



Article Influences of Dimethyl Phthalate on Bacterial Community and Enzyme Activity in Vertical Flow Constructed Wetland

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Abstract: Dimethyl phthalate (DMP), belonging to the family of Phthalate esters (PAEs), is a plasticizer and has been widely used in the world for many years. Nowadays, it has become a ubiquitous environmental pollutant and is listed as an environmental priority pollutant by China's Environmental Monitoring Center. The purpose of this study is to estimate the responses of the bacterial community and enzyme activity to DMP contamination in three vertical flow constructed wetlands (VFCW), namely the constructed wetland A (planted with Pennisetum sinese Roxb), constructed wetland B (planted with Pennisetum purpureum Schum.), and constructed wetland C (unplanted), respectively. The results showed that the relative percentages of some genera associated with nitrogen metabolism and the function of degrading aromatic hydrocarbons were increased by DMP contamination, such as Dechloromonas agitata, Pleomorphomonas sp., Denitratisoma oestradiolicum, Plasticicumulans lactativorans, Novosphingobium sp., Alicycliphilus denitrificans, and Thauera sp. Meanwhile, principal coordinate analysis (PCA) analysis showed that the addition of DMP divided 12 samples into two groups as followed: one was the DMP group containing a-1, a-2, b-1, b-2, c-1 and c-2 while the other was no DMP group including A-1, A-2, B-1, B-2, C-1 and C-2. It indicated that DMP was the main reason for this change. In addition, by monitoring the activity of substrate enzymes, the activity of urease, phosphatase, catalase, and invertase in the wetlands before and after the experiment, these were significantly higher in the upper layer than in the lower layer and maintained high activity. Ultimately, the average influent concentration of DMP in three VFCWs was 8.12 mg/L and the average removal efficiency of the effluent was over 90%. Our results suggested that DMP was an important factor affecting the microbial community structure of wetland and the upper layer of the VFCW was the main site for the degradation of DMP. VFCW has great potential for the removal of the high concentration of DMP and it can be a good choice for the treatment of PAEs.

Keywords: dimethyl phthalate (DMP); bacterial community; enzyme activity; vertical flow constructed wetland (VFCW)

1. Introduction

Constructed wetland (CW) is a low-cost, eco-technology and a prospective wastewater treatment technology, which relies on biological, physical, and chemical processes to remove various pollutants [1]. In the past few years, vertical flow constructed wetland (VFCW) has become a preferable choice for wastewater treatment [2]. It showed good performance in removing nitrogen [3] and organic compounds [4,5] due to their excellent oxygen transfer characteristics and small area. Microorganisms play a key role in the



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Copyright: © 2021 by the authors. Licensee MDPI, Basel, Switzerland. This article is an open access article distributed under the terms and conditions of the Creative Commons Attribution (CC BY) license (https:// creativecommons.org/licenses/by/ 4.0/). treatment of pollutants in CW. In recent years, the related research studies on CW microorganisms mainly include the diversity of microorganisms [6], microbial enzyme activity [7], and the influence of environmental conditions (temperature, pH, etc.) on microbial population and activity [8]. Some researchers propose that the number of microorganisms, bacterial diversity, and the activities of enzymes in the wetland are closely relevant to the removal of pollutants [9–11]. The traditional research methods for the microbial diversity of CW are mainly to separate, cultivate, and identify the microbial species through pure culture technology, but 85~99% of the microorganisms in nature are not pure culture, so the traditional research methods may cause the wetland microbial diversity to be seriously underestimated [12]. In order to avoid the limitations of traditional methods, molecular biological analysis techniques, represented by polymerase chain reaction temperature gradient gel electrophoresis (PCR-TGGE) [13], polymerase chain reaction denaturing gradient gel electrophoresis (PCR-DGGE) [14,15], DNA molecular fingerprinting technology [16], and high-throughput sequencing technology [17], have been widely used in the study of microbial diversity in constructed wetlands in the field of environmental science. Phthalate esters (PAEs) have many types, complex properties, are mostly difficult to degrade, are not volatile, and have carcinogenic and mutation-causing characteristics, and potentially harmful to the environment characteristics, thus attracting wide attention [18]. Dimethyl phthalate (DMP), a known persistent organic pollutant and one of the PAEs, is a universal pollutant [18,19]. China has also classified dibutyl phthalate (DBP), dimethyl phthalate (DMP), and diester of phthalate (DEHP), as priority pollutants for environmental control in China. Liao et al. [20] reported that DMP, DBP, and DEHP were the main components of PAE pollutants in the soil of West Lake Scenic Spot that in Hangzhou city, Zhejiang Province, China, and the contribution rate of total content ranged from 95.65% to 99.22%. PAEs can undergo a series of reactions, such as hydrolysis, adsorption, photochemical reactions, and microbial degradation in water. Under natural conditions, the main process of PAE disappearance in the environment should be the biodegradation reaction [21–23]. The content of PAEs in Chinese river water environments is usually in the order of 10^{-1} - $10^2 \,\mu$ g/L [24], but the concentration of PAEs in wastewater discharged from some chemical plants and nearby rivers is 10-300 ug/L, while the concentration of PAEs in wastewater discharged from plasticizer production plants is up to 30 mg/L [25]. In this study we used constructed wetland technology to treatment DMP with a concentration of 8.12 mg/L. At present, there are few studies on the application of constructed wetlands in the treatment of PAEs [26–28], especially in the treatment of high-concentration PAES [26], and all of these have different removal effects. The main objectives of this study are: (1) study the effect of DMP on the microbial diversity and composition in VFCW; (2) understand the dominant bacterial in processing DMP and to determine enzyme activity for assessing the potential of DMP removal in VFCW.

2. Materials and Methods

2.1. Construction of VFCW

Three parallel VFCWs were constructed in the experiment, which were CW A (planted *Pennisetum sinese* Roxb), CW B (planted *Pennisetum purpureum* Schum.), CW C (unplanted). Each CW was a cement structure with a specification of 200 cm (length) \times 100 cm (width) \times 130 cm (height), as shown in Figure 1. Three CWs were irrigated effluents with three peristaltic pumps. Studies have shown that gravel was an ideal substrate for the treatment of dibutyl phthalate sewage [29], so the substrate in these CWs consisted mainly of grit. Three CWs were placed in barbed wire chambers with glass roofs.



Figure 1. Schematic diagram of profile and plan of vertical flow constructed wetlands (VFCWs).

2.2. Inflow Water Quality

The experimental sewage is synthetic sewage, and the composition simulation is close to domestic sewage. Parameter values are as followed: the variation range of dissolved oxygen (DO) and pH values are 4–6 mg/L and 6.4–6.7. The average values of total nitrogen (TN), total phosphorous (TP), chemical oxygen demand (COD), and DMP are 40 ± 5.5 mg/L, 4 ± 1.1 mg/L, 330 ± 40.9 mg/L, and 8.12 ± 0.69 mg/L. The water quality indexes including of COD, TN, and TP are measured according to the national standard method. DMPs are analyzed by HPLC (Agilent 1260). Dimethyl phthalate standard (>99.2%) is purchased from Merck. The UV detector analysis wavelength of the HPLC (Agilent 1260) is set at 254 nm. The reversed phase chromatography is carried out on the EclipsexDB-C18 column with methanol and ultra-pure water mixture (50:50) as a mobile phase at a flow rate of 1.0 mL/min. The column temperature is constant at 23 °C. The automatic sampling system took 20μ L samples at a time and injected them into the column for analysis. The relative standard deviation between the analysis results and the standard sample is less than 2%.

2.3. Operation and Management of VFCW

Three VFCWs were carried out at 20 cm/d hydraulic loading. The sewage contained DMP was added on the 15th of October; three systems were irrigated continuously with DMP sewage for 8 h every day. When the microorganism in the wetland is stable, the water quality monitoring work begins. On the 11th of November, we collected the water from the outlet position and analyzed every two weeks. This experiment ended on the 7th of February of the following year, with a total period of 114 days.

2.4. Sample Collection and Analysis

2.4.1. Substrate Collection and Analysis

The substrate sampling was made by sampling in a five-point method for analysis of microorganism and enzyme activity. Five points were collected for each wetland matrix, and each point was divided into two layers. The upper layer was 0–40 cm, and the lower layer was 40–80 cm. Finally, the same layer samples were mixed evenly to be the final substrate samples. Before the experiment, the upper layer substrate sample of the three wet systems were named by A-1, B-1, and C-1, respectively. The lower layer substrate sample of the three wet systems were named by A-2, B-2, and C-2, respectively. Similarly, at the end of experiment, the same method was adopted for substrate sampling. The upper layer of the three systems was numbered a-1, b-1, and c-1, and the lower layer was numbered a-2, b-2, and c-2. Part of the substrate samples collected were stored in a refrigerator of 4 °C for the detection of enzyme activity, and other parts of the samples were stored in a refrigerator

of -40 °C for the experimental analysis of microbe. After the substrate was collected, the samples were sent to the laboratory with an ice box for subsequent extracting DNA.

2.4.2. Methods for the Determination of Substrate Enzyme Activity

Urease was determined by the phenol sodium-sodium hypochlorite colorimetric method. Phosphatase was determined by the phenyl phosphate disodium sodium colorimetric method. Catalase was determined by the potassium permanganate titration method and invertase was determined by the sodium thiosulfate titration method.

2.4.3. Water Sample Collection and Analysis

Water samples were collected every 14 days.

2.5. Extraction and Detection of DNA from Soil Samples

Mobio Power Soil DNA Isolation Kits (MO BIO Laboratories, Inc, Carlsbad, CA, USA) were selected to extract soil sample DNA. DNA was stored in an ultra-low temperature refrigerator at -80 °C.

2.6. Microbial Diversity and Richness Index

The bacterial diversity index is a comprehensive index to study the number of species and the number of individuals of the community. According to the number of sample bands in the electrophoretic pattern and the intensity (gray scale) of each band, the indexes of bacteria in the Shannon–Wiener diversity index (H), evenness index (E), and richness index (S) in each sample were analyzed. The calculation formula of each indicator is as follows:

$$H' = -\sum_{i=1}^{s} (p_i \ln p_i) = -\sum_{i=1}^{s} (N_i / N) \ln(N_i / N)$$
(1)

In the formula, p_i represents the ratio of the strength of a single strip in the sample to the total strength of all the strips in the sample. N represents the abundance of all the bands in a single lane of DGGE map. N_i represents the abundance of band *i*. S represents the sum of the number of bands in a sample.

2.7. Statistical Analysis

Quantity one software was used to analyze the number and density of each sample. GCTA software was used for PCA analysis. The statistical analyses were performed in Excel 2007 and SPSS(IBM) 26.0 software. Mean values and standard errors of correlation analysis were calculated. The Duncan method was used to analyze and compare the multiple comparisons. Because of the large scale of wetlands, no treatment duplication was set up, and the statistical analysis duplication in this paper is the absolute duplication of sampling.

3. Results

3.1. DMP Removal in VFCW

The DMP removal efficiency of the three wetlands was 100% on the 1st of November and DMP could hardly be detected in the effluent. However, after two weeks, the removal rate of DMP dropped rapidly. After November 29th, the removal rate of DMP increased



and became stable (Figure 2). At the end of experiment, the removal rate of DMP in these three CWs was more than 90%. The three VFCWs all performed well in DMP removal.

Figure 2. Dimethyl phthalate (DMP) removal efficiency in three VFCW.

3.2. Dominant Microorganisms Based on DMP Contamination

DGGE analysis of the PCR amplification products of each sample can be separated by the electrophoresis bands with different numbers and positions, so as to identify the differences of microbial community structure in the samples with different treatments. According to the principle that DGGE can be separated from DNA of the same length with different sequences, each stripe corresponds roughly to a dominant flora or Operational Taxonomic Unit (OTU) in the community. The fluorescence intensity after the band staining reflects the abundance of this type of bacteria. The brighter the stripe signal, the more bacteria there are in the genus. Consequently, it reflects the species and quantity of bacteria in the wetland. The analysis results of DGGE gel electrophoresis (Figure 3) showed that there were 40 types of dominant bacterial flora in CWs, and analyses the DGGE gel bands recovery sequence (Table 1). Generally, the homology of 16S rDNA sequence is less than 98%, which can be considered as belonging to different species of bacteria. If the homology is less than 93–95%, it can be considered as belonging to different genera. DMP was added into CW to induce the growth and reproduction of the microbial community related to the decomposition of the substance. The results indicated that after adding DMP in these three CWs, the dominant bacteria in substrate were *Dechloromonas agitata*, *Labilithrix luteola*, Plasticicumulans lactativorans, Pleomorphomonas sp., Novosphingobium sp., Denitratisoma *oestradiolicum, Alicycliphilus denitrificans, Thauera* sp., and *Levilinea saccharolytica*. From the view of microorganism, VFCW has great potential to degrade DMP.

Table 1. BLAST results of selected DGGE bands from the DGGE	profiles.
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Band Number	Length of Fragment (bp)	Closest Related Strain	Accession Number	Similarity (%)
Band1 (B-1)	169	Uncultured bacterium	KF182847	100
Band2 (C-1)	193	uncultured Verrucomicrobiales bacterium	LN625193	97
Band3 (b-1)	194	Dechloromonas agitata	KF800710	100
Band4 (B-1)	169	Uncultured bacterium	AB661277	100
Band5 (C-1)	194	Thauera sp.	AB757831	100
Band6 (A-1)	194	Chlamydiales bacterium	JN606074	96
Band7 (c-1)	193	Labilithrix luteola	NR_126182	97
Band11 (c-1)	189	Prevotellaceae bacterium	AB298732	98
Band13 (b-1)	194	Plasticicumulans lactativorans	NR_118276	99
Band14 (b-2)	169	Pleomorphomonas sp.	KF983816	99
Band18 (A-1)	169	Clostridium beijerinčkii	KJ194928	100
Band23 (c-1)	169	Novosphingobium sp.	KP284177	99
Band24 (C-1)	189	Chitinophaga terrae	KP076216	98
Band28 (C-1)	169	Reyranella graminifolii	NR 126180	97
Band29 (c-2)	194	Uncultured bacterium	KJ461512	99
Band32 (b-1)	194	Denitratisoma oestradiolicum	KF810120	98
Band33 (B-2)	173	Uncultured bacterium	GU738861	97
Band35 (b-1)	194	Alicycliphilus denitrificans	KM210246	100
Band39 (a-2)	194	Thauera sp.	AB920830	99
Band40 (b-2)	170	Levilinea saccharolytica	NR_040972	99

Note: Bands are numbered according to Figure 3. Similarity represents the percentage identity shared with the sequences in the GenBank databases.



Figure 3. Denaturing gradient gel electrophoresis (DGGE) profile of 16S rDNA V3 fragments amplified from a microbial community of three wetlands. (Note: A-1, A-2, B-1, B-2, C-1 and C-2 were substrate samples before adding DMP in the three constructed wetlands (CWs). The samples of a-1, a-2, b-1, b-2, c-1, and c-2 were substrate samples after adding DMP in the three CWs. In the sample code, the uppercase English letter refers to the sample before adding DMP, and the lowercase letter refers to the sample after adding DMP (the letters A and a represent the wetland A of planted *Pennisetum sinese* Roxb; B and b represent the wetland B of planted *Pennisetum purpureum* Schum.; C and c represent the blank control wetland C; the Arabic numerals represent the sampling depth: 1 represents the upper substrate of 0–40 cm, and 2 represents the lower layer substrate of 40–80 cm. Serial numbers correspond to bands excised from the DGGE and sequenced results are listed in Table 1).

3.3. Microbial Community Grouping Based on DMP Contamination

DMP has a great influence on bacterial diversity in the substrate of VFCW (Figure 4). Each point in the figure represents a sample, and the same color points are from same wetland. The closer the distance between the two points, the smaller the difference in community composition between the two points; otherwise, the greater the difference. Principal coordinate analysis (PCA) was conducted to investigate the dissimilarities among the 12 samples. The first axis explained 27.6% variance of species and 25.3% was explained by the second axis. These 12 samples formed two groups as follows: one was the DMP group containing a-1, a-2, b-1, b-2, c-1, and c-2 while the other was no DMP group including A-1, A-2, B-1, B-2, C-1, and C-2. Seen from Figure 4, there was a high similarity among these samples of each group, while there was a great difference in the similarity among the groups. It indicated that the microbial flora structure changed, and the samples before and after the experiment showed different flora compositions, which were directly related to the presence or absence of DMP. Seen from Table 2, Shannon–Wiener and richness of 12 samples all showed the same rule: the upper layer was higher than the lower layer, and the addition of DMP has different effects on the three wetlands. Adding DMP reduced the microbial diversity of the planted vegetation wetland and increased the microbial diversity of the unplanted wetland. This also proved that planting vegetation contributes to the construction of wetland microbial community diversity, and the addition of DMP can narrow the microbial diversity gap between planted vegetation wetland and unplanted

wetland. In the future, we can further study the long-term effects of DMP on constructed wetland vegetation.



Figure 4. Principal coordinate analysis (PCA) score plot of different wetland samples based on weighted UniFrac metrics. Two categories were clustered, including the DMP group and the NO DMP group.

Samples	Shannon-Wiener	Richness
A-1	3.33	29
A-2	3.22	26
B-1	3.38	31
B-2	3.07	23
C-1	3.08	23
C-2	2.73	16
a-1	3.13	24
a-2	3.00	21
b-1	3.23	27
b-2	3.30	28
c-1	3.30	29
c-2	2.90	19

Table 2. Shannon-Wiener diversity and richness index of substrate in VFCW.

3.4. Substrate Enzyme Activity Based on DMP Contamination

The activity of these enzymes in the substrate had the same changed rules, all of which showed that the enzyme activities of the substrate at upper layer were significantly higher than those at the lower layer (p = 0.000 < 0.01). The enzyme activities showed a negative correlation with depth. In this study, the activities of urease, phosphatase, catalase, and invertase in VFCW after adding DMP were higher than that of those before adding DMP. At the end of experiment, the urease, phosphatase, catalase, and invertase activities at upper layer of the three wetlands reached up to 172.8 µg/(g·24 h), 98.3 µg/(g·24 h), 0.0223 nmol/g, 1.92 mL/(g·24 h), respectively. Before adding DMP, these enzyme activities in wetlands were 99.8 µg/(g·24 h), 79.8 µ g/(g·24 h), 0.022 nmol/g, 0.91 mL/(g·24 h) respectively. In general, the enzyme activity and microbial metabolism of the three wetlands were active in the experimental period.

4. Discussion

PAEs, as a kind of environmental endocrine disruptor, have potential threat to human health [30,31] and ecological environment. As a sewage treatment technology, CW has been

used to treat various types of sewage, but its removal efficiency of PAEs, especially the high concentration PAEs, has been rarely studied [28]. DMP, a known persistent organic pollutant and one of the PAEs, is a universal pollutant [18]. The removal rate of DMP analysis showed that the DMP removal efficiency of the three wetlands was 100% on the 1st of November; however, the removal rate of DMP has been decreased on the 15th of November. After November 29th, the removal rate of DMP increased and tended to be stable. That is because substrate adsorbed DMP significantly at the beginning of VFCW operation. However, when the substrate adsorption tends to saturation, the DMP removal rate decreased sharply, and the ability of the wetland to remove DMP will not be restored until the microorganisms in the wetland reach a stable state. Some studies have shown that under natural conditions, the main process of PAEs disappearance in the environment should be a biodegradation reaction [18,28]. As an ecological treatment technology, a constructed wetland should have a certain ability to remove PAEs. Previous study has shown that the influent concentrations of DBP in the composite wetland were 9.84 mg/L and the average removal efficiency of effluent was over 79% [26], while in reed bed sewage treatment process, the removal rate of diethyl phthalate (DEP) was as high as 94-99% [32], which was similar to our study. In our study, the removal rates of DMP in these three CWs were more than 90%. That is due to the fact that there were some functional bacteria and active enzymes in these three wetlands (seen from Tables 1 and 3).

			Before Adding DMP		After Adding DMP	
			Upper Layer	Lower Layer	Upper Layer	Lower Layer
urease activity	[µg/(g·24 h)]	А	$99.8\pm7.08^{\text{ b}}$	10.2 ± 1.34 $^{\rm a}$	172.8 ± 4.39 $^{\rm a}$	$25.8\pm1.27^{\text{ b}}$
		В	124.5 ± 4.27 a	14.2 ± 6.82 ^a	167.7 \pm 2.77 $^{\mathrm{a}}$	$30.1\pm3.03~^{\mathrm{ab}}$
		С	80.5 ± 2.18 c	8.1 ± 1.26 a	150.8 ± 2.50 $^{\rm b}$	32.9 ± 2.18 a
phosphatase activity	[µg/(g·24 h)]	А	79.8 ± 14.52 $^{\rm a}$	0.3 ± 0.16 ^a	98.3 ± 0.12 ^b	$4.8\pm0.14^{\text{ b}}$
		В	82.4 ± 3.78 $^{\mathrm{a}}$	0.2 ± 0.09 ^a	102.4 \pm 2.26 $^{\mathrm{a}}$	6.7 ± 0.41 ^a
		С	$42.2\pm18.28^{\text{ b}}$	0.1 ± 0.03 $^{\rm a}$	$86.3\pm2.15~^{\rm c}$	$2.8\pm0.11~^{c}$
catalase activity	(nmol/g)	А	$0.0211 \pm 0.001 \; ^{\rm a}$	0.009 ± 0.003 ^a	$0.0223 \pm 0.0012~^{a}$	$0.0077 \pm 0.0009~^{\rm a}$
		В	0.022 ± 0.003 $^{\rm a}$	$0.007 \pm 0.001~^{ m ab}$	0.0213 ± 0.000 ^{ab}	$0.0071 \pm 0.0002 \ ^{\rm a}$
		С	0.017 ± 0.003 $^{\rm a}$	$0.004 \pm 0.001 \ ^{\rm b}$	$0.0205\pm 0.0002~^{\rm b}$	$0.0067 \pm 0.0008 \ ^{\rm a}$
invertase activity	[mL/(g·24h)]	А	0.91 ± 0.085 $^{\rm a}$	$0.01 \pm 0.005 \ ^{\rm b}$	1.92 ± 0.05 $^{\rm a}$	$0.90\pm0.10~^{\mathrm{ab}}$
		В	0.84 ± 0.392 ^a	0.2 ± 0.087 $^{\mathrm{a}}$	1.85 ± 0.03 ^a	1.0 ± 0.17 ^a
		С	0.48 ± 0.087 $^{\rm a}$	$0.10\pm0.065~^{\mathrm{ab}}$	1.85 ± 0.04 $^{\rm a}$	0.73 ± 0.08 ^b

Table 3. Enzyme activity of substrate in the three VFCWs.

Note: Mean \pm standard deviation, where a, b, and c in the table represent the significant differences among different CWs, respectively.

Microorganisms are the most important drivers of soil biochemical processes and are very sensitive to changes in the soil environment, which can be used for evaluation changes in soil quality during pollution [32]. Microbial sequencing and cluster analysis were performed to analyze the structure of the microbial community in 12 samples of three wetlands. PCA analysis showed bacterial communities in the 12 samples were divided into two groups: the no DMP group and the DMP group. This result indicated that DMP could be the important factor altering microbial community structure. This is similar to a previous study in DMP contamination on abundance and diversity of microbes in black soil [18,33]. There are a large number of functional microbial flora in the soil, which catalyze the biochemical reactions in the soil and directly participate in the cycling process of nutrient elements [34,35]. Their quantity and activity are directly related to the ecosystem function of the soil. The results of the author's experiments showed that some functional bacteria appeared, such as, Dechloromonas agitata [36], Pleomorphomonas sp. [37], and Denitratisoma oestradiolicum [38]. These bacteria were involved in the nitrogen cycle and became the dominant bacteria in the DMP treatment wetland, indicating that the high concentration of DMP had a positive effect on the microorganisms involved in the nitrogen cycle in the CW. Meanwhile, Plasticicumulans lactativorans [39], Novosphingobium sp. [40],

Alicycliphilus denitrificans [41], and Thauera sp. [42] have the ability to degrade polycyclic aromatic hydrocarbon compounds and can be widely used in sewage treatment as the dominant flora. It also showed that DMP could be used as a carbon source to facilitate the growth and reproduction of related functional microorganisms. Some studies showed that the inhibitory effect of phthalates on microorganisms was enhanced with the increase of pollutant mass fractions, showing a good dose–effect relationship [43,44]. Meanwhile, as a biocatalyst, the enzyme in CW plays an important role in the decomposition of organic pollutants. It has been found that microorganisms and substrate enzymes play a key role in the treatment of organic pollutants in composite VFCWs [26]. These reported that soil enzymes can accelerate the chemical reaction of organic substances [45]. Dehydrogenase, catalase, phosphatase, urease, protease, and other enzymes in the soil can be responded to PAEs [43,46–48]. In the enzyme activity analysis that was performed after adding DMP, the urease, phosphatase, catalase, and invertase activity in wetlands did not decrease; on the contrary, these enzyme activities after adding DMP were higher than those before adding DMP in the three wetlands. It is similar with the study on the increase of enzyme activities of urease, phosphatase, and dehydrogenase that was observed in the treatments with DBP addition [49]. At the same time, the enzyme activities of the substrate at the upper layer are significantly higher than those at the lower layer (p = 0.000 < 0.01), and the microbial diversity in the upper layer is also higher than at the lower layer (Table 2). This result is similar to previous studies in which with the increasing substrate depth, the enzyme activity decreased, and the number of bacteria reflected by the spatial law was consistent with the enzyme activity [50].

5. Conclusions

(1) The results of this study demonstrated that DMP could change the bacterial community structure, and increase the relative percentages of some genera associated with nitrogen metabolism and the function of degrading aromatic hydrocarbons, such as *Dechloromonas agitata*, *Pleomorphomonas* sp., *Denitratisoma oestradiolicum*, *Plasticicumulans lactativorans*, *Novosphingobium* sp., *Thauera* sp., and *Alicycliphilus denitrificans*. It showed that DMP (8.12 mg/L) could be used as a carbon source to facilitate the growth and reproduction of some functional microorganisms. In the future, we will further study the influence of different concentration levels of DMP on wetland microorganisms.

(2) DMP maintained the higher activities of urease, phosphatase, catalase, and invertase in these three CWs, and stronger activities of these four enzymes occurred at the upper layer of the three CWs. In addition, the activities of urease, phosphatase, catalase, and invertase in the upper layer are significantly higher than in the lower layer, indicating that VFCW has great potential to degrade high concentration DMP. Furthermore, at the end of the experiment, the removal rates of DMP in these three CWs are more than 90%.

(3) The results also showed that the upper layer is the main site for the degradation of DMP. It provides theoretical and data support for the treatment of DMP in CWs.

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