

Article

Enough Is Enough? Searching for the Optimal Sample Size to Monitor European Habitats: A Case Study from Coastal Sand Dunes

Simona Maccherini ¹, Giovanni Bacaro ², Enrico Tordoni ^{2,*}, Andrea Bertacchi ³, Paolo Castagnini ¹, Bruno Foggi ⁴, Matilde Gennai ⁴, Michele Mugnai ⁴, Simona Sarmati ¹ and Claudia Angiolini ¹

¹ Department of Life Sciences, University of Siena, Via P.A. Mattioli 4, 53100 Siena, Italy; simona.maccherini@unisi.it (S.M.); paolo.castagnini@unisi.it (P.C.); simona.sarmati@unisi.it (S.S.); claudia.angiolini@unisi.it (C.A.)

² Department of Life Sciences, University of Trieste, Via L. Giorgieri 10, 34127 Trieste, Italy; gbacaro@units.it

³ Department of Agricultural, Food and Agro-Environmental Sciences, Via del Borghetto 80, 56124 Pisa, Italy; andrea.bertacchi@unipi.it

⁴ Department of Biology, University of Florence, Via G. La Pira 4, 50121 Firenze, Italy; bruno.foggi@unifi.it (B.F.); matizgen@gmail.com (M.G.); michele.mugnai@unifi.it (M.M.)

* Correspondence: etordoni@units.it

Received: 29 January 2020; Accepted: 30 March 2020; Published: 2 April 2020



Abstract: A robust survey method that samples the main characteristics of plant assemblages is needed to assess the conservation status of European habitat in the Natura 2000 network. A measure of variability, called pseudo-multivariate dissimilarity-based standard error (MultSE), was recently proposed for assessing sample-size adequacy in ecological communities. Here, we used it on coastal sand dune systems in three Special Areas of Conservation (SACs) in Tuscany. Our aim was to assess the minimum number of replicates necessary to adequately characterize sand dune environments in terms of differences between habitats and SACs, after a preliminary baseline assessment of plant diversity. Analysis of α and β diversity indicated that especially between habitats the three SACs protect different plant communities. The study of the MultSE profiles showed that the minimum number of replicates needed to assess differences among habitats varied between 10 and 25 plots. Two-way PERMANOVA and SIMPER analysis on the full and reduced datasets confirmed that SACs and habitats host different plant communities, and that the contribution of the target species remained unchanged even with a reduced sample size. The proposed methodological approach can be used to develop cost-effective monitoring programs and it can be useful for plant ecologists and biodiversity managers for assessing ecosystem health and changes.

Keywords: coastal sand dune habitats; habitat directive 92/43/ECC; multivariate pseudo-standard error; plant diversity; sampling effort; species assemblage

1. Introduction

The Habitats Directive (92/43/EEC) obliges Member States of the European Union (EU) to monitor the conservation status of habitats and species listed in the Directive Annexes, and to report the results every 6 years [1,2]. In the Natura 2000 network, it is therefore essential to assess the actual distribution, natural variation and information on the quality of habitats in each site, and at the same time to provide solid data useful for objectively and quantitatively evaluating changes due to any conservation and/or restoration activities [1]. Nevertheless, no robust exhaustive method is available to detect the main characteristics of plant assemblages (presence and abundance) and to monitor habitat health along with habitat changes and conservation status. Probabilistic samplings and representativeness

assessment have not yet been used in monitoring schemes of European habitats [3], although some attempts based on species accumulation curves have been made [4]. Monitoring schemes have been established for many different purposes, and three aspects, namely sampling design, sample size and type of statistical analysis, are regarded as generally relevant in determining the scientific quality of the information derived from biodiversity monitoring [5]. Efficient sampling design is essential for accuracy, i.e., correspondence between real and measured biodiversity trends [6]. The sample size, namely the number of measurements made, is central for data precision (i.e., the ability to measure the same value under identical conditions). Finally, appropriate statistical analysis is needed to translate the data collected into useful information with relative uncertainty, which also depends on sampling design [7]. In plant ecology, when comparing attributes of plant communities in space or in time, it is fundamental to estimate how adequate a sample is for capturing the species diversity, taxonomic composition and relative abundance of the entire survey population, avoiding bias and dependency on sample size [8]. Sampling effort can influence the possibility of differentiating ecological communities [8–11], to the detriment of monitoring for community conservation and restoration purposes. In relation to habitat comparisons, these are usually evaluated through multivariate differences in the composition of plant communities (e.g., Anderson and Santana-Garcon [11], Tordoni et al. [12]), classifying assemblages and inferring species-environment relationships [8]. Anderson and Santana-Garcon [11] recently proposed a measure of precision for dissimilarity-based multivariate analysis of ecological communities called pseudo multivariate dissimilarity-based standard error (MultSE) for assessing sample-size adequacy within ecological communities. This statistic, which is the multivariate analog of the univariate standard error, measures the variability in the position of the sample centroid under repeated sampling for a given sample size in the space of a chosen dissimilarity matrix [11].

Here we apply this measure of multivariate precision in the context of habitat monitoring, aiming to simulate the effect of sampling size reduction on the discrimination power of statistical analysis based on multivariate (species composition) characteristics of biodiversity useful for discriminating habitats.

As a working example, we focused on a recent monitoring of protected coastal sand dune habitats in Special Areas of Conservation (SACs) of the Natura 2000 network in Tuscany, central Italy. Coastal sand dunes are usually characterized by marked vegetation zonation; the different zones often host rare or exclusive species important for dune formation and stabilization because they enhance sand deposition [13]. These habitats have suffered a heavy loss of biodiversity and fragmentation in recent decades, chiefly due to human encroachment in the form of tourism, urban sprawl and shoreline erosion whose consequences for biodiversity and related ecosystem services have been severe [14–19]. Furthermore, biological invasions pose a serious threat to sand dune ecosystems, threatening local plant diversity and related functional aspects and may lead to long-term alterations [12,20–22].

The general aim of this study was therefore to assess the number of replicates needed to adequately characterize sand dune environments in terms of differences between habitat types, SACs and habitat types within SACs, after obtaining a baseline assessment of plant diversity at habitat and site scale. We postulated that the decrease in MultSE would be more evident for species assemblages closer to the drift line than for those in the landward part of the beach.

2. Methods

2.1. Study Area and Sampling Design

The study was performed in three SACs on the Tyrrhenian coast of central Italy: (1) Dune Costiere del Parco dell'Uccellina (PU; SAC code IT51A0015, centroid coordinates: 11.0736E, 42.6361N; 158 ha), (2) Dune Litoranee di Torre del Lago (TL; IT5170001, 10.253889E, 43.828611N; 123 ha), and (3) Selva Pisana (SP; IT5170002, 10.306389E, 43.710278N; 9657 ha) (Figure 1). Geologically, the sites are mainly composed of Quaternary sand sediments, mostly Holocene [23]. Macrobioclimate is Mediterranean with upper meso-mediterranean thermotype and ombrotype ranging from upper dry (PU) to upper humid (TL and SP) [24]. EU habitat maps were detected from the available information provided by the HaSCITu (Habitat in the Sites of Conservation Interest in Tuscany) program of the Tuscan Regional

Administration (<http://www.regione.toscana.it/-/la-carta-degli-habitat-nei-siti-natura-2000-toscana>) where dune habitats form an intricate shifting mosaic hard to map and included in large patches that create a serious difficulty when planning sampling design. We solved these problems by adopting the European Nature Information System (EUNIS), at the third classification level, generally more inclusive respect to EU habitats because mainly based on physiognomic and physical attributes (i.e., B1.3—Shifting coastal dunes, including EU habitats 2110 and 2120). Moreover, EUNIS is a pan-European system for hierarchical habitat classification and its commonly accepted nomenclature facilitates comparison of the results between European countries [25,26].

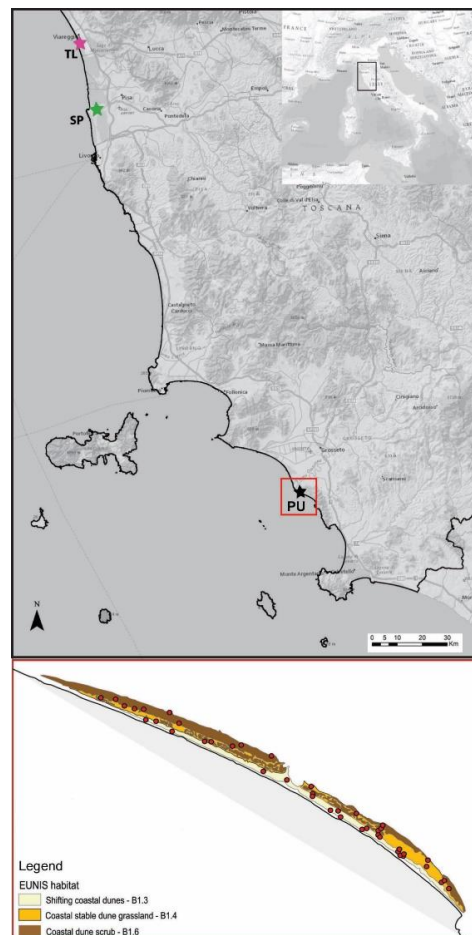


Figure 1. Locations of the Special Areas of Conservation (SACs) on the Tyrrhenian coast of Tuscany (colored stars) with respect to the Italian peninsula (upper right insert). Lower panel shows an example from a portion of the SAC “Dune costiere del Parco dell’ Uccellina” with European Nature Information System (EUNIS) habitats, red points indicate plots.

Three EUNIS habitat types (for more information see Davies et al. [27]), were mapped in the dune systems: (a) shifting coastal dunes (B1.3, including EU habitats 2110—embryonic shifting dunes—and 2120—shifting dunes along the shoreline with *Ammophila arenaria*); (b) coastal stable dune grasslands (B1.4, including EU habitats 2210—*Crucianellion maritimae* fixed beach dunes—and 2230—*Malcolmietalia* dune grasslands) and (c) coastal dune scrub (B1.6, including habitats 2230—*Malcolmietalia* dune grasslands—2240—*Brachypodietalia* dune grasslands with annuals and 2250—Coastal dunes with *Juniperus* sp. pl.). The target species for each EUNIS habitat type are reported in Table 1. 206 squared plots of 4 m² were randomly allocated in proportion to the surface area of the EUNIS habitats in each SAC (1 plot/ha for B1.3 and B1.4 and 1 plot/3-ha for B1.6; see Table 2 for further details) where occurrence and abundance (% visual cover estimation) of vascular plant species was recorded.

Table 1. The three EUNIS habitat types with corresponding EU Habitats. For each EUNIS habitat type, the target species are indicated according to the Italian interpretation manual of the Habitats Directive [28]. Asterisk denotes priority habitats according to Habitats Directive.

EUNIS Habitat	EU Habitat (Directive 92/43/EEC)	Target Species
Shifting coastal dunes (B1.3)	2110- Embryonic shifting dunes 2120 -Shifting dunes along the shoreline with <i>Ammophila arenaria</i> (white dunes)	<i>Ammophila arenaria</i> , <i>Anthemis maritima</i> , <i>Calystegia soldanella</i> , <i>Cyperus capitatus</i> , <i>Echinophora spinosa</i> , <i>Elymus farctus</i> , <i>Eryngium maritimum</i> , <i>Euphorbia peplis</i> , <i>Euphorbia paralias</i> , <i>Lotus creticus</i> , <i>L. cytoides</i> , <i>Medicago marina</i> , <i>Othantus maritimus</i> , <i>Pancratium maritimum</i> , <i>Polygonum maritimum</i> , <i>Solidago litoralis</i> , <i>Sporobolus arenarius</i> , <i>Stachys maritima</i>
Coastal stable dune grassland (B1.4)	2210 -Fixed coastal dunes with herbaceous vegetation (grey dunes) 2230 - <i>Malcolmietalia</i> dune grasslands	<i>Anthemis mixta</i> , <i>Bromus diandrus</i> , <i>Corrigiola telephifolia</i> , <i>Corynephorus divaricatus</i> , <i>Crucianella maritima</i> , <i>Cutandia maritima</i> , <i>Daucus pumilus</i> , <i>Eva xpygmaea</i> , <i>Helichrysum stoechas</i> , <i>Lagurus ovatus</i> , <i>Lupinus angustifolius</i> , <i>Malcolmia ramosissima</i> , <i>Maresia nana</i> , <i>Matthiola tricuspidata</i> , <i>Medicago littoralis</i> , <i>Ononis variegata</i> , <i>Pancratium maritimum</i> , <i>Phleum arenarium</i> , <i>Polycarpon diphylum</i> , <i>Pseudorhiza pumila</i> , <i>Pycnocomon rutifolium</i> , <i>Seseli tortuosum</i> , <i>Silene canescens</i> , <i>S. gallica</i> , <i>S. niceensis</i> , <i>Sonchus bulbosus</i> , <i>Thesium humile</i> , <i>Vulpia fasciculata</i>
Coastal dune scrub (B1.6)	2230 - <i>Malcolmietalia</i> dune grasslands 2240 - <i>Brachypodietalia</i> dune grasslands with annuals 2250* - Coastal dunes with <i>Juniperus</i> spp.	<i>Aetheorhiza bulbosa</i> , <i>Aira elegans</i> , <i>Andryala integrifolia</i> , <i>Anthyllis barba-jovis</i> , <i>Asparagus acutifolius</i> , <i>Brachypodium distachyum</i> , <i>Briza maxima</i> , <i>Clematis flammula</i> , <i>Corynephorus divaricatus</i> , <i>Corrigiola telephifolia</i> , <i>Cutandia maritima</i> , <i>Evax pygmaea</i> , <i>Galium divaricatum</i> , <i>Juniperus communis</i> , <i>Juniperus macrocarpa</i> , <i>J. turbinata</i> , <i>Lagurus ovatus</i> , <i>Lonicera implexa</i> , <i>Lotus angustissimus</i> , <i>Lupinus angustifolius</i> , <i>Malcolmia ramosissima</i> , <i>Maresia nana</i> , <i>Matthiola tricuspidata</i> , <i>Medicago littoralis</i> , <i>Myrtus communis</i> , <i>Ononis variegata</i> , <i>Ornithopus compressus</i> , <i>Phillyrea angustifolia</i> , <i>P. latifolia</i> , <i>Phleum arenarium</i> , <i>Pistacia lentiscus</i> , <i>Polycarpon diphylum</i> , <i>Plantago lagopus</i> , <i>P. bellardii</i> , <i>Prasium majus</i> , <i>Pseudorhiza pumila</i> , <i>Rhamnus alaternus</i> , <i>Rubia peregrina</i> , <i>Rumex bucephalophorus</i> , <i>Ruscus aculeatus</i> , <i>Silene canescens</i> , <i>S. nicaensis</i> , <i>S. gallica</i> , <i>Smilax aspera</i> , <i>Tuberaria guttata</i> , <i>Vulpia membranacea</i>

Table 2. Descriptive statistics of the study areas (area, number of plots and plant species richness) according to EUNIS habitat type, SAC and EUNIS type within SAC.

Term	Levels	Name	Area (Ha)	N° Plots	Average Richness (Min-Max)
HABITAT	B1.3	Shifting coastal dunes	54.70	54	7 (2–12)
	B1.4	Stable dune grasslands	121.94	124	10 (3–21)
	B1.6	Coastal dune scrubs	42.33	28	9 (2–17)
SAC	TL	Torre del Lago	81.26	83	11 (4–21)
	PU	Parco dell’Uccellina	73.33	54	7 (3–12)
	SP	Selva Pisana	64.39	69	8 (2–17)
HABITAT:SAC	B1.3:PU	Shifting coastal dunes: Parco dell’Uccellina	20.11	21	7 (3–12)
	B1.3:SP	Shifting coastal dunes: Selva Pisana	19.26	17	6 (2–11)
	B1.3:TL	Shifting coastal dunes: Torre del Lago	15.32	16	8 (4–11)
	B1.4:PU	Stable dune grasslands: Parco dell’Uccellina	20.08	21	6 (3–11)
	B1.4:SP	Stable dune grasslands: Selva Pisana	40.96	38	8 (3–16)
	B1.4:TL	Stable dune grasslands: Torre del Lago	60.91	65	12 (6–21)
	B1.6:PU	Coastal dune scrub: Parco dell’Uccellina	33.14	12	8 (4–12)
	B1.6:SP	Coastal dune scrub: Selva Pisana	41.62	14	9 (2–17)
	B1.6:TL	Coastal dune scrub: Torre del Lago	5.03	2	12 (11–12)

2.2. Workflow of the Analysis

Habitats and SACs were first characterized in terms of species richness and compositional similarity. These diversity characteristics were obtained from our whole dataset and served as a baseline for evaluating how much information was lost when sampling size was reduced. The following analyses were performed.

2.2.1. Description of Diversity Patterns

We first evaluated overall sampling efficiency and diversity, computing classical sample-based rarefaction curves (RC) and spatial-explicit rarefaction curves (SER, see Chiarucci et al. [29] for more details on methodology) using the function available in Bacaro et al. [30], and those available in the ‘vegan’ R package [31]. Unlike RCs which do not account for spatial autocorrelation, SERs take adjacency of sampling units into account and consequently the spatial structure of the data [32]. RCs and SERs were calculated for the three habitats separately (across SACs) and at SAC scale (across habitats), RC was also computed for the whole dataset (Random curve). The consistency of species diversity patterns across spatial scales (plot, habitat and site) was also assessed using additive partitioning techniques [33,34] in R package ‘vegan’ [31]. To test for significance, a null model was generated, permutating the original data matrix 999 times to assess deviation from random expectations.

2.2.2. Species Composition Variation across Spatial Scales

To investigate how SAC (fixed effect, three levels: PU, TL and SP), EUNIS habitats (fixed effect, three levels: B1.3, B1.4 and B1.6) and their interaction determined community composition, we performed a permutational multivariate analysis of variance (PERMANOVA, [35]) on the whole dataset, where: (a) Bray-Curtis dissimilarities were calculated on $\log(x + 1)$ -transformed abundance data and (b) Jaccard dissimilarity on species occurrences (i.e., p/a matrix). All tests were performed using 9999 permutations of residuals under a reduced model and $\alpha = 0.05$; this method yields the best power and the most accurate type I error for multi-factorial designs [36]. The significant interaction term was then investigated using a posteriori pairwise comparisons with the PERMANOVA t statistic and 9999 permutations. We also calculated the pseudo multivariate variance component (expressed

as percentages) for each source of variation. The analysis was performed using the PERMANOVA routine in the PRIMER v6 computer program [37], including the add-on package PERMANOVA+ [38]. After PERMANOVA, the same data underwent similarity percentage analysis (SIMPER, [39]) to identify the species that contribute most to the average Bray–Curtis dissimilarity between habitats across SACs.

2.2.3. Measurement of Multivariate Precision and Associated Dissimilarities (MultSE)

MultSE was calculated according to Equations (1) and (2) using the code and functions available in [11] and the method described therein. To compute MultSE, the community composition matrix was $\log(x + 1)$ transformed and Bray–Curtis dissimilarity calculated. This statistic was calculated between habitats (pooling all SACs) and between SACs (pooling all habitats). It can be considered a direct analogue of univariate SE, but is based on the chosen dissimilarity measure, thus providing a powerful tool to examine the relative precision of a sampling procedure. This statistic was calculated as follows:

$$\text{MultSE} = \sqrt{V/n} \quad (1)$$

where V is a multivariate measure of pseudo variance in the space of the chosen dissimilarity measure:

$$V = \frac{1}{(n-1)} \sum_{i=1}^{(n-1)} \sum_{j=(i+1)}^n d_{ij}^2/n \quad (2)$$

where n is the number of sampling units and d represents the squared distance between individual sampling points to their centroid, given a chosen dissimilarity measure. We computed 95% confidence intervals by a double resampling method based on permutations for means calculation and bias-adjusted bootstrap-based error bars (5000 resamples). As in the case of its univariate counterpart, when the profile of MultSE as a function of increasing sampling size reaches an asymptote, this measure can be considered indicative of adequate sampling effort. Beyond this threshold, in fact, the relationship becomes flat and only negligible changes in sampling precision can be expected. The breaking point of the MultSE profile was estimated using R package ‘segmented’ [40,41]. The statistic is unbiased if and only if the sampling procedure is representative of the statistical population and an equal probability is given to each sample by appropriate sampling methods [11]. In a similar way, we computed the expected increase in community dissimilarity related to sampling effort. In other words, using simple randomization procedures, we randomly extracted an increasing number of replicates (from 1 to $n-1$) 999 times, and we calculated the average Bray–Curtis and Jaccard dissimilarity profiles that indicate the centroid of the species assemblage, for a given sampling size.

2.2.4. Effect of Sampling Size Reduction

In order to describe how a reduction in sampling size affects the ecological conclusions obtained from the analysis of the complete dataset, and in particular the ability to: (1) characterize the species composition of single habitats, (2) discriminate compositional differences between habitats, and (3) provide an acceptable level of sampling precision under reduced sampling effort, we resampled plots virtually by means of a stratified random sampling approach. The plots were resampled from the whole species assemblage, using the number of plots derived from MultSE estimation for each level of the crossed factor SAC \times habitat. Then this subset was used to compute PERMANOVA and do SIMPER analysis as described above. PERMANOVA and SIMPER were calculated at each randomization (999), and the resulting statistics were compared with those of the whole dataset. All the analyses, except for PERMANOVA, were computed using R 3.6.1 [42].

3. Results

Rarefaction curves (Figure 2) did not reach an asymptote: their comparison revealed that coastal dune scrubs (B1.6) accumulated more species than shifting coastal dunes (B1.3) and dune grasslands (B1.4) across sampling sites; whereas, among SACs, TL showed the highest plant richness followed by SP and PU. It is worth noting that the RC of dune scrubs lay above the random curve, whereas the corresponding SER did not, again corroborating the need to include the spatial structure of the data in order to avoid biased results. Additive partitioning showed that overall diversity in each site was mainly due to variation among habitats rather than plots (Figure 3).

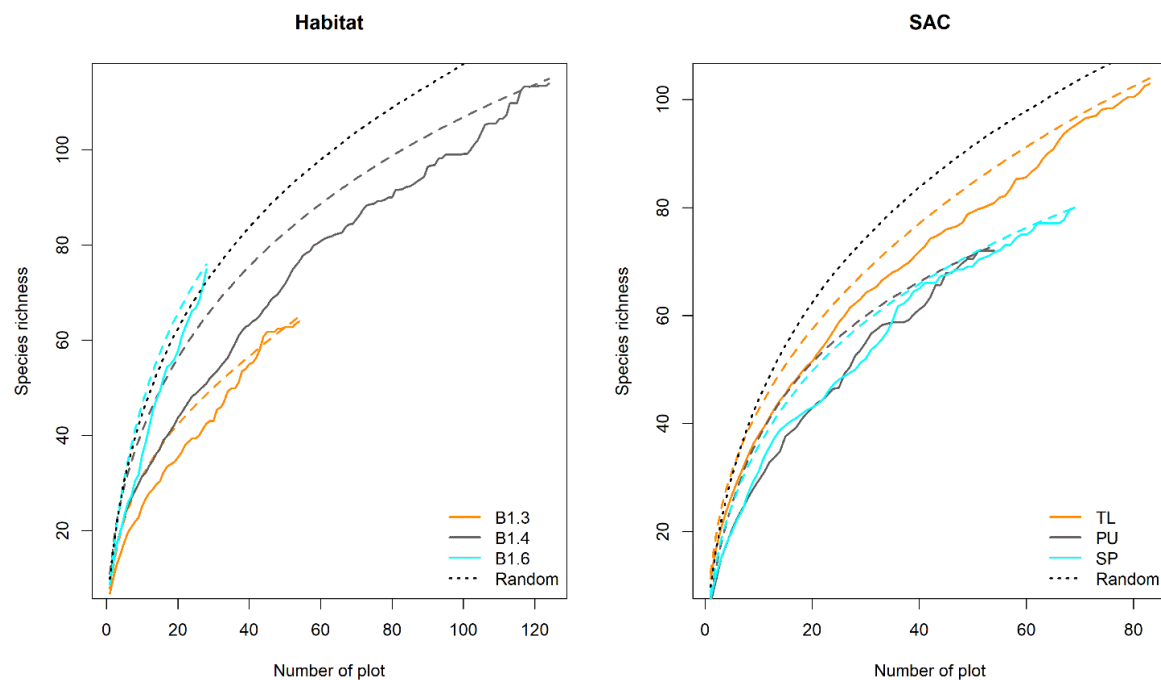


Figure 2. Spatially-explicit rarefaction curves (SERs, solid lines) and plot-based rarefaction curves (RCs, dashed lines) calculated among habitats (left) and SACs (right). The dotted line (Random) represents the RC calculated from the whole species pool. Please note that this curve was truncated according to the maximum number of plots detected in each factor. TL = Torre del Lago; PU = Parco Uccellina; SV = Selva Pisana.

Two-way PERMANOVA revealed that all sources of variation significantly affected community composition; pairwise comparisons for the interaction $SAC \times habitat$ were significant for all pairs examined, except coastal dune scrub of SP and TL when considering abundance data (Table 3, Table S1). Figure 4 and Figure S1 summarize the relationship between MultSE and the number of replicates among habitats in each SAC. Based on this, we estimated that among habitats, approximately 10 plots were enough to grasp overall diversity in the study area, even though slight differences could be detected in relation to habitat and SAC (Table 4). The dissimilarity profiles computed among habitats, pooling all SACs, flattened out at about 20 plots (Figure 5), suggesting an effect of the SAC on the plant species pool. Notably, there were different patterns of diversity accumulation when abundances or incidence matrix were used, especially in habitat B1.6. Interestingly, the overall signal remained constant in the reduced dataset which was approximately one-quarter of the original dataset (54 vs. 206 plots, Table 5) whether we considered species abundance or species presence/absence. EUNIS habitat type accounted for the highest variance component in both datasets (original vs. reduced dataset), further corroborating the output in Figure 3; on the other hand, there were slight differences in the role of SAC and its interaction with habitat, depending on the type of data (abundance vs. presence/absence data). The contribution of the single species characterizing each habitat was concordant in the two datasets (Table S2 and Figure S2 of supplementary material).

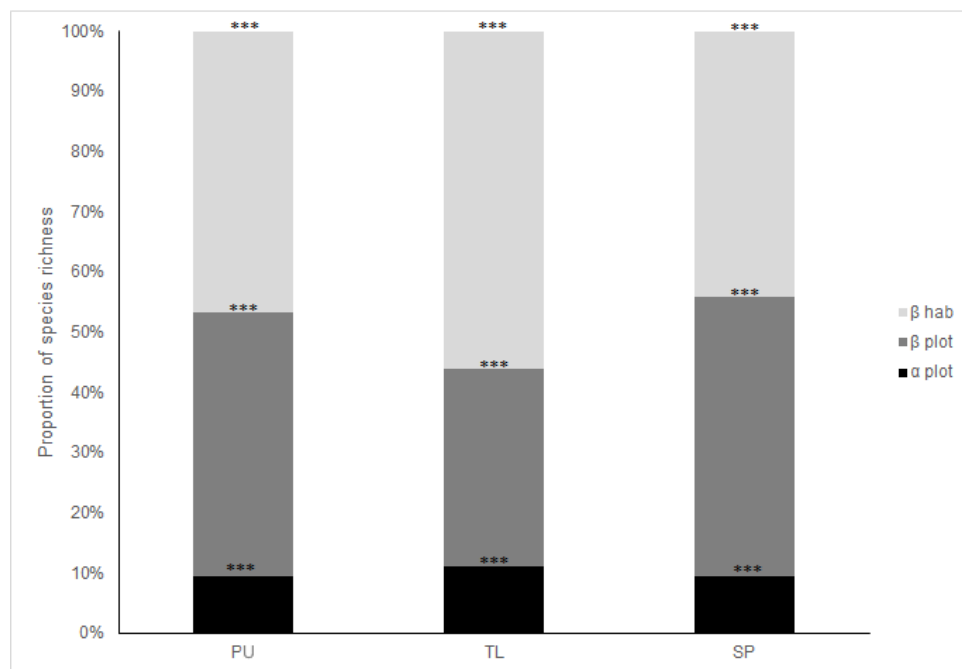


Figure 3. Additive partitioning of diversity (% of total species richness) in each sampling site across different spatial scales (within plot— α plot; habitat— β plot, and site— β hab). A null model was also computed to assess significant departures from random expectations (***) $p < 0.001$; ** $p < 0.01$; * $p < 0.05$). TL = Torre del Lago; PU = Parco Uccellina; SV = Selva Pisana.

Table 3. PERMANOVA results on percentage cover of species and occurrence data in 206 dune plots. SAC—Special Areas of Conservation. The main effects and their interactions were tested separately under a null model. *** $p \leq 0.001$.

Source of Variation	df	MS	F	Variance Components (%)
<i>Abundance</i>				
SAC	2	14,817	6.17 ***	15.58
Habitat	2	42,946	17.86 ***	25.68
SAC \times Habitat	4	8988	3.74 ***	16.97
Residual	197	2405		41.77
Total	205			
<i>presence/absence</i>				
SAC	2	19,916	8.99 ***	18.53
Habitat	2	40,571	18.31 ***	24.86
SAC \times Habitat	4	8653	3.91 ***	16.70
Residual	197	2215		39.91
Total	205			

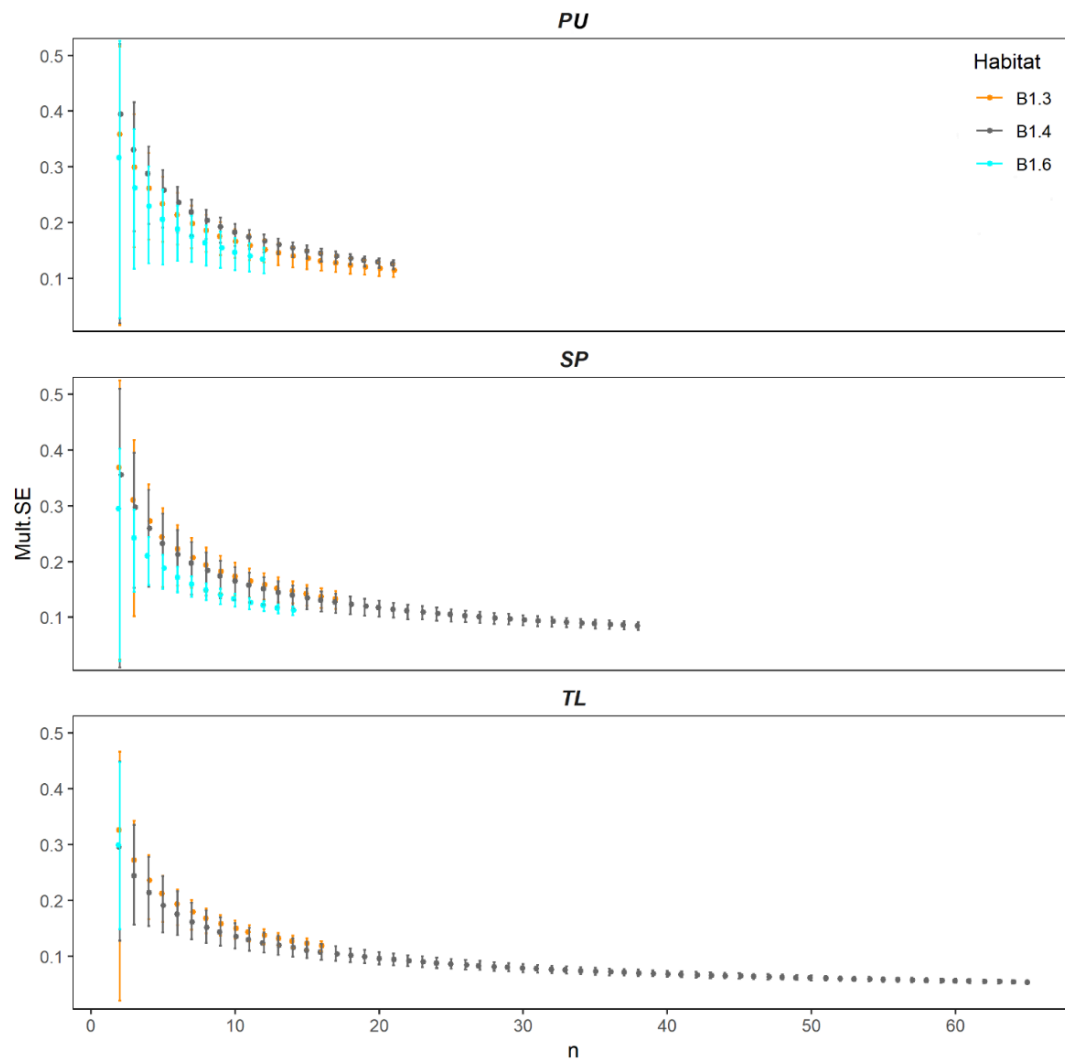


Figure 4. Profile of MultSE based on Bray–Curtis dissimilarity for each EUNIS habitat type within the three sampling areas. n = number of plots. TL = Torre del Lago; PU = Parco Uccellina; SV = Selva Pisana.

Table 4. Estimated number of plots and associated standard error indicating where precision stabilizes considering habitat, SAC and their interaction. Breaking points of MultSE (i.e., where the linear relation changes) were estimated by segmented relationships.

Levels	Name	Abundance	Presence/Absence
B1.3	Shifting coastal dunes	9 ± 0.3	10 ± 0.4
B1.4	Stable dune grasslands	25 ± 2.2	14 ± 0.4
B1.6	Coastal dune scrubs	7 ± 0.3	7 ± 0.3
TL	Torre del Lago	12 ± 0.4	11 ± 0.4
PU	Parco dell'Uccellina	9 ± 0.3	9 ± 0.4
SP	SelvaPisana	10 ± 0.4	10 ± 0.4
B1.3:PU	Shifting coastal dunes: Parco dell'Uccellina	6 ± 0.3	6 ± 0.3
B1.3:SP	Shifting coastal dunes: SelvaPisana	5 ± 0.3	5 ± 0.3
B1.3:TL	Shifting coastal dunes: Torre del Lago	5 ± 0.3	5 ± 0.3
B1.4:PU	Stable dune grasslands: Parco dell'Uccellina	6 ± 0.3	6 ± 0.3
B1.4:SP	Stable dune grasslands: SelvaPisana	8 ± 0.3	8 ± 0.3
B1.4:TL	Stable dune grasslands: Torre del Lago	10 ± 0.4	10 ± 0.4
B1.6:PU	Coastal dune scrub: Parco dell'Uccellina	4 ± 0.2	4 ± 0.2
B1.6:SP	Coastal dune scrub: Selva Pisana	5 ± 0.3	5 ± 0.2
B1.6:TL	Coastal dune scrub: Torre del Lago	-	-

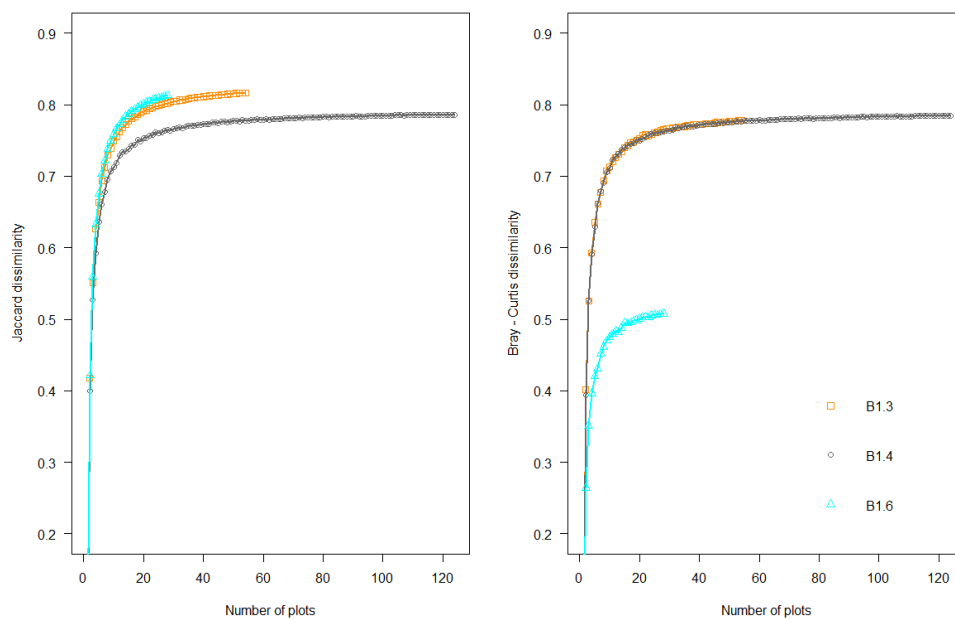


Figure 5. Profile of overall diversity across habitats based on Jaccard dissimilarity (**left**) and Bray–Curtis dissimilarity (**right**).

Table 5. Summary statistics of PERMANOVA results derived from 999 random resampling of the original dataset based on the plot numbers given by the decay of MultSE both using abundance and presence/absence data.

Term	Statistic	Abundance			Presence/Absence		
		F	R ²	Rate of Significance ($p < 0.05$)	F	R ²	Rate of Significance ($p < 0.05$)
SAC	Min.	2.62	0.06	100%	2.31	0.08	100%
	1st quart.	3.86	0.09		3.11	0.10	
	Median	4.31	0.01		3.38	0.10	
	3rd quart.	4.77	0.11		3.65	0.11	
	Max.	6.84	0.15		5.22	0.15	
Habitat	Min.	4.67	0.13	100%	2.74	0.09	100%
	1st quart.	7.20	0.19		4.03	0.13	
	Median	8.11	0.21		4.47	0.14	
	3rd quart.	9.06	0.23		4.93	0.15	
	Max.	12.97	0.30		7.31	0.20	
SAC × Habitat	Min.	1.11	0.06	94.3%	1.15	0.08	98.7%
	1st quart.	1.67	0.09		1.53	0.10	
	Median	1.88	0.01		1.67	0.10	
	3rd quart.	2.13	0.11		1.83	0.11	
	Max.	3.38	0.15		2.70	0.14	

4. Discussion

4.1. Diversity Patterns Across Habitats and SACs

Our analysis provided evidence that overall, the three EUNIS habitats host different plant communities and that the three SACs protect different dune vegetation. This information is confirmed by the diversity patterns and variations in species composition between different scales of analysis. Regarding diversity, the rarefaction curves described how species richness differed substantially between habitats, where an increasing number of species for a given sample size was detected moving from shifting dunes (B1.3) to dune grasslands (B1.4) and dune scrub (B1.6). Other studies

have shown a correlation between species diversity and the coast-to-inland environmental gradient. Indeed, species richness generally increases as one moves from the annual communities of the upper beach (more unstable habitats and stressful conditions) to the fixed dunes (more stable environments) along the psammophile sequence [43–46] and our results are in line with these findings. Dune scrub has higher richness, as this habitat is dominated by shrub communities characterized by open areas with many annual species, as already shown by Acosta et al. [44]. However, interesting differences in rarefaction curve patterns were observed with respect to those of Ciccarelli and Bacaro [46] for the same study area and habitats: while these authors observed an asymptotic pattern for all curves, our rarefaction curves showed a constantly increasing trend. Sampling design and sampling size can be considered the main factors responsible for these differences: while our study was based on a randomly chosen plots, the study of the authors was based on contiguous transects. The overall sampling effort was also different: only 206 plots in our study versus a total of 980 plots. The total number of species collected per habitat in the two studies was nevertheless comparable, implying good sampling design efficiency.

α and β diversity partitioning across spatial scales showed substantial similarity for PU and SP, where β_{habitat} and β_{plot} components gave the same contribution to total gamma diversity. Notably, β_{habitat} in TL showed the highest relative contribution to total γ diversity, meaning that in this SAC, communities are clearly distinguished from each other: TL has a more stable coastal configuration, allowing a more ample dune system with well-defined habitats. The second-largest variation was found at EUNIS habitat level in relation to the strong environmental gradients in coastal dunes. The latter ensure the development of floristically different vegetation types that host species with a narrow ecological range [47–50] and the existence of vegetation zonation which inevitably controls not only diversity patterns [13,44] but also community structure [26].

Based on the sea-inland environmental gradient, the species that contribute to average dissimilarities between EUNIS habitats are target species for EU dune habitats (key species or diagnostic species *sensu* Biondi et al. [28], Angiolini et al. [50], Sperandii et al. [26]). *Elymus farctus* and *Ammophila arenaria* are considered constructor species of embryonic and mobile dunes, respectively, linked to shifting coastal dune habitat (B1.3). *Helichrysum stoechas*, *Seseli tortuosum* and *Vulpia fasciculata* are considered structural species of fixed beach dune garrigues or grasslands, characterizing B1.4. Woody species such as *Juniperus oxycedrus* subsp. *macrocarpa*, *Smilax aspera* and *Pistacia lentiscus* are typical of fixed coastal dunes dominated by *Juniperus* sp. pl. such as in dune scrubs (B1.6). At the European scale, this is mostly true for communities of mobile dunes, even if at regional/local scale, as in our case study, communities of fixed dunes also have similar floristic compositions [47,51]. This confirms that in habitats with strong environmental gradients, local variability is more important in shaping communities than larger scale variability. This was also found in Tuscan badland environments, where salinity and erosion create a complex mosaic of habitat types recognized at a local scale [52,53].

4.2. Spatial Variation of Reduced Dataset

The differences between habitats and SACs and between habitats within SACs remained appreciable with the reduced dataset which was approximately one-quarter of the original dataset (54 vs. 206 plots). This suggests that a reduced sample can also capture the structure and composition of plant communities and environmental gradients, at least in the type of community considered here. A reduced dataset can, therefore, provide information on habitat conditions and be useful for monitoring habitat conservation status over time, since ecological groups such as target species contribute substantially to ecosystem structure and function, being particularly responsive to threats and habitat modifications [26,50,54]. Obviously, in order to ensure its representativeness, the reduced dataset must be based on information obtained by adequate sampling methods.

As expected, a smaller number of replicates was enough to distinguish B1.3 (shifting coastal dunes) from other EUNIS habitats. In line with Acosta et al. [44] and Angiolini et al. [50], this habitat showed the lowest number of species per plot. Conversely, greater sampling effort was needed in

dune grasslands (B1.4), probably due to the presence of more stable and heterogeneous communities favored by deeper, more evolved soil along with lower exposure to the already mentioned limiting factors [13,55]. We also observed that a greater number of replicates was needed to characterize habitat when site variation was not considered. This was somehow expected, since SACs contribute to community composition, even if the greatest percentage of variance is accounted for by habitat type (see Table 3). In addition, another explanation for the low number of replicates needed to characterize dune community structure can be explained by the relatively low number of species thanks to species' highly restricted ecological preferences and specific functional features [22].

4.3. The Lesson We Learned

To date, few monitoring programs reach the standards necessary (e.g., survey design, hypothesis formulation, statistical power) to be considered statistically unbiased [7]. In recent years, the use of a probabilistic sampling design has proved useful in monitoring natural vegetation both for collecting reliable quantitative information and for representing of different physiognomic vegetation types, also allowing for generalizations [56]. In this light, it has now been widely acknowledged that it is not appropriate to examine biodiversity patterns such as species abundance by preferential sampling (e.g., Diekmann et al. [57]; Lájer [58]). It has also been shown that preferential sampling may lead to biased results by narrowing the environmental gradient or artificially restricting the species pool which may cause overestimation of rare and underestimation of common ones such as generalist or alien species [4,59]. Nonetheless, an urgent need for a quantitative measure of sampling adequacy in plant communities is advocated, especially by conservation technicians and plant ecologists. Quantitatively speaking, for instance, Stohlgren et al. [60] found that as few as ten replicates of 1 m² are satisfactory to detect fine-scale species richness patterns along an elevation gradient in the Rocky Mountains, and similar outcomes have been reported for semi-natural vegetation in Eastern Europe [61]. Our results agree with these studies suggesting that sampling effort in dune ecosystems could be moderately reduced, because biodiversity patterns remain quite stable and detection bias is relatively low. As a cautionary note, however, it worth noting that this method may be inappropriate for taxa with low detectability such as rare or cryptic species [61]; in this case, a more intense sampling effort is needed (e.g., Bried and Pellet [62]).

5. Conclusions

To the best of our knowledge, this is the first attempt to apply a MultSE approach to terrestrial habitats. MultSE proved useful for characterizing sampling adequacy and habitat features in a cost-effective way and highlighted that the three SACs protect different plant communities. The methodology proposed here evaluated different aspects of the monitoring of plant communities: in particular, it offers a flexible solution for plant ecologists and biodiversity managers wishing to optimize sampling design for habitat monitoring, facilitating the assessment of habitat quality and conservation status over time as specified in Art. 11 and 17 of EEC Directive 92/43. The type of response variable (abundance or occurrence) affects the restoration and conservation monitoring [63] and sampling costs, but this does not seem to occur for dune habitats (at least as far as the ability to distinguish different habitats and SACs is concerned). In any case, sand dune environments are usually characterized by a few abundant species that structure community composition facilitating the collection of abundance data which is often preferred for a quantitative assessment of the effects of conservation measures or habitat changes through time. Thus, before implementing habitat surveys, we recommend plant ecologists and biodiversity managers to consider the following aspects:

- conduct pilot studies testing different sampling probabilistic methods;
- wisely plan sampling efforts taking into account resource availability (i.e., time and costs);

- approaches based on plant functional traits and remote sensing may provide novel insights on ecosystem functioning, the latter revealed to be also a cost-effective way to handle biodiversity measurements and to predict species changes through time.

In conclusion, we advocate the use of the present methodological approach in other habitat types and geographical areas in order to test its reproducibility and effectiveness and to develop cost-effective monitoring programs for other European protected areas under Habitats Directive.

Supplementary Materials: The following are available online at <http://www.mdpi.com/1424-2818/12/4/138/s1>, Figure S1: Profile of *pseudo* multivariate dissimilarity-based standard error (MultSE) based on Jaccard dissimilarity for each EUNIS habitat type within the three sampling areas. PU = Parco dell'Uccellina, SP = Selva Pisana, TL = Torre del Lago, Figure S2: SIMPER output derived from 999 random resampling of the original community based on plot numbers given by decay of MultSE. Bar plots represent the proportion of resampling for each species, red dashed lines indicate an 80% contribution, Table S1: Results of PERMANOVA pairwise test for the interaction SAC \times habitat. PU—Parco dell'Uccellina—TL—Torre del Lago—SP—Selva Pisana; B1.3—shifting coastal dunes, B1.4—stable dune grassland—B1.6—coastal dune scrub", Table S2: "SIMPER output reporting species contribution to average between-group dissimilarity (Average), the corresponding standard deviation (SD), average abundances per group (Av.a, Av.b), and the cumulative contribution of the species (Cum. Contr.). NB. only species contributing up to 50% are reported".

Author Contributions: Conceptualization and methodology, S.M., B.F., C.A.; formal analysis, E.T., G.B., S.M.; data curation, A.B., P.C., S.S., M.G., M.M.; writing—original draft preparation, E.T., S.M., G.B., C.A.; writing—review and editing, A.B., P.C., B.F., S.S., M.G., M.M. All authors have read and agreed to the published version of the manuscript.

Funding: This research was supported by the Tuscany Region through the project Monito-Rare (Monitoraggio di specie e habitat terrestri di interesse comunitario ai sensi della Direttiva Habitat 92/43/CEE).

Acknowledgments: We thank Daniela Ciccarelli, Giulio Ferretti and Daniele Viciani for their help during fieldwork.

Conflicts of Interest: The authors declare no conflict of interest.

References

1. Evans, D.; Arvela, M. *Assessment and Reporting under Article 17 of the Habitats Directive. Explanatory Notes & Guidelines for the Period 2007–2012; Final Version*; European Commission: Brussels, Belgium, 2011.
2. Alberdi, I.; Nunes, L.; Kovac, M.; Bonheme, I.; Cañellas, I.; Rego, F.C.; Dias, S.; Duarte, I.; Notarangelo, M.; Rizzo, M.; et al. The conservation status assessment of Natura 2000 forest habitats in Europe: Capabilities, potentials and challenges of national forest inventories data. *Ann. For. Sci.* **2019**, *76*, 34. [\[CrossRef\]](#)
3. Ellwanger, G.; Runge, S.; Wagner, M.; Ackermann, W.; Neukirchen, M.; Frederking, W.; Müller, C.; Ssymank, A.; Sukopp, U. Current status of habitat monitoring in the European Union according to Article 17 of the Habitats Directive, with an emphasis on habitat structure and functions and on Germany. *Nat. Conserv.* **2018**, *29*, 57–78. [\[CrossRef\]](#)
4. Del Vecchio, S.; Fantinato, E.; Silan, G.; Buffa, G. Trade-offs between sampling effort and data quality in habitat monitoring. *Biodivers. Conserv.* **2019**, *28*, 55–73. [\[CrossRef\]](#)
5. Yoccoz, N.G.; Nichols, J.D.; Boulinier, T. Monitoring of biological diversity in space and time. *Trends Ecol. Evol.* **2001**, *16*, 446–453. [\[CrossRef\]](#)
6. Balmford, A.; Green, R.E.; Jenkins, M. Measuring the changing state of nature. *Trends Ecol. Evol.* **2003**, *18*, 326–330. [\[CrossRef\]](#)
7. Legg, C.J.; Nagy, L. Why most conservation monitoring is, but need not be, a waste of time. *J. Environ. Manag.* **2006**, *78*, 194–199. [\[CrossRef\]](#)
8. Cao, Y.; Larsen, D.P.; Hughes, R.M.; Angermeier, P.L.; Patton, T.M. Sampling effort affects multivariate comparisons of stream assemblages. *J. N. Am. Benthol. Soc.* **2002**, *21*, 701–714. [\[CrossRef\]](#)
9. Marignani, M.; Del Vico, E.; Maccherini, S. Spatial scale and sampling size affect the concordance between remotely sensed information and plant community discrimination in restoration monitoring. *Biodivers. Conserv.* **2007**, *16*, 3851–3861. [\[CrossRef\]](#)
10. Schmera, D.; Erős, T. The role of sampling effort, taxonomical resolution and abundance weight in multivariate comparison of stream dwelling caddisfly assemblages collected from riffle and pool habitats. *Ecol. Ind.* **2011**, *11*, 230–239. [\[CrossRef\]](#)

11. Anderson, M.J.; Santana-Garcon, J. Measures of precision for dissimilarity-based multivariate analysis of ecological communities. *Ecol. Lett.* **2015**, *18*, 66–73. [\[CrossRef\]](#)
12. Tordoni, E.; Napolitano, R.; Maccherini, S.; Da Re, D.; Bacaro, G. Ecological drivers of plant diversity patterns in remnants coastal sand dune ecosystems along the northern Adriatic coastline. *Ecol. Res.* **2018**, *33*, 1157–1168. [\[CrossRef\]](#)
13. Maun, M.A. *The Biology of Coastal Sand Dunes*; Oxford University Press: New York, NY, USA, 2009.
14. Dolan, A.H.; Walker, I.J. Understanding vulnerability of coastal communities to climate change related risks. *J. Coast. Res.* **2006**, *39*, 1316–1323.
15. Coombes, E.G.; Jones, A.P.; Sutherland, W.J. The biodiversity implications of changes in coastal tourism due to climate change. *Environ. Conserv.* **2008**, *35*, 319–330. [\[CrossRef\]](#)
16. Miller, T.E.; Gornish, E.S.; Buckley, H. Weather and coastal vegetation: Effects of storms and drought. *Plant Ecol.* **2010**, *206*, 97–104. [\[CrossRef\]](#)
17. Ciccarelli, D. Mediterranean coastal dune vegetation: Are disturbance and stress the key selective forces that drive the psammophilous succession. *Estuar. Coast. Shelf Sci.* **2015**, *165*, 247–253. [\[CrossRef\]](#)
18. Ciccarelli, D.; Pinna, M.S.; Alquini, F.; Cogoni, D.; Ruocco, M.; Bacchetta, G.; Sarti, G.; Fenu, G. Development of a coastal dune vulnerability index for Mediterranean ecosystems: A useful tool for coastal managers? *Estuar. Coast. Shelf Sci.* **2017**, *187*, 84–95. [\[CrossRef\]](#)
19. Bertacchi, A. Dune habitats of the Migliarino—San Rossore—Massaciuccoli Regional Park (Tuscany—Italy). *J. Maps* **2017**, *13*, 322–331. [\[CrossRef\]](#)
20. Carboni, M.; Santoro, R.; Acosta, A.T.R. Are some communities of the coastal dune zonation more susceptible to alien plant invasion? *J. Plant Ecol.* **2010**, *3*, 139–147. [\[CrossRef\]](#)
21. Novoa, A.; González, L.; Moravcová, L.; Pyšek, P. Constraints to native plant species establishment in coastal dune communities invaded by *Carpobrotus edulis*: Implications for restoration. *Biol. Conserv.* **2013**, *164*, 1–9. [\[CrossRef\]](#)
22. Tordoni, E.; Petruzzellis, F.; Nardini, A.; Savi, T.; Bacaro, G. Make it simpler: Alien species decrease functional diversity of coastal plant communities. *J. Veg. Sci.* **2019**, *30*, 498–509. [\[CrossRef\]](#)
23. Carmignani, L.; Lazzarotto, A.; Brogi, A.; Conti, P.; Cornamusini, G.; Costantini, A.; Sandrelli, P. *Carta geologica della Toscana (1: 250 000). Regione Toscana, Direzione Generale delle Politiche Territoriali e Ambientali, Servizio Geologico; Litografia Artistica Cartografica: Firenze, Italy, 2004.*
24. Pesaresi, S.; Galdenzi, D.; Biondi, E.; Casavecchia, S. Bioclimate of Italy: Application of the worldwide bioclimatic classification system. *J. Maps* **2014**, *10*, 538–553. [\[CrossRef\]](#)
25. Medvecká, J.; Jarolímek, I.; Senko, D.; Svitok, M. Fifty years of plant invasion dynamics in Slovakia along a 2,500 m altitudinal gradient. *Biol. Invasions* **2014**, *16*, 1627. [\[CrossRef\]](#)
26. Sperandii, M.G.; Bazzichetto, M.; Acosta, A.T.R.; Barták, V.; Malavasi, M. Multiple drivers of plant diversity in coastal dunes: A Mediterranean experience. *Sci. Total Environ.* **2019**, *652*, 1435–1444. [\[CrossRef\]](#)
27. Davies, C.E.; Moss, D.; Hill, M.O. *EUNIS Habitat Classification Revised*; Report to European Environment Agency; European Topic Centre on Nature Protection and Biodiversity: Paris, France, 2004.
28. Biondi, E.; Blasi, C.; Burrascano, S.; Casavecchia, S.; Copiz, R.; Del Vico, E.; Galdenzi, D.; Gigante, D.; Lasen, C.; Spampinato, G.; et al. Manuale Italiano di interpretazione degli habitat della Direttiva 92/43/CEE (Italian Interpretation Manual of the 92/43/EEC Directive habitats). Available online: <http://vnr.unipg.it/habitat/index.jsp> (accessed on 15 September 2019).
29. Chiarucci, A.; Bacaro, G.; Rocchini, D.; Ricotta, C.; Palmer, M.; Scheiner, S. Spatially constrained rarefaction: Incorporating the autocorrelated structure of biological communities into sample-based rarefaction. *Commun. Ecol.* **2009**, *10*, 209–214. [\[CrossRef\]](#)
30. Bacaro, G.; Rocchini, D.; Ghisla, A.; Marcantonio, M.; Neteler, M.; Chiarucci, A. The spatial domain matters: Spatially constrained species rarefaction in a Free and Open Source environment. *Ecol. Complex.* **2012**, *12*, 63–69. [\[CrossRef\]](#)
31. Oksanen, J.; Blanchet, G.; Friendly, M.; Kindt, R.; Legendre, P.; McGlinn, D.; Minchin, P.R.; O'Hara, R.B.; Simpson, G.L.; Solymos, P.; et al. Vegan: Community Ecology Package. R package Version 2.5-6; Available online: <https://cran.r-project.org/web/packages/vegan/index.html> (accessed on 15 September 2019).
32. Bacaro, G.; Altobelli, A.; Cameletti, M.; Ciccarelli, D.; Martellos, S.; Palmer, M.W.; Ricotta, C.; Rocchini, D.; Scheiner, S.M.; Tordoni, E.; et al. Incorporating spatial autocorrelation in rarefaction methods: Implications for ecologists and conservation biologists. *Ecol. Indic.* **2016**, *69*, 233–238. [\[CrossRef\]](#)

33. Lande, R. Statistics and partitioning of species diversity, and similarity among multiple communities. *Oikos* **1996**, *76*, 5–13. [CrossRef]
34. Crist, T.O.; Veech, J.A.; Gering, J.C.; Summerville, K.S. Partitioning species diversity across landscapes and regions: A hierarchical analysis of α , β , and γ -diversity. *Am. Nat.* **2003**, *162*, 734–743. [CrossRef] [PubMed]
35. Anderson, M.J. A new method for non-parametric multivariate analysis of variance. *Austral Ecol.* **2001**, *26*, 32–46.
36. Anderson, M.J.; TerBraak, C.J.F. Permutation tests for multi-factorial analysis of variance. *J. Stat. Comput. Sim.* **2003**, *73*, 85–113. [CrossRef]
37. Clarke, K.R.; Gorley, R.N. *PRIMER v6: Users Manual/Tutorial*; PRIMER-E: Plymouth, UK, 2006.
38. Anderson, M.J.; Gorley, R.N.; Clarke, K.R. *PERMANOVA+ for PRIMER: Guide to Software and Statistical Methods*; PRIMER-E: Plymouth, UK, 2008; p. 214.
39. Clarke, K.R. Non-parametric multivariate analyses of changes in community structure. *Aust. J. Ecol.* **1993**, *18*, 117–143. [CrossRef]
40. Muggeo, V.M.R. Estimating regression models with unknown break-points. *Stat. Med.* **2003**, *22*, 3055–3071. [CrossRef] [PubMed]
41. Muggeo, V.M.R. Segmented: An R Package to Fit Regression Models with Broken-Line Relationships. *R. News* **2008**, *8/1*, 20–25.
42. R Core Team. R: A Language and Environment for Statistical Computing. R Foundation for Statistical Computing: Vienna, Austria, 2020. Available online: <https://www.R-project.org> (accessed on 15 September 2019).
43. Foster, B.L.; Tilman, D. Dynamic and static views of succession: Testing the descriptive power of the chronosequence approach. *Plant Ecol.* **2000**, *146*, 1–10. [CrossRef]
44. Acosta, A.; Carranza, M.L.; Izz, C.F. Are there habitats that contribute best to plant species diversity in coastal dunes? *Biodivers. Conserv.* **2009**, *18*, 1087–1098. [CrossRef]
45. Kuiters, A.; Kramer, K.V.; Van der Hagen, H.G.J.M.; Schaminée, J.H.J. Plant diversity, species turnover and shifts in functional traits in coastal dune vegetation: Results from permanent plots over a 52-year period. *J. Veg. Sci.* **2009**, *20*, 1053–1063. [CrossRef]
46. Ciccarelli, D.; Bacaro, G. Quantifying plant species diversity in coastal dunes: A piece of help from spatially constrained rarefaction. *Folia Geobot.* **2016**, *51*, 129–141. [CrossRef]
47. Carranza, M.L.; Acosta, A.T.R.; Stanisci, A.; Pirone, G.; Ciaschetti, G. Ecosystem classification for EU habitat distribution assessment in sandy coastal environments: An application in central Italy. *Environ. Monit. Assess.* **2008**, *140*, 99–107. [CrossRef]
48. Calvão, T.; Pessoa, M.F.; Lidon, F.C. Impact of human activities on coastal vegetation: A review. *Emir. J. Food Agric.* **2013**, *25*, 926.
49. Angiolini, C.; Landi, M.; Pieroni, G.; Frignani, F.; Finioia, M.G.; Gaggi, C. Soil chemical features as key predictors of plant community occurrence in a Mediterranean coastal ecosystem. *Estuar. Coast. Shelf Sci.* **2013**, *119*, 91–100. [CrossRef]
50. Angiolini, C.; Bonari, G.; Landi, M. Focal plant species and soil factors in Mediterranean coastal dunes: An undisclosed liaison? *Estuar. Coast. Shelf Sci.* **2018**, *211*, 248–258. [CrossRef]
51. Del Vecchio, S.; Fantinato, E.; Janssen, J.A.M.; Bioret, F.; Acosta, A.; Prisco, I.; Tzonev, R.; Marcenò, R.; Rodwell, J.; Buffa, G. Biogeographic variability of coastal perennial grasslands at the European scale. *Appl. Veg. Sci.* **2018**, *21*, 312–321. [CrossRef]
52. Maccherini, S.; Marignani, M.; Gioria, M.; Renzi, M.; Rocchini, D.; Santi, E.; Torri, D.; Tundo, J.; Honnay, O. Determinants of plant community composition of remnant biancane badlands: A hierarchical approach to quantify species-environment relationships. *Appl. Veg. Sci.* **2011**, *14*, 378–387. [CrossRef]
53. Torri, D.; Rossi, M.; Brogi, F.; Marignani, M.; Bacaro, G.; Santi, E.; Tordoni, E.; Amici, V.; Maccherini, S. Badlands and the dynamics of human history, land use, and vegetation through centuries—Chapter IV. In *Badlands Dynamics in a Context of Global Change*; Nadal-Romero, E., Martínez-Murillo, J.F., Khun, N., Eds.; Elsevier: Amsterdam, The Netherlands, 2018; pp. 111–153.
54. De l Vecchio, S.; Slaviero, A.; Fantinato, E.; Buffa, G. The use of plant community attributes to detect habitat quality in coastal environments. *AoB Plants* **2016**, *8*, plw040. [CrossRef]

55. Wiedemann, A.M.; Pickart, A.J. Temperate zone coastal dunes. In *Coastal Dunes: Ecology and Conservation. Ecological Studies: Analysis and Synthesis*; Martínez, M.L., Psuty, N.P., Eds.; Springer: Heidelberg, Germany, 2004; pp. 53–65.
56. Chelli, S.; Simonetti, E.; Campetella, G.; Chiarucci, A.; Cervellini, M.; Tardella, F.M.; Tomasella, M.; Canullo, R. Plant diversity changes in a nature reserve: A probabilistic sampling method for quantitative assessments. *Nat. Conserv.* **2019**, *34*, 145–161. [[CrossRef](#)]
57. Diekmann, M.; Kühne, A.; Isermann, M. Random vs non-random sampling: Effects on patterns of species abundance, species richness and vegetation-environment relationships. *Folia Geobot.* **2007**, *42*, 179. [[CrossRef](#)]
58. Lájer, K. Statistical tests as inappropriate tools for data analysis performed on non-random samples of plant communities. *Folia Geobot.* **2007**, *42*, 115–122. [[CrossRef](#)]
59. Swacha, G.; Botta-Dukát, Z.; Kački, Z.; Pruchniewicz, D.; Żołniercz, L.A. Performance comparison of sampling methods in the assessment of species composition patterns and environment–vegetation relationships in species-rich grasslands. *Acta Soc. Bot. Pol.* **2017**, *86*, 3561. [[CrossRef](#)]
60. Stohlgren, T.J.; Chong, G.W.; Kalkhan, M.A.; Schell, L.D. Rapid assessment of plant diversity patterns: A methodology for landscapes. *Environ. Monit. Assess.* **1997**, *48*, 25–43. [[CrossRef](#)]
61. Loos, J.; Hanspach, J.; von Wehrden, H.; Beldean, M.; Moga, C.I.; Fischer, J. Developing robust field survey protocols in landscape ecology: A case study on birds, plants and butterflies. *Biodivers. Conserv.* **2015**, *24*, 33–46. [[CrossRef](#)]
62. Bried, J.T.; Pellet, J. Optimal design of butterfly occupancy surveys and testing if occupancy converts to abundance for sparse populations. *J. Insect. Conserv.* **2012**, *16*, 489–499. [[CrossRef](#)]
63. Maccherini, S.; Bacaro, G.; Giovannetti, G.; Angiolini, C.; Chiarucci, A. Analysing methodological issues in short-term monitoring of rare European beech forests restoration. *Plant Biosyst.* **2019**, *153*, 60–67. [[CrossRef](#)]



© 2020 by the authors. Licensee MDPI, Basel, Switzerland. This article is an open access article distributed under the terms and conditions of the Creative Commons Attribution (CC BY) license (<http://creativecommons.org/licenses/by/4.0/>).