-year (2011) base case and the future-year (2030) emission scenarios. This is the recommended regulatory approach to determine future-year ozone design values, and is based on the assumption that the model can predict ozone responses to changes in emissions more accurately than it can predict absolute ozone concentrations.

Tables 1 and 2 show the ozone NAAQS attainment status (70 ppb H4MDA8) for the three study areas without any additional controls. Table 1 provides the list of monitors in the two California air basins and state of Maryland where future-year ozone design values were calculated using the air quality modeling results and MATS, and the future-year ozone design values at these monitors. Table 2 quantifies the number of monitors in Table 1 which are in or not in attainment.

Air Basin	Monitor ID	Monitor Name	2030 DVF (ppb)
SoCAB	Los Angeles_0016	Glendora-Laurel	83.4
SoCAB	Los Angeles_0113	West Los Angeles-VA Hospital	58.7
SoCAB	Los Angeles_1103	Los Angeles-North Main Street	58
SoCAB	Los Angeles_2005	Pasadena-S Wilson Avenue	70.7
SoCAB	Los Angeles_6012	Santa Clarita	84.3
SoCAB	Orange_0007	Anaheim	58.6
SoCAB	Riverside_0012	Banning Airport	82.8
SoCAB	Riverside_6001	Perris	78.2
SoCAB	San Bernardino_0005	Crestline	94.8
SoCAB	San Bernardino_1004	Upland	85.7
SoCAB	San Bernardino_2002	Fontana-Arrow Highway	89.1
SJV	Fresno_5001	Clovis-N Villa Avenue	76.3
SJV	Kern_0232	Oildale-3311 Manor Street	70.2
SJV	Merced_0003	Merced-S Coffee Avenue	67.3
SJV	San Joaquin_1002	Stockton-Hazelton Street	57.8
SJV	Stanislaus_0005	Modesto-14th Street	63.4
SJV	Tulare_0009	Sequoia and Kings Canyon Natl Park	78.4
MD	Cecil_0003	Fair Hill Natural Resource Management Area	60.5
MD	Harford_1001	Edgewood	68.1
MD	Baltimore_1007	Padonia	60.7
MD	Frederick_0037	Frederick Airport	59.4
MD	Anne Arundel_0014	Davidsonville	58.2
MD	Calvert_0011	Calvert	58.5

Table 1. List of monitoring sites selected for analysis and future-year design values (DVF) at these sites.

Table 2. Number of DVF attainment areas in the monitoring sites selected for analysis.

	Total Listad	DVF in Attainment	DVF Non-Attainment	
	Iotal Listed	Number (%)	Number (%)	
South Coast Air Basin (SoCAB)	11	3 (27%)	8 (73%)	
San Joaquin Valley (SJV)	7	4 (57%)	3 (43%)	
Maryland State (MD)	6	6 (100%)	0	

A matrix of 25 (5  $\times$  5) future-year emission scenarios for each subdomain provides 25 future-year ozone design values at the selected monitors as a function of  $NO_x$  and VOC emissions in the air basins of interest. This is a relatively coarse resolution, so it is necessary to use an interpolation technique, such as kriging, to obtain smooth contours of the ozone design values (ozone isopleths). Kriging is an advanced geostatistical procedure that generates an estimated surface from a scattered set of points. Although kriging was developed originally for applications in geostatistics, it is a general method of statistical interpolation that can be applied within any discipline to sampled data from random fields. It differs from simpler interpolation methods in that it uses the spatial correlation between sampled points to interpolate the values in the spatial field. Kriging has been used in other studies to develop ozone response surfaces [15,16]. PyKrige, a Python kriging library, was used for this purpose in the present study. The interpolation was performed using ordinary kriging, which is the most commonly used kriging algorithm and produces interpolation values by relying on an unknown mean value, allowing local influences due to nearby neighboring values. The original  $5 \times 5$  matrix of values were interpolated to  $81 \times 81$  data points, representing reductions in anthropogenic emissions in the California air basins and state of Maryland in increments of 1% from 0% to 80%. Standard Python libraries were then used to prepare the ozone isopleths.

Ozone isopleths ("EKMA diagrams") for the selected monitors in the South Coast Air Basin are shown in Figure 2, while Figure 3 shows ozone isopleths for selected monitors in Fresno and Tulare, CA, and the state of Maryland. The x-axis shows the VOC anthropogenic emissions, and the y-axis the NO<sub>x</sub> anthropogenic emissions. Thus, the y-axis and x-axis represent anthropogenic emissions ranging from 20% to 100% of the base future-year air basin anthropogenic emissions (i.e., each tick mark represents 20%). Therefore, the 2030 DVF is at the 100% NO<sub>x</sub> and 100% VOC point, in the top right corner of the diagram. The color scheme for the ozone design values ranges from dark brown (85 ppb and higher) to dark blue (55 ppb and lower). The light blue region represents a level of about 70 ppb and provides a measure of the reductions in NO<sub>x</sub> and/or VOC required to reach attainment at a specific monitor.



Figure 2. Cont.



**Figure 2.** Future-year ozone design value isopleths (ppb) for monitors in the South Coast Air Basin (SoCAB): (a) Glendora, (b) Los Angeles-North Main, (c) Pasadena, (d) Anaheim, (e) Banning, and (f) Crestline.



**Figure 3.** Future-year ozone design value isopleths (ppb) for monitors in (**a**) Fresno, CA; (**b**) Tulare, CA; (**c**) Baltimore, MD; and (**d**) Anne Arundel, MD.

### 4. Discussion

Table 3 provides the future-year projected NO<sub>x</sub> and VOC emission totals for the non-attainment areas, and the estimated NO<sub>x</sub> or VOC emission levels needed to attain the standards in tons/day and as a reduction percentage of the base future-year emissions. In the SoCAB, 3 of the 11 monitors (West Los Angeles, Los Angeles-North Main, and Anaheim) showed attainment of the 2015 ozone NAAQS level at projected 2030 SoCAB anthropogenic emissions without further controls. Table 3

shows that monitors in Glendora and Pasadena could meet the NAAQS with NO<sub>x</sub> or VOC reductions, while on a percentage basis, less NO<sub>x</sub> reductions are required than VOC reductions. Moving eastward in the SoCAB, VOC reductions to 80% will not reduce ozone to the NAAQS level. In Riverside, NO<sub>x</sub> reductions are projected to bring the area into compliance. The northwest and most eastern areas of the SoCAB (i.e., Santa Clarita, Reseda, Crestline, and San Bernardino) are not projected to achieve compliance with 80% NO<sub>x</sub> or VOC anthropogenic reductions. In the SJV area, four of seven monitors show attainment of the 2015 NAAQS level at projected 2030 anthropogenic emissions without further controls. To achieve attainment at the two monitors projecting non-attainment in the SJV (Fresno and Tulare), NO<sub>x</sub> reductions would be effective, while VOC reductions would not be useful. At all the Maryland monitors, future-year ozone levels are in attainment of the 2015 NAAQS level without further controls. The Maryland area is sensitive to changes in NO<sub>x</sub> and shows little to no sensitivity to VOC changes.

	NO <sub>x</sub>			VOC			
	2030 (Tons/Day)	70 ppb Ozone Attainment (Tons/Day)	% Reduction	2030 (Tons/Day)	70 ppb Ozone Attainment (Tons/Day)	% Reduction	
Los Angeles_0016, Glendora	264.3	85	68%	357	82	77%	
Los Angeles_2005, Pasadena	264.3	200	24%	357	300	16%	
Riverside_0012, Banning	264.3	111	58%	357	NA		
San Bernardino_0005, Crestline	264.3	NA		357	NA		
Fresno_5001	153.6	90	41%	541.7	NA		
Tulare_0009	153.6	100	35%	541.7	NA		

**Table 3.** List of non-attainment area monitoring sites in relation to future-year ozone design values at the projected  $NO_x$  and VOC levels, and  $NO_x$  or VOC level needed in tons/day and as a percentage of reduction to attain 70 ppb 4th highest maximum daily 8-h average (H4MDA8).

To compare this study's ozone isopleths to those reported in the SCAQMD 2016 AQMP [9], the modeling platform and domain are shown in Table 4 for both studies.

	2016 AQMP	Present Study		
Modeling Base Year	2012	2011		
Future Year	2031	2030		
Time Frame	May 1–30 September	1 May-30 September		
Chemical Transport Model	CMAQ version 5.0.2 SAPRC07 with version "c" toluene updates	CAMx 6.4 with CB6r4		
Meteorological Model	Meteorological Model WRF version 3.6 with updated land use			
Spin-up	5 days	10 days		
On-Road Emission	EMFAC2014	EMFAC2014		
Off-Road Emission	Single package Category Specific Calculation	MOVES2014		
Modeling Domain	Grid: 154 by 102 624 km by 408 km	Grid: 66 by 105 792 km by 1260 km		
Grid Resolution	4 km by 4 km	12 km by 12 km		
Vertical Grid	18 layers with 14 layers below 2000 m agl and 50 hPa as top boundary	25 vertical layers (layer-collapsed from 35 WRF layers)		
Boundary Conditions	Model for Ozone and Related Chemical Tracers (MOZART)	36 km from GEOS-Chem simulation		

**Table 4.** Numerical modeling platforms and domains for 2016 Air Quality Management Plan (AQMP) and this study.

Table 5 compares the additional reductions in future-year emission required to achieve attainment from the SCAQMD AQMP isopleths for 2031 and the 2030 isopleths from this study. For most of the

monitors, the difference in the amount of  $NO_x$  or VOC (in tons/day or percent reduction) needed to achieve the ozone NAAQS level from the two sets of isopleths is within +25%. The SCAQMD AQMP isopleths show more  $NO_x$ -limited trends than the present study, possibly due to differences between the inputs and models used in the two studies (see Table 4).

**Table 5.** NO<sub>x</sub> and VOC levels required to achieve future-year (2030 for this study and 2031 for the AQMP) 70 ppb ozone design values at various monitors in the SoCAB. The monitor groups are ordered from east to west with monitors in each group ordered from north to south. NA means not achievable. For this paper, the two studies are considered similar when the differences between the reductions required to achieve attainment are within +25%.

		NO <sub>x</sub>			VOC	
	Tons/Day (% Reduction)			Tons/Day (% Reduction)		
	Present Study	AQMP	Similar	Present Study	AQMP	Similar
Base (tons/day)	264	214	Yes	357	362	Yes
Santa Clarita	NA	130 (39%)	No	NA	NA	Yes
Reseda	NA	160 (25%)	No	NA	7 (98%)	Yes
West Los Angeles-VA Hospital	Attainment	Attainment	Yes	Attainment	Attainment	Yes
Los Angeles-North Main Street	Attainment	Attainment	Yes	Attainment	Attainment	Yes
Pico Rivera	Attainment	Attainment	Yes	Attainment	Attainment	Yes
Azusa	211 (20%)	150 (30%)	Yes	325 (9%)	200 (45%)	No
Glendora-Laurel	85 (68%)	83 (61%)	Yes	102 (71%)	NA	No
Pomona	150 (43%)	110 (49%)	Yes	234 (34%)	90 (75%)	No
La Habra	Attainment	Attainment	Yes	Attainment	Attainment	Yes
Anaheim	Attainment	Attainment	Yes	Attainment	Attainment	Yes
Upland	56 (79%)	88 (59%)	Yes	NA	NA	Yes
Fontana-Arrow Hwy	NA	80 (63%)	No	NA	NA	Yes
Riverside	80 (70%)	100 (53%)	Yes	NA	NA	Yes
Mira Loma	-	124 (42%)	-	-	40 (89%)	-
Lake Elsinore	200 (24%)	212 (1%)	Yes	180 (50%)	258 (29%)	Yes
Mission Viejo	Attainment	Attainment	Yes	Attainment	Attainment	Yes
Crestline	NA	90 (58%)	No	NA	NA	Yes
San Bernardino	75 (72%)	108 (50%)	Yes	NA	NA	Yes
Redlands	NA	85 (60%)	No	NA	NA	Yes
Banning	111 (58%)	104 (51%)	Yes	NA	NA	Yes

## 5. Conclusions

This study developed future-year (2030) VOC-NO<sub>x</sub> isopleths of H4MDA8 ozone design values at selected monitors in the SoCAB and SJV in California, and Maryland. Photochemical grid modeling for a large number of VOC and NO<sub>x</sub> emission reduction scenarios (from 0% to 80%) were conducted to develop the isopleths of ozone design values versus VOC and NO<sub>x</sub> emissions in each region. The latest version of the photochemical grid model, CAMx 6.4, was used in this modeling study.

The modeling results showed that only 27% of the selected monitors in the SoCAB would reach attainment of the 2015 ozone NAAQS level in 2030 without further controls. In the SJV, 57% of the monitors would reach attainment without further controls. All monitors in Maryland were projected to reach attainment of the 2015 ozone NAAQS level in 2030 at base projected emission levels. The ozone isopleths for the SoCAB and SJV were used to determine the amount of additional controls that would be required to attain the standards in 2030. This analysis showed that the areas in the western and central portions of the basins could achieve attainment with NO<sub>x</sub> or VOC reductions or a combination of both, while areas between the central and eastern locations could achieve attainment with NO<sub>x</sub> reductions, but VOC reductions are not useful. Monitors in the northwest and easternmost areas of the SoCAB are not predicted to achieve attainment with an additional 80%-reduction in future-year anthropogenic NO<sub>x</sub> or VOC. In the SJV, additional NO<sub>x</sub> reductions are effective in achieving attainment at the two monitors projected to be in non-attainment in 2030, while VOC reductions are not effective. The Maryland area is sensitive to changes in NO<sub>x</sub> and shows little to no sensitivity to VOC changes, but all monitors are projected to be in attainment without further controls in 2030.

The SoCAB ozone isopleths developed in this study were compared with those reported in the SCAQMD 2016 AQMP. While there are several differences between the two modeling studies, the results are qualitatively similar (within +25%) for most of the monitors in the relative amounts of additional NO<sub>x</sub> and/or VOC reductions needed to achieve the ozone NAAQS level. This study shows that monitors in five areas will not attain the 70 ppb NAAQS with up to 80% reductions in NO<sub>x</sub> or VOC, while the SCAQMD isopleths show that NO<sub>x</sub> controls are effective in bringing these areas into attainment.

The results from this study provide insight into designing potential control strategies for ozone attainment in future years in areas currently in non-attainment. Additional photochemical modeling using these strategies can then provide confirmation of the effectiveness of the controls.

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