



Article Mapping Soil Organic Carbon Stock and Uncertainties in an Alpine Valley (Northern Italy) Using Machine Learning Models

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Abstract: In this study, we conducted a comprehensive analysis of the spatial distribution of soil organic carbon stock (SOC stock) and the associated uncertainties in two soil layers (0-10 cm and 0-30 cm; SOC stock 10 and SOC stock 30, respectively), in Valchiavenna, an alpine valley located in northern Italy (450 km²). We employed the digital soil mapping (DSM) approach within different machine learning models, including multivariate adaptive regression splines (MARS), random forest (RF), support vector regression (SVR), and elastic net (ENET). Our dataset comprised soil data from 110 profiles, with SOC stock calculations for all sampling points based on bulk density (BD), whether measured or estimated, considering the presence of rock fragments. As environmental covariates for our research, we utilized environmental variables, in particular, geomorphometric parameters derived from a digital elevation model (with a 20 m pixel resolution), land cover data, and climatic maps. To evaluate the effectiveness of our models, we evaluated their capacity to predict SOC stock 10 and SOC stock 30 using the coefficient of determination (\mathbb{R}^2). The results for the SOC stock 10 were as follows: MARS 0.39, ENET 0.41, RF 0.69, and SVR 0.50. For the SOC stock 30, the corresponding R^2 values were: MARS 0.45, ENET 0.48, RF 0.65, and SVR 0.62. Additionally, we calculated the rootmean-squared error (RMSE), mean absolute error (MAE), the bias, and Lin's concordance correlation coefficient (LCCC) for further assessment. To map the spatial distribution of SOC stock and address uncertainties in both soil layers, we chose the RF model, due to its better performance, as indicated by the highest R² and the lowest RMSE and MAE. The resulting SOC stock maps using the RF model demonstrated an accuracy of RMSE = 1.35 kg m^{-2} for the SOC stock 10 and RMSE = 3.36 kg m^{-2} for the SOC stock 30. To further evaluate and illustrate the precision of our soil maps, we conducted an uncertainty assessment and mapping by analyzing the standard deviation (SD) from 50 iterations of the best-performing RF model. This analysis effectively highlighted the high accuracy achieved in our soil maps. The maps of uncertainty demonstrated that the RF model better predicts the SOC stock 10 compared to the SOC stock 30. Predicting the correct ranges of SOC stocks was identified as the main limitation of the methodology.

Keywords: SOC stock; DSM; machine learning models; uncertainty mapping

1. Introduction

Soil is an essential resource that offers numerous benefits for sustainable development, especially in the domains of food security and environmental regulation. One of its critical services is the storage of soil organic carbon (SOC), which is pivotal for both climate change mitigation and adaptation. Moreover, SOC plays a vital role in water management, enhancing soil capacity to address both floods and droughts [1,2]. Poor soil management can lead to significant disruptions in soil parameters and characteristics, resulting in changes in SOC stocks. These changes, in turn, can cause the release of substantial amounts of carbon into the atmosphere. Sequestering carbon in the soil is a valuable method for controlling greenhouse gas levels in the atmosphere [3]. Studies indicate that this approach has the potential to capture approximately 0.8 to 1.5 billion metrics tons of carbon annually.



Citation: Agaba, S.; Ferré, C.; Musetti, M.; Comolli, R. Mapping Soil Organic Carbon Stock and Uncertainties in an Alpine Valley (Northern Italy) Using Machine Learning Models. *Land* **2024**, *13*, 78. https://doi.org/10.3390/ land13010078

Academic Editors: Maria Fantappiè, Giuseppe Lo Papa, Calogero Schillaci and Giuliano Langella

Received: 30 November 2023 Revised: 28 December 2023 Accepted: 5 January 2024 Published: 10 January 2024



Copyright: © 2024 by the authors. Licensee MDPI, Basel, Switzerland. This article is an open access article distributed under the terms and conditions of the Creative Commons Attribution (CC BY) license (https:// creativecommons.org/licenses/by/ 4.0/). As a result, there is a high demand for accurate information and maps explaining the actual SOC stock and the soil capacity for SOC sequestration [4].

Mountain ecosystems are characterized by a substantial amount of biological and cultural diversity, and they play a crucial role in providing essential services such as water and food security, and energy generation, as well as aesthetic and spiritual qualities [5]. According to the IPCC's WGII Sixth Assessment Report on Mountains [6], it is assumed that these ecosystems are extremely vulnerable to global changes. Mountainous soil is naturally vulnerable, and it is increasingly sensitive to changes in the environment [7]. Understanding the spatial distribution of SOC stocks in alpine mountains is essential for developing sustainable management strategies and environmental policies that can effectively address future global changes. However, this task remains difficult due to the complex morphology of these environments, which makes the collection of soil data challenging. Furthermore, comprehensive soil maps and information in the alpine mountains are scarce [8]. In Italy, where a significant portion of the land area is covered by mountains, the monitoring and assessment of the functionality of mountain soils becomes crucial. Detailed and accurate maps of SOC ensure that local and global decision makers have access to precise information. From a pedological perspective, soils in the Italian Alps show diversity due to variations in factors related to pedogenesis [9]. These factors are associated with the differing landscape, including diverse climatic conditions, geological substrates, geomorphological processes, and the heterogeneity in land use and land cover (LU/LC) [10]

The development of geographic information systems (GISs), remote sensing, and mathematical algorithms have improved the techniques of digital soil mapping (DSM), which is suitable for mapping soil parameters in mountainous areas. In recent years, there has been a surge in studies that focus on mapping soil properties by applying various strategies such as geostatistics and machine learning [11]. These methodologies have sought to overcome the limitations of traditional methods, which are time-consuming and labor-intensive and cannot capture the real variability of soil properties in complex environments. The machine learning models can be used to gain an understanding of the complex interactions between soil properties and environmental factors and generate accurate predictions and maps [12]. In scientific research focused on SOC in alpine mountains, the primary approach involves examining the connections between SOC and environmental factors. These factors typically include topography, vegetation cover, and climate parameters, which serve as the main variables employed in DSM techniques. Yang et al. in 2016 employed boosted regression trees (BRTs) and random forest (RF) to model and map the SOC content of the Tibetan plateau. The two models showed good results, explaining about 70% of the SOC spatial distribution [13]; vegetation cover and the topographic variables were the most important covariates for SOC prediction. The mapping of SOC stock of several land cover types was carried out in the Bernese Alps, Switzerland, using different approaches [7]. The results of this research showed that, except for Regression Kriging, all interpolation approaches exhibited little variability in the RMSE of the expected SOC stock [7]. The spatial distribution of SOC stock in the Andossi plateau, Valchiavenna, was mapped at high resolution using Regression Kriging with geomorphometric parameters. A detailed vegetation map was produced to improve the model performance [14]. The geomorphometry influences soil formation and the storage of SOC in mountainous environments because it controls many factors of pedogenesis; for example, in the upper part of the slope, water and soil sediment (including organic matter) are lost without being compensated. On the other hand, at the foot of the slopes, sediment inputs lead to soil accretion. Southern exposures are warmer and drier, and vegetation tends to be thermophilic or xerophilic, while northern exposures are colder [15].

The diversity of geomorphometric conditions influences the spatial distribution of soil properties; therefore, geomorphometry is a mandatory variable in DSM methodology. Most of the research cited [8,14] pointed out the need to enhance mapping methods to gather precise and comprehensive data on mountainous areas [1,16,17].

Uncertainty mapping is a critical step in the DSM approach, although it is not yet used in all DSM papers. Soil maps are a simplified representation of a more complex reality. As a result, no model is error-free, and no map is 100% accurate [18,19]. The causes of uncertainty in DSM are diverse; we may highlight four major sources of uncertainty: (a) errors related to soil sampling and laboratory measurements; (b) uncertainty of soil geospatial position measurement; (c) uncertainties in covariate calculation; and (d) errors linked to modeling approaches. These lead to several errors in DSM outcomes. Statistical analyses of uncertainties and their mapping are strong tools for assessing map errors; they are critical for soil map users since they provide additional information about the error average that should be considered during the decision-making process [19–21].

The main objectives of our research were to compare four machine learning models as DSM techniques, using geomorphometric and climatic variables, as well as land cover as covariates. To: (i) map the SOC stocks of two layers (0–10 cm: SOC stock 10 and 0–30 cm: SOC stock 30), (ii) estimate the associated uncertainties in an alpine valley, as well as understand the spatial distribution of SOC stock and the uncertainties within each land cover.

2. Materials and Methods

2.1. Study Area

Valchiavenna is a valley in the Central Alps, located in the province of Sondrio, Lombardy. It has a north-south orientation and covers an area of 450 km²; it is characterized by a varied landscape; the elevation changes from around 200 to 3279 m a.s.l. The morphology of the valley is linked to the action of water and glaciers, which act at different times and in different ways. Glacial erosion is responsible for the transverse U-shaped profiles of the valley and its hanging sides. In addition, fluvial erosion forms have influenced and frequently re-shaped previous glacial morphologies. Valchiavenna has a considerable range of lithologies with crystalline-acidic character, mainly of metamorphic origin, and subordinately igneous rocks (late-Alpine Pluton intrusive body of Val Màsino and Val Bregaglia), as well as mesozoic cover and the group of mafic and ultramafic rocks (ophiolitic complex). In restricted areas (Pian dei Cavalli and the Andossi plateau), there are outcrops of sedimentary rocks of carbonate type. According to the classification of climates by Köppen (1936), the climate of Valchiavenna is Cfb (humid temperate with maximum summer rainfall), with an average annual precipitation in the range of 1000–1400 mm. The average annual temperature at the foot of the valley is 12.8 °C, as measured by the Chiavenna meteorological station at 333 m a.s.l.; in the upper part of the valley, at Montespluga station (1908 m a.s.l.), the mean annual temperature drops to 2.7 °C. Valchiavenna has a high diversity in terms of vegetation and land use, from meadows and arable land in the lower parts to oak forests, coniferous forests, and finally, alpine grasslands at high altitudes. Various soil types are present in the study area, classified as: Leptosols, Regosols, Cambisols, Umbrisols, Podzols, and Histosols (according to the World Reference Base (WRB) for Soil Resources) [22]. The soils in this study area are mostly coarse-textured (sandy loam; sometimes loam or loamy sand), often with a high content of rock fragments. In general, soil thickness ranges from 20 to 90 cm.

2.2. DSM Approaches in SOC Stock Mapping

To achieve our objectives, we performed the following steps:

- Soil survey and laboratory analyses.
- Calculation of SOC stock at each sampling point.
- Calculation of environmental covariates.
- Preparation of the covariate maps (with a spatial resolution of 20×20 m).
- Extraction of the environmental covariates at each soil sampling point.
- Environmental covariates selection, using a statistical correlation matrix.
- Comparison of different machine learning models to estimate the SOC stocks.
- SOC stock mapping.

• Obtaining estimation uncertainty maps.

2.2.1. Soil Survey and Data Collection Strategy

The sampling was scattered across 18 topographic transects, chosen according to the physical nature of the study area. The main sampling transects were on the north–south axis, corresponding to the main orientation of Valchiavenna, and in transverse directions (generally east–west) along its secondary valleys (Figure 1). The position of each soil profile was chosen based on elevation (approximately every 300 m). Since changes in altimetry across the topographic transects are associated with changes in the landscape (geomorphometry and vegetation), this sampling method provides an accurate representation of the valley landscape and its pedological variability. All the sampled soil profiles were georeferenced using a high-accuracy GPS. After the description of the profile, soil samples were collected from each horizon. At the end of the pedological survey, 110 soil profiles were described, for a total of 496 soil samples. The density of sampling points by km² is represented in Tables S3 and S4 (see Supplementary Materials).



Figure 1. Geographical position of the study area and soil profile location.

2.2.2. Laboratory Analysis Methods

The soil samples collected in the field were air-dried and sieved through a 2 mm sieve. The standard laboratory analyses were performed on the fine earth.

Soil pH was measured potentiometrically in a soil-to-water ratio of 1:2.5. The organic carbon was determined via oxidation with $K_2Cr_2O_7$ in an acid environment: for samples very rich in organic matter and those taken from Histosols, we measured OM via incineration in a muffle furnace at 550 °C (LOI). The sieving and sedimentation method (pipette method) was used to obtain textural fractions: coarse sand (2.0–0.1 mm), fine sand (0.1–0.05 mm), silt (0.05–0.002 mm), and clay (<0.002 mm).

As the main objective of this work was to map the SOC stock by soil layers, the calculation per unit area was carried out as follows: the SOC content and that of the rock fragments (described in the field) of each soil layer was calculated; then the bulk density of the fine earth of each layer (BD1 and BD2, 0–10 cm and 10–30 cm depths, respectively; Equations (1) and (2)) was estimated using pedotransfer functions (unpublished) obtained in a detailed study of soils on the Andossi plateau (upper Valchiavenna). After obtaining these data, we calculated the SOC stock of each soil layer using Equation (3).

BD1 =
$$-0.293\ln(SOC) + 1.253 (n = 110; R^2 = 0.08)$$
 (1)

BD2 =
$$-0.242\ln(SOC) + 1.2002 \left(n = 66; R^2 = 0.66\right)$$
 (2)

SOC stock
$$\left(\text{kgm}^{-2} \right) = \left(1 - \frac{\text{vrf}}{100} \right) \times \frac{\text{ht} \times \text{BD} \times \text{SOC}}{10}$$
 (3)

where: SOC = organic carbon content (%); BD = bulk density (g cm⁻³); ht = horizon thickness (cm); and vrf = volumetric rock fragments content (%).

2.2.3. Environmental Covariates

The environmental variables used as covariates are illustrated in Table 1. The covariates were calculated with different methodologies and transferred to raster layers with a 20 m spatial resolution in a GIS environment, using the open-source software QGIS 3.16.1. We used three different types of environmental covariates, geomorphometric, climatic, and land cover, as follows:

- Geomorphometric covariates: To calculate these covariates, we used the digital terrain model (DTM), delivered from the regional geo-portal of Lombardy (www.geoportale. regione.lombardia.it (accessed on 15 October 2022), and extracted 16 morphometric parameters. The calculation was carried out in QGIS 3.16.1 using the integrated SAGA tool.
- Climatic covariates: We used mean annual air temperature (T) and precipitation (P) delivered from WorldClim (www.worldclim.org (accessed on 5 January 2023) with a spatial resolution of 1 km². We applied a statistical downscaling technique using a 30-year time series of climatic data registered at seven meteorological stations in Valchiavenna, to obtain climatic covariate maps with the same spatial resolution as the other environmental variables (20 m). Working in an alpine valley, the downscaling technique was based on statistical correlations between climatic variables and elevation, and also with latitude and longitude [23]. The results of the correlations were used to obtain T and P maps of the area, correcting the estimated values for slope and exposure, which have a direct impact on microclimatic conditions in mountainous environments [24]. The equations used for climate downscaling are explained in the Supplementary Materials (Equations (S1)–(S5)).
- Land cover covariates: We used the most recent land cover maps of Lombardy, related to agricultural and forestry use (DUSAF 7.0) [25], and identified six land cover classes in the study area: broadleaf forests, coniferous forests, grasslands (low elevation), prairies (high elevation), peatlands, and rocky soils.

Coveriatos Namos	Abbreviations -	Main Statistics					
Covariates Names		Min	Mean	Median	Max	SD	
Elevation (m)	Elv	197	1558.57	1664.21	3262	723.48	
Slope (°)	Slp	0	31.75	32.93	80.08	15.40	
Northness Index	N_ind	-0.99	-0.14	-0.31	1	0.74	
Eastness Index	E_ind	-0.99	-0.05	-0.07	0.99	0.67	
Profile Curvature	Pr_cur	-0.277	-0.000118	-0.00003	0.208	0.007	
Plan Curvature	Pl_cur	-14.224	0.000095	0.00062	8.503	0.045	
Min Curvature	Min_cur	-0.666	-0.010872	-0.00515	0.242	0.023	
Log Curvature	Log_cur	-0.919	-0.000248	-0.00004	0.680102	0.039003	
General Curvature	Gen_cur	-1.426	0.000063	0	1.167034	0.07111	
Max Curvature	Max_cur	-0.309	0.010903	0.00539	0.483	0.022	
Transversal Curvature	Tra_cur	-0.773112	0.000311	0.00007	0.829	0.04	
Total Curvature	Tot_cur	0	0.000986	0.00015	0.319	0.003	

Table 1. Main statistics of climate and geomorphometric covariates extracted from the 20 m DTM.

Covariatas Namas	Abbreviations	Main Statistics					
Covariates maines		Min	Mean	Median	Max	SD	
Tang Curvature	Tan_cur	-0.269201	0.000099	0.000071	0.298031	0.014142	
Terrain Ruggedness Index	TRI	0.0013	11.09	10.27	94.32	7.119	
Terrain Position Index	TPI	-81.178	0.0055	-0.0012	65.2903	4.351	
Flow Accumulation	Fl_Acc	0	106.35	3	61576	1109.12	
Vector Ruggedness Measure	VRM	0	0.09	0.06	0.75	0.06	
Topographic Wetness Index	TWI	2.808	7.944	7.324	19.311	2.715	
Mean annual Temperature (°C)	Т	1.62	4.97	3.12	14.61	3.74	
Mean annual Precipitations (mm)	Р	514.8	1278.56	1268.6	1531.1	132.39	

Table 1. Cont.

2.2.4. Covariate Selections and Modeling Approaches

We used the statistical variable selection strategy, which is a mandatory step in DSM, to improve the models' performance and guard against noise and overfitting problems. Firstly, we created a correlation matrix between the different continuous variables, and when pairs showed a correlation coefficient >0.8 we removed one member of the pair. We chose this strategy as it is not a time-consuming methodology with a good performance. All the categorical variables (land cover) were used in the modeling by using binary (0/1) indicator variables for each category.

To understand the differences in the distribution of SOC stock according to the land cover, we used the one-way ANOVA with the post hoc Tukey HSD test (Tables S1 and S2). The statistical analysis and modeling were performed using R software version 4.3.0 (R Development Core Team, 2021). For the DSM approach, we built different machine learning models: MARS, ENET, RF, and SVR using the "Caret" and "Train" packages of the R software version 4.3.0 [26]. We also applied hyperparameter tuning to automatically select the best model structures according to the lowest prediction errors. We applied data standardization (Z score normalization) to models (SVR and ENET) that require this preprocessing step. For the hyperparameter optimization, we applied the grid search for each algorithm:

- Multivariate adaptive regression splines (MARS). In 1991, Friedman unveiled a new methodology that amalgamated linear regression with spline mathematical modeling through binary recursive partitioning [27]. This method constructs a model step by step, assessing variable importance and regularization to unimportant covariates. MARS is flexible, identifying complex nonlinear interactions between input variables, and it requires minimal pre-processing. Until now, the MARS model has not been widely applied in soil property prediction [28,29].
- Elastic net model (ENET). The model was introduced by Zou and Hastie in 2005 [30]. Similar to Lasso and Ridge Regression, it employs a regulation and variable selection technique, choosing the most advantageous combination of the two models. For studies with few observations and a high number of predictors, it is advised to use this model [30–32].
- Random forest (RF). Proposed by Breiman in 2001 [33], RF is the most used machine learning algorithm in DSM, as it has proven effective in mapping soil properties over an extensive variety of data sources and scales of soil heterogeneity. The model uses decision trees for training, combining them to produce single predictions for each observation in the datasets using an out-of-bag (OOB) strategy [34].
- Support vector machine (SVM). An effective machine learning method for mapping soil properties, largely used by soil mappers in recent years [35,36]; it is a kernel-based model, highly used to analyze nonlinear relationships over a high-dimensional induced feature space. SVM uses decision surfaces specified by a kernel function [37].

In the DSM approach, SVM is frequently used for classification, but it is also used for regression predictions.

2.2.5. Prediction Validation and Uncertainties Mapping

A 10-fold cross-validation was employed to assess the model. In DSM, cross-validation is frequently employed since it splits the data into several training and test datasets. Moreover, it is advisable to utilize the cross-validation technique when conducting studies in regions where data collection is limited, such as mountainous areas [38,39]. We employed the following metrics to validate the models: the mean absolute prediction error (MAE), the root-mean-squared error (RMSE), the coefficient of determination (R²), Lin's concordance correlation coefficient (LCCC), and bias. To map the uncertainties, we used the standard deviation (SD) of 50 runs, as proposed by the Global Map Project [19,20,34]; in addition, the zonal statistics was applied to understand the uncertainty distribution under the different land cover types. To better grasp how SOC stock is spread out across the valley, we examined its distribution based on certain geomorphometric parameters like slope, aspect, and elevation. We created boxplots to show the SOC stock within various classes of these parameters (Figures S6–S8 in Supplementary Materials).

3. Results

3.1. SOC Stock Statistical Analysis

The SOC stock values at 10 and 30 cm soil layers are summarized in Table 2. The results illustrate that the soils in our study area store a significant amount of SOC, especially in the top 30 cm, where the average is 8.72 kg m⁻². The mean SOC stock for 30 cm is approximately twice as high as that for 10 cm, which averaged 4.29 kg m⁻². The SD results reveal a high variability in the SOC stock data, indicating a high spatial heterogeneity in the distribution of SOC stock in our study area. This is a result of the high pedodiversity characterizing the Valchiavenna valley.

Coll Dronortion	Statistical Metrics						
Son riopentes	Min	1st Qu	Median	Mean	3rd Qu	Max	SD
SOC stock 10 (kg m ⁻²)	0.02	2.88	4.00	4.29	5.55	9.31	2.10
SOC stock 30 (kg m ⁻²)	0.03	5.13	7.27	8.72	10.93	29.90	5.51

Table 2. Analytical data of Valchiavenna soils.

The correlation matrix of the SOC stock with the environmental covariates is shown in Figure 2. Many variables are highly correlated, as, for example, temperature and elevation. Notably, parameters such as Pl_cur, Max_cur, and TPI exhibit significant correlations. Additionally, climatic factors such as T and P are shown to exert control over the SOC stock. It is important to note that while the Pearson correlation coefficient indicates this relationship, its capacity to explain the complex statistical dynamics of the relationship between SOC stock and the environmental parameters remains limited.

The boxplot of the distribution of SOC stock by land cover types (Figure 3) shows that by far the highest SOC storage is found in peatlands, while the lowest is found in high-altitude, thin, and skeletal soils. In the other cases, SOC storage is comparable, but that of soils in coniferous forests is on average lower than that of broadleaf forests and natural or cultivated grasslands. The results of the ANOVA analysis confirm this statistically: Tukey's HSD test shows that the greatest differences are found between rocky soils and peatlands (Tables S1 and S2).



Figure 2. Correlation matrix of SOC stock and the environmental covariates.



Figure 3. Boxplot of SOC stock distribution by different land cover types (BF: broadleaf forests; CF: coniferous forests; GR: grasslands; PR: prairies; PT: peatlands; RS: rocky soils): (a) SOC stock 0–10 cm; (b) SOC stock 0–30 cm. The boxplots represent the following metrics: the median, first and third quantile (Q1, Q3), maximum, minimum values, and outliers.

3.2. Model Validation and SOC Stock Prediction

The model validation results, obtained from an average of 50 training trials of the models, are shown in Table 3 and Figures S2 and S3 (see Supplementary Materials). For

both soil layers, the RF model demonstrated the best validation results, with the highest R^2 and LCC and the lowest RMSE, MAE, and bias close to zero. However, the errors of SOC stock prediction are higher for the SOC stock 30 (MAE = 2.48 kg m⁻²), compared to the SOC stock 10 (MAE = 1.10 kg m⁻²).

Model Performance		Machine Learning Models					
		MARS	ENET	RF	SVR		
SOC stock 10 (kg m ⁻²)	RMSE	1.63	1.61	1.35	1.50		
	R ²	0.39	0.41	0.69	0.50		
	MAE	1.25	1.23	1.10	0.98		
	LCCC	0.55	0.56	0.66	0.59		
	Bias	0.75	-1.25	0.01	-0.025		
SOC stock 30 (kg m ⁻²)	RMSE	3.47	3.97	3.36	3.46		
	R ²	0.45	0.48	0.65	0.62		
	MAE	2.67	3.01	2.48	2.25		
	LCCC	0.62	0.64	0.73	0.70		
	Bias	0.52	-0.67	0.03	-0.56		

Table 3. Validation performance of the different investigated machine learning models.

The SVR model also showed good results, better than for ENET and MARS, which were almost equal in performance. However, the results of bias illustrated that ENET notably underestimated the SOC stock, while the MARS model tended to overestimate it.

The results showed that the RF model performed well in predicting SOC stock in both soil layers, with particularly good results in the 0–10 cm compared to the 0–30 cm layer. When we compared MAE with the average SOC stock values (0–10 cm: 4.29 kg m⁻², 0–30 cm: 8.72 kg m⁻²), the RF model displayed a 26% error rate for SOC stock at 10 cm and a 28% error rate for SOC stock at 30 cm. Interestingly, the SVR consistently showed better results of MAE compared to the RF model. The results of bias illustrated that while ENET model highly underestimate the SOC stock, the MARS model show an important overestimation.

The order of importance of the predictors (Figure 4 and Figures S4 and S5 in Supplementary Materials) changes from one model to another, depending on the type of model and its structure. The MARS and ENET models used fewer variables than RF and SVR. In the RF model, land cover was the most important predictor, followed by climate parameters and several geomorphometric variables (mainly curvatures).



Figure 4. Predictors' importance of SOC stock mapping using RF model (PT: peatlands; RS: rocky soils; GR: grasslands; PR: prairies; BF: broadleaf forests; CF: coniferous forests): (**a**) SOC stock 0–10 cm; (**b**) SOC stock 0–30 cm.

3.3. Maps of SOC Stock and Uncertainty Estimation

We employed the RF model to represent the spatial distribution of SOC stock and the associated uncertainties (Figure 5) since it produced the best prediction results (Table 3).



Figure 5. Maps of SOC stock and associated uncertainties in Valchiavenna: (**a**) SOC stock 10 distribution map; (**b**) SOC stock 30 distribution map; (**c**) uncertainty map for SOC stock 10 (SD 10); (**d**) uncertainty map for SOC 30 (SD 30).

The prediction maps of SOC stocks show a similarity in the spatial pattern of the two soil layers considered. The central region of the valley has a higher storage of organic carbon: these are areas covered by broadleaf forests, coniferous forests, and grasslands, located at medium altitudes. The lowest values correspond to high-altitude and sloping areas, where the vegetation is sparse, and the soil is thin and rich in rock fragments (Figures S6 and S7 in Supplementary Materials). The valley floor areas show a different behavior: for the 0–10 cm layer they have stock values comparable to those of the forest areas, while for the 0–30 cm layer they have significantly lower stockage. This difference arises from the management of soils on the valley floor, which are alternated between grassland and arable land. Mechanically ploughing the soil results in a substantial loss of organic matter due to oxidation. Additionally, soils managed as grassland are also subjected to ploughing after a limited number of years. The value of the SOC stock, estimated cartographically, is obviously greater for the 0–30 cm layer (2.9 to 19.5 kg m⁻², with an average of 7.73 kg m⁻²), than for the 0–10 cm layer (0.8 to 6.8 kg m⁻², with an average of 3.72 kg m⁻²). Comparing these estimations to the observed data in Table 2, the average SOC stock across the entire area is lower than the observed average SOC stock values.

The maps displaying the uncertainty (obtained as the variance from 50 repetitions of the estimates) of SOC stock 10 has a range between 0.01 and 0.15 kg m⁻². For SOC stock 30 the error varies between 0.04 and 0.54 kg m⁻². These results indicate that there are generally low levels of uncertainty, underscoring the model's stability.

A statistical analysis of the uncertainty distribution across different land cover types, as shown in Figure 6, reveals that errors in SOC stock predictions tend to be higher at elevated altitudes, especially in areas with significant slopes and rocky soils.



Figure 6. Average of SOC stock uncertainty distribution under different land cover types (RS: rocky soils; GR: grasslands; CF: coniferous forests; PR: prairies; BF: broadleaf forests; PT: peatlands).

4. Discussion

4.1. Models' Performance

The performance of machine learning models can vary because each model operates differently, due to its unique structure. The choice of variables, which differs from model to model, has a significant impact on how well the model performs. For instance, in the MARS model, which presented the least accurate predictions, only a few variables were chosen, resulting in a loss of information about the relationship between SOC stock and environmental factors. In contrast, RF and SVR used a more extensive set of variables, leading to much better model performances. Our research obtained results consistent with previous scientific work on predicting and mapping soil properties such as SOC stock,

demonstrating the robust performance of the RF model. In complex tropical landscapes, RF rivalled the predicting power of the boosted regression tree (BRT) algorithm, skillfully handling data variability and mitigating irrelevant factors [40]. Similarly, in a study using Sentinel-1 and Sentinel-2 for soil mapping, RF competed effectively among machine learning methods for SOC prediction, highlighting its promise when coupled with multi-source sensor data [41]. In a study focused on employing machine learning for SOC prediction in agriculture, XGBoost demonstrated exceptional accuracy. In the same study, the RF model also performed admirably. Furthermore, the integration of Sentinel-1 and Sentinel-2 data significantly enhanced the precision of these predictions [42]. Another research project focused on predicting SOC content using RF, k-nearest neighbors (kNNs), SVM, artificial neural network (ANN), and ensembles. RF stood out, with excellent predictive performance [43]. Similarly, the work of Zhang et al. (2022) aims to map the SOC distribution in China using machine learning; when comparing models, RF emerged as superior, with higher R² and lower RMSE values across soil depths (0–10, 10–20, 20–30, and 30–40 cm) [32].

4.2. SOC Stock Spatial Distribution: The Main Drivers and Uncertainties

The results of the RF model show that the main environmental drivers of SOC stocks in Valchiavenna are land cover types, climate, and geomorphometric variables (slope, curvatures, and TWI). These results are in agreement with previous studies [16,17], which have shown that SOC stocks in mountain environments are strongly influenced by vegetation cover and climatic conditions.

Previous research has shown that the type of land cover and habitat significantly influence the storage of SOC stock in alpine mountains. Consequently, the type of vegetation is a crucial parameter because it directly impacts the storage of organic carbon in the soil [44]. Our results illustrate that peatlands, grasslands, and coniferous forests can store considerably more carbon in the soil compared to broadleaf forests and prairies. Our results also show that the SOC stock is significantly influenced by climatic conditions. Air temperature has a strong influence as it controls the rate of mineralization of organic matter in the SOC balance and therefore affects the output rate. In alpine ecosystems, there is a negative relationship between temperature and SOC storage, at least beyond the belt of natural grasslands, where thin and rocky soils have only sparse and discontinuous vegetation, in addition SOC stocks increase with elevation in these areas [45]. As a result of ongoing climate change, increases in temperature are expected to reduce SOC storage. Mitigation measures favoring carbon sequestration strategies (protection or restoration of peatlands, afforestation, sustainable grassland cultivation, etc.) should focus on the most fragile mountain ecosystems [44,46].

Precipitation also controls the dynamics of SOC storage in mountain soils: it is essential for net primary production (NPP) and has an impact on soil moisture, pH, and respiration [47]. However, research assessing the impact of changes in precipitation on the soil SOC budget is still limited [47,48]. Geomorphometric and topographic factors have an important influence on SOC stock spatial distribution, although their impact is generally less significant than climatic factors. The importance of geomorphometrical predictors on the spatial distribution of SOC stock differs between the top 10 cm and the top 30 cm of soil. For example, Figure 3 illustrates that the wetness index has a more notable effect on SOC stock distribution in the upper 10 cm compared to the upper 30 cm, as it is related to the soil water content, which influences indirectly the SOC stock. Slope and aspect control the solar radiation and soil moisture: steeper slopes often experience higher rates of erosion, which can result in reduced soil development and SOC storage; aspect influences the exposure to sunlight, affecting vegetation growth and decomposition rates, which, in turn, impact SOC accumulation. The landforms, such as the curvatures, control the zones of SOC erosion and deposition. Previous research has already demonstrated the relationship between SOC stock variability and geomorphometry [17,36,49]. The SOC stock maps and analyses of its distribution by topographical parameters (Figures S6–S8) confirm a strong link between topographic attributes and SOC stock levels. Specifically, elevations

between 1500 and 2500 m show higher SOC stocks, declining beyond 2500 m (Figure S6). Additionally, areas with steep slopes tend to have a lower SOC stock (Figure S7). Our study highlights those regions situated to the east and north exhibit high carbon stock (Figure S8). These areas, characterized by the highest precipitation, the coldest temperatures of the valley, and vegetation such as peatlands, grasslands, and coniferous forests, consistently demonstrate elevated SOC storage (Figure S1).

By examining the uncertainty maps, it appears that there is a correlation between topographical parameter attributes and prediction errors. Higher altitudes with significant slopes and rocky soils exhibit greater prediction errors, particularly in regions with complex topography and shallow soils prone to erosion. In contrast, valleys and low-lying areas display lower uncertainties due to their more uniform soils. Peatlands stand out with notably increased uncertainty, especially for SOC stock 30. This is attributed to the limited number of peatland soils sampled and to the variability of their soil characteristics in our study area. Further, the analysis of uncertainty distribution by land use indicates a threefold higher uncertainty in predicting SOC stock 30 compared to SOC stock 10. This underscores the complexity of the SOC stock prediction and highlights the need for more data acquisition and model calibration.

It is essential to note that when comparing the SOC stock ranges depicted in the final maps with those observed in the actual data, a noticeable trend emerges. The RF model appears to impose a limitation on the SOC stock range. For instance, in the observed dataset, the SOC stock 10 spans from 0.02 to 9.31 kg m⁻²; however, in the generated map, this range contracts to 0.85 to 6.75 kg m⁻². Similarly, examining the SOC stock 30 in the map, the range shifts from 2.85 to 19.50 kg m⁻², while in the observed data it increases from 0.03 to 29.90 kg m⁻². The differences in SOC stock ranges between the model's maps, and the actual data highlight the fact that the model does not perform perfectly for soils with very high or very low SOC stock amounts. This mismatch in accuracy is due to several factors that are partly, but not solely, due to the modeling process. The complicated mountain landscape makes the modeling harder, and the difficulties in collecting data in this area make the challenges higher. The complex terrain and the problems with obtaining representative samples both contribute to this issue. The SOC stock uncertainty maps reveal insights into predictive accuracy across diverse land covers and depths. These findings contribute to our understanding of carbon dynamics and underscore challenges in modeling complex terrains and land covers. Our research demonstrates that the Valchiavenna stocks a high amount of SOC. According to EIONET-SOIL data [50], Italian soils have an average SOC stock (0–30 cm) of 5.63 kg m⁻², compared to 8.72 kg m⁻² of the Valchiavenna soils in the same soil layer; this means that the soils of this valley provide important ecosystem services that should be taken into consideration to mitigate and adapt the impact of climate change and that it is necessary to manage soils carefully and protect them from degradation to avoid the loss of SOC, especially under climatic change scenarios.

5. Conclusions

The machine learning models applied in our research showed different performances, which is important in the context of DSM approaches to better understand the suitable modeling techniques. The RF model showed the best performance results compared to the other models. The results highlight the crucial role that machine learning models play in accurately capturing the complex relationships between SOC stock and environmental factors. Our research indicates that land cover and climatic factors are the most important predictors of SOC stock spatial distribution; geomorphometric parameters (slope, curvatures, and TWI) also demonstrated a significant impact in our mountainous environments. While the machine learning application yielded promising results in predicting the spatial distribution of SOC stock, the methodology revealed significant limitations, particularly in accurately estimating the entire range of SOC stock values.

The future development of this work may involve enhancing data collection in areas where uncertainties are great: the precision and accuracy of the output's maps might be improved by a future data-gathering design for the models' validation. Using additional predictors such as parent material maps and the history of land use may also improve the quality of the maps. The use of future projection scenarios of climate and land use changes would be a way to include temporal data to enhance knowledge of SOC dynamics over time in this environment, for the adoption of sustainable land management strategies. Therefore, the next step of this work is the prediction of SOC stocks under future climate change scenarios using machine learning and climatic models.

This study contributes to the understanding of SOC dynamics and mapping at a local scale: the knowledge of SOC stocks can be used by decision makers to protect regions with high actual carbon storage potential, such as mountain forests, peatlands, and grasslands, or zones at high risk of losing SOC stock, such as the upper belts of the valley. Finally, our research offers valuable information into the distribution of soil organic carbon stock in mountainous areas and can be used to assess ecosystem services, environmental management strategies, and support plans to mitigate climate change in these areas.

Supplementary Materials: The following supporting information can be downloaded at: https://www.mdpi.com/article/10.3390/land13010078/s1, References cited in [25,51].

Author Contributions: Methodology, S.A., C.F., M.M. and R.C.; Software, S.A. and M.M.; Validation, R.C.; Resources, S.A. and M.M.; Data curation, S.A., C.F., M.M. and R.C.; Writing—original draft, S.A.; Writing—review & editing, C.F. and R.C.; Supervision, C.F. and R.C.; Project administration, R.C. All authors have read and agreed to the published version of the manuscript.

Funding: This research received no external funding.

Data Availability Statement: The data are available upon request.

Conflicts of Interest: The authors declare no conflict of interest.

References

- 1. Baruck, J.; Nestroy, O.; Sartori, G.; Baize, D.; Traidl, R.; Vrščaj, B.; Bräm, E.; Gruber, F.E.; Heinrich, K.; Geitner, C. Soil classification and mapping in the Alps: The current state and future challenges. *Geoderma* **2016**, *264*, 312–331. [CrossRef]
- 2. Romeo, R.; Vita, A.; Manuelli, S.; Zanini, E.; Freppaz, M.; Stanchi, S. Understanding Mountain Soils: A Contribution from Mountain Areas to the International Year of Soils; FAO: Rome, Italy, 2015.
- Hartemink, A.E.; Gerzabek, M.H.; Lal, R.; McSweeney, K. Soil Carbon Research Priorities. In Soil Carbon; Hartemink, A.E., McSweeney, K., Eds.; Springer International Publishing: Berlin/Heidelberg, Germany, 2014; pp. 483–490. [CrossRef]
- Lal, R.; Smith, P.; Jungkunst, H.F.; Mitsch, W.J.; Lehmann, J.; Nair, P.R.; McBratney, A.B.; Sá, J.C.d.M.; Schneider, J.; Zinn, Y.L.; et al. The carbon sequestration potential of terrestrial ecosystems. *J. Soil Water Conserv.* 2018, 73, 145A–152A. [CrossRef]
- Alfthan, B.; Gjerdi, H.; Puikkonen, L.; Schoolmeester, T.; Andresen, M.; Gjerdi, H.L.; Jurek, M.; Semernya, L. Mountain Adaptation Outlook Series: Synthesis Report; UN Environment & GRID-Arendal: Arendal, Norway, 2018.
- 6. Adler, C.P.; Weste, I.; Bhatt, C.; Huggel, G.E. Climate Change 2022—Impacts, Adaptation and Vulnerability: Working Group II Contribution to the Sixth Assessment Report of the Intergovernmental Panel on Climate Change; Cambridge University Press: Cambridge, UK, 2023. [CrossRef]
- Hoffmann, U.; Hoffmann, T.; Jurasinski, G.; Glatzel, S.; Kuhn, N. Assessing the spatial variability of soil organic carbon stocks in an alpine setting (Grindelwald, Swiss Alps). *Geoderma* 2014, 232-234, 270–283. [CrossRef]
- Lagacherie, P.; McBratney, A. Chapter 1. Spatial soil information systems and spatial soil inference systems: Perspectives for Digital Soil Mapping. In *Developments in Soil Science*; Elsevier: Amsterdam, The Netherlands, 2007; Volume 31, pp. 3–22.
- 9. D'amico, M.E.; Freppaz, M.; Leonelli, G.; Bonifacio, E.; Zanini, E. Early stages of soil development on serpentinite: The proglacial area of the Verra Grande Glacier, Western Italian Alps. *J. Soils Sediments* **2014**, *15*, 1292–1310. [CrossRef]
- 10. D'Amico, M.E.; Freppaz, M.; Filippa, G.; Zanini, E. Vegetation influence on soil formation rate in a proglacial chronosequence (Lys Glacier, NW Italian Alps). *CATENA* **2014**, *113*, 122–137. [CrossRef]
- Wang, D.; Li, X.; Zou, D.; Wu, T.; Xu, H.; Hu, G.; Li, R.; Ding, Y.; Zhao, L.; Li, W.; et al. Modeling soil organic carbon spatial distribution for a complex terrain based on geographically weighted regression in the eastern Qinghai-Tibetan Plateau. *CATENA* 2020, 187, 104399. [CrossRef]
- 12. Ferré, C.; Caccianiga, M.; Zanzottera, M.; Comolli, R. Soil–plant interactions in a pasture of the Italian Alps. *J. Plant Interact.* 2020, 15, 39–49. [CrossRef]
- 13. Yang, R.-M.; Zhang, G.-L.; Liu, F.; Lu, Y.-Y.; Yang, F.; Yang, F.; Yang, M.; Zhao, Y.-G.; Li, D.-C. Comparison of boosted regression tree and random forest models for mapping topsoil organic carbon concentration in an alpine ecosystem. *Ecol. Indic.* **2016**, *60*, 870–878. [CrossRef]

- 14. Ballabio, C.; Fava, F.; Rosenmund, A. A plant ecology approach to digital soil mapping, improving the prediction of soil organic carbon content in alpine grasslands. *Geoderma* **2012**, *187–188*, 102–116. [CrossRef]
- 15. Baize, D. Naissance et Évolution des Sols: La Pédogenèse Expliquée Simplement; Quae Editions: Versailles, France, 2021; pp. 1–160.
- 16. Dorji, T.; Odeh, I.O.; Field, D.J.; Baillie, I.C. Digital soil mapping of soil organic carbon stocks under different land use and land cover types in montane ecosystems, Eastern Himalayas. *For. Ecol. Manag.* **2014**, *318*, 91–102. [CrossRef]
- 17. Li, Y.; Liu, W.; Feng, Q.; Zhu, M.; Yang, L.; Zhang, J. Effects of land use and land cover change on soil organic carbon storage in the Hexi regions, Northwest China. *J. Environ. Manag.* **2022**, *312*, 114911. [CrossRef]
- Vaysse, K.; Lagacherie, P. Using quantile regression forest to estimate uncertainty of digital soil mapping products. *Geoderma* 2017, 291, 55–64. [CrossRef]
- Heuvelink, G. Uncertainty quantification of GlobalSoilMap products. In Proceedings of the GlobalSoilMap. Basis of the Global spatial soil information system prodect of the 1st Globalsoilmap Conference, Orléans, France, 7–9 October 2013; pp. 335–340. [CrossRef]
- Peralta, G.; Di Paolo, L.; Luotto, I. Global Soil Organic Carbon Sequestration Potential Map—GSOCseq v.1.1.; FAO: Rome, Italy, 2022. [CrossRef]
- 21. Nations, Y.; Olmedo, G.F.; Reiter, S. Soil Organic Carbon Mapping Cookbook, 2nd ed.; FAO: Rome, Italy, 2018.
- 22. IUSS Working Group WRB. World Reference Base for Soil Resources. In *International Soil Classification System for Naming Soils and Creating Legends for Soil Maps*, 4th ed.; International Union of Soil Sciences (IUSS): Vienna, Austria, 2022.
- 23. Bc, H.; Rg, C. Climate downscaling: Techniques and application. Clim. Res. 1996, 7, 85–95.
- 24. Belloni, S.; Pelfini, M. Il gradiente termico in Lombardia, Dipartimento di scienze terra del università di Milano. *Acqua-Aria* **1987**, *4*, 441–447.
- DUSAF 7.0—Uso e Copertura del Suolo 2023—Geoportale della Lombardia. Available online: https://www.geoportale. regione.lombardia.it/news/-/asset_publisher/80SRILUddraK/content/dusaf-7.0-uso-e-copertura-del-suolo-2023 (accessed on 20 April 2023).
- Kuhn, M. Caret: Classification and Regression Training. R Package Version 6.0-86. Available online: https://CRAN.R-project. org/package=caret (accessed on 20 March 2023).
- 27. Friedman, J.H. Estimating Functions of Mixed Ordinal and Categorical Variables Using Adaptive Splines. 1991. Available online: https://apps.dtic.mil/sti/citations/ADA590939 (accessed on 25 October 2022).
- Rentschler, T.; Gries, P.; Behrens, T.; Bruelheide, H.; Kühn, P.; Seitz, S.; Shi, X.; Trogisch, S.; Scholten, T.; Schmidt, K. Comparison of catchment scale 3D and 2.5D modelling of soil organic carbon stocks in Jiangxi Province, PR China. *PLoS ONE* 2019, 14, e0220881. [CrossRef] [PubMed]
- 29. Wang, L.-J.; Cheng, H.; Yang, L.-C.; Zhao, Y.-G. Soil organic carbon mapping in cultivated land using model ensemble methods. *Arch. Agron. Soil Sci.* 2022, *68*, 1711–1725. [CrossRef]
- Zou, H.; Hastie, T. Regularization and variable selection via the elastic net. J. R. Stat. Soc. Stat. Methodol. Ser. B 2005, 67, 301–320. [CrossRef]
- Sirsat, M.; Cernadas, E.; Fernández-Delgado, M.; Barro, S. Automatic prediction of village-wise soil fertility for several nutrients in India using a wide range of regression methods. *Comput. Electron. Agric.* 2018, 154, 120–133. [CrossRef]
- Zhang, J.; Schmidt, M.G.; Heung, B.; Bulmer, C.E.; Knudby, A. Using an ensemble learning approach in digital soil mapping of soil pH for the Thompson-Okanagan region of British Columbia. *Can. J. Soil Sci.* 2022, 102, 579–596. [CrossRef]
- 33. Breiman, L. Random Forests. *Mach. Learn.* 2001, 45, 5–32. [CrossRef]
- 34. Wadoux, A.M.-C.; Minasny, B.; McBratney, A.B. Machine learning for digital soil mapping: Applications, challenges and suggested solutions. *Earth-Sci. Rev.* 2020, *210*, 103359. [CrossRef]
- 35. Khaledian, Y.; Miller, B.A. Selecting appropriate machine learning methods for digital soil mapping. *Appl. Math. Model.* **2020**, *81*, 401–418. [CrossRef]
- Were, K.; Bui, D.T.; Dick, Ø.B.; Singh, B.R. A comparative assessment of support vector regression, artificial neural networks, and random forests for predicting and mapping soil organic carbon stocks across an Afromontane landscape. *Ecol. Indic.* 2015, 52, 394–403. [CrossRef]
- 37. Cortes, C.; Vapnik, V. Support-vector networks. Mach. Learn. 1995, 20, 273–297. [CrossRef]
- Piikki, K.; Wetterlind, J.; Söderström, M.; Stenberg, B. Perspectives on validation in digital soil mapping of continuous attributes— A review. Soil Use Manag. 2020, 37, 7–21. [CrossRef]
- 39. Tajik, S.; Ayoubi, S.; Zeraatpisheh, M. Digital mapping of soil organic carbon using ensemble learning model in Mollisols of Hyrcanian forests, northern Iran. *Geoderma Reg.* 2020, 20, e00256. [CrossRef]
- 40. Ließ, M.; Schmidt, J.; Glaser, B. Improving the Spatial Prediction of Soil Organic Carbon Stocks in a Complex Tropical Mountain Landscape by Methodological Specifications in Machine Learning Approaches. *PLoS ONE* **2016**, *11*, e0153673. [CrossRef]
- Zhou, T.; Geng, Y.; Chen, J.; Pan, J.; Haase, D.; Lausch, A. High-resolution digital mapping of soil organic carbon and soil total nitrogen using DEM derivatives, Sentinel-1 and Sentinel-2 data based on machine learning algorithms. *Sci. Total. Environ.* 2020, 729, 138244. [CrossRef]
- Nguyen, T.T.; Pham, T.D.; Nguyen, C.T.; Delfos, J.; Archibald, R.; Dang, K.B.; Hoang, N.B.; Guo, W.; Ngo, H.H. A novel intelligence approach based active and ensemble learning for agricultural soil organic carbon prediction using multispectral and SAR data fusion. *Sci. Total. Environ.* 2022, 804, 150187. [CrossRef]

- Zeraatpisheh, M.; Ayoubi, S.; Mirbagheri, Z.; Mosaddeghi, M.R.; Xu, M. Spatial prediction of soil aggregate stability and soil organic carbon in aggregate fractions using machine learning algorithms and environmental variables. *Geoderma Reg.* 2021, 27, e00440. [CrossRef]
- 44. Yigini, Y.; Panagos, P. Assessment of soil organic carbon stocks under future climate and land cover changes in Europe. *Sci. Total. Environ.* **2016**, 557–558, 838–850. [CrossRef] [PubMed]
- 45. Ma, M.; Chang, R. Temperature drive the altitudinal change in soil carbon and nitrogen of montane forests: Implication for global warming. *CATENA* **2019**, *182*, 104126. [CrossRef]
- Odebiri, O.; Mutanga, O.; Odindi, J.; Peerbhay, K.; Dovey, S.; Ismail, R. Estimating soil organic carbon stocks under commercial forestry using topo-climate variables in KwaZulu-Natal, South Africa. *South Afr. J. Sci.* 2020, 116, 1–8. [CrossRef] [PubMed]
- 47. Parton, W.J.; Scurlock, J.M.O.; Ojima, D.S.; Schimel, D.S.; Hall, D.O.; Scopegram Group Members. Impact of climate change on grassland production and soil carbon worldwide. *Glob. Chang. Biol.* **1995**, *1*, 13–22. [CrossRef]
- Puche, N.J.B.; Kirschbaum, M.U.F.; Viovy, N.; Chabbi, A. Potential impacts of climate change on the productivity and soil carbon stocks of managed grasslands. *PLoS ONE* 2023, *18*, e0283370. [CrossRef]
- Chen, S.; Liang, Z.; Webster, R.; Zhang, G.; Zhou, Y.; Teng, H.; Hu, B.; Arrouays, D.; Shi, Z. A high-resolution map of soil pH in China made by hybrid modelling of sparse soil data and environmental covariates and its implications for pollution. *Sci. Total. Environ.* 2019, 655, 273–283. [CrossRef]
- 50. Panagos, P.; Hiederer, R.; Van Liedekerke, M.; Bampa, F. Estimating soil organic carbon in Europe based on data collected through an European network. *Ecol. Indic.* 2013, 24, 439–450. [CrossRef]
- 51. Available online: https://www.worldclim.org/ (accessed on 29 November 2023).

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