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Investigating the Effects of the Built Environment on PM_{2.5} and PM₁₀: A Case Study of Seoul Metropolitan City, South Korea

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Abstract: Air pollution has a major impact on human health and quality of life; therefore, its determinants should be studied to promote effective management and reduction. Here, we examined the influence of the built environment on air pollution by analyzing the relationship between the built environment and particulate matter (i.e., PM_{2.5} and PM₁₀). Air pollution data collected in Seoul in 2014 were spatially mapped using geographic information system tools, and PM_{2.5} and PM₁₀ concentrations were determined in individual neighborhoods using an interpolation method. PM_{2.5} and PM₁₀ failed to show spatial autocorrelation; therefore, we analyzed the associations between PM fractions and built environment characteristics using an ordinary least squares regression model. PM_{2.5} and PM₁₀ exhibited some differences in spatial distributions, suggesting that the built environment has different effects on these fractions. For instance, high PM₁₀ concentrations were associated with neighborhoods with more bus routes, bus stops, and river areas. Meanwhile, both PM_{2.5} and PM₁₀ were more likely to be high in areas with more commercial areas and multi-family housing, but low in areas with more main roads, more single-family housing, and high average gross commercial floor area. This study is expected to contribute to establishing policies and strategies to promote sustainability in Seoul, Korea.

Keywords: air pollution; built environment; geographic information system; particulate matter

1. Introduction

Korea, particularly its capital Seoul, has experienced substantial economic growth; however, such development has also had adverse effects on the environment, including a rise in air pollution [1]. Air pollution is recognized as a serious problem in metropolitan cities in Korea and worldwide. According to the World Health Organization (WHO), about 4.2 million people die annually from air pollution-related diseases. Moreover, approximately 91% of the world's population lives in areas with air pollutant levels that exceed the WHO standards [2]. Increased exposure to higher concentrations of air pollutants has a detrimental effect on health and can trigger respiratory, cardiovascular, and lung diseases [3,4]. Therefore, it is necessary to identify the influence of the built environment on air pollution to build safe neighborhoods that are protected from air pollution.

Dust can be classified into total suspended particles and particulate matter (PM) fractions based on particle size, such as PM₁₀ (<10 µm in diameter) and PM_{2.5} (<2.5 µm in diameter). Prolonged exposure to high levels of PM may increase the risk of diabetes and respiratory disease [5,6]. Meanwhile, exposure to PM_{2.5} is associated with heart and respiratory diseases [7], and the WHO has designated PM_{2.5} as a Group 1 carcinogen [8].

According to the Organisation for Economic Co-operation and Development (OECD) statistics, the average PM_{2.5} level in Korea was 32 µg/m³ in 2015, more than three times the WHO standard of

10 $\mu\text{g}/\text{m}^3$, representing the worst $\text{PM}_{2.5}$ level among the OECD countries [9]. Many studies within a wide array of fields, such as public health, environment, urban planning, transportation, and medicine, have investigated both PM_{10} and $\text{PM}_{2.5}$ with the aim of investigating air pollution problems [10–13]. However, in Korea, because $\text{PM}_{2.5}$ has only been measured and managed since 2013, there is relatively little research on $\text{PM}_{2.5}$ compared to PM_{10} [14,15].

From the studies on PM, the association of PM with the built environment is of interest. However, there is controversy regarding the relationship between land use and PM concentrations. For instance, some studies have demonstrated a positive relationship between $\text{PM}_{2.5}$ and residential area [11,16,17], while others have found lower PM concentrations in residential areas [18–20]. In particular, Chen et al. [21] and Rivera et al. [22] showed that PM_{10} was high in residential areas and especially high in high-density residential areas. Commercial areas have also been shown to have a positive relationship with PM levels [10,18–20]. For instance, in a study on air pollution around roads, the number of commercial facilities (e.g., restaurants) had a greater influence on $\text{PM}_{2.5}$ than the general presence of commercial land use [23]. Meanwhile, most studies have found high PM_{10} and $\text{PM}_{2.5}$ concentrations in industrial areas [12,13,16–18,20,21,23–25]. Finally, many studies have found lower PM_{10} and $\text{PM}_{2.5}$ levels in green areas [10,19,26,27]. However, McCarty and Kaza [28] found high PM_{10} levels in green areas and Weichenthal et al. [20] showed the same result for $\text{PM}_{2.5}$. Other studies have assessed the association of PM with other land use variables (e.g., river area [11,20,23,24] and urbanization area [25]). These findings showed that $\text{PM}_{2.5}$ and PM_{10} have a dependency on the typology of measurement site but not the metal content [29]. Built environment attributes related to transportation are also associated with PM concentration. Traffic volume on roads was found to have a significant effect on PM [16,23,27,30]. Regarding vehicle type, truck density and truck route length were found to be positively correlated with $\text{PM}_{2.5}$ [20,31,32]. Moreover, Rivera et al. [22] showed that $\text{PM}_{2.5}$ was higher in regions with more vehicles and motorcycles. In addition, road type has been found to affect PM. Liu et al. [24] found that $\text{PM}_{2.5}$ levels were positively associated with highway intensity in Shanghai, China. Several studies have revealed a positive association between road length and PM [11,12,16,17,21]. In a study of intersections in Toronto, Canada, Weichenthal et al. [10] revealed a positive relationship between intersections and $\text{PM}_{2.5}$. By contrast, Hankey and Marshall [16] found a negative correlation between intersections and $\text{PM}_{2.5}$ in Minneapolis, USA. In addition, road width [13,18], bus route length [10,16], and the number of bus stops [16] have been examined as determinants of PM.

Development density has also been studied as a determinant of PM [11,13,18,30]. Weichenthal et al. [18] found a positive relationship between average building height and $\text{PM}_{2.5}$. Meanwhile, Wolf et al. [11] showed that $\text{PM}_{2.5}$ levels were higher in areas with more buildings. In addition, Tang et al. [13] found high PM_{10} levels in areas with high building volumes. Oh and Chung [30] further divided the number of buildings into residential areas, commercial areas, and industrial areas to study the impact of development density on PM_{10} by land type. The number of buildings in residential and commercial areas had positive relationships with PM_{10} ; however, the number of buildings in industrial areas had a negative relationship. Meanwhile, Oh and Chung [30] examined the relationship between PM_{10} and gross building floor area of individual residential, commercial, and industrial areas. The gross residential and commercial floor areas were positively associated with PM_{10} , while the gross industrial floor area was negatively associated.

Although many studies have examined the relationship between PM and the built environment, the results show dependency on the spatiotemporal characteristics unique to each study; therefore, continuous research on the relationship between the built environment and PM is needed. Moreover, there is little research on $\text{PM}_{2.5}$ specific to Korea. Therefore, we investigated the association between PM and the built environment in neighborhoods in Seoul, Korea. We classified PM into $\text{PM}_{2.5}$ and PM_{10} to further investigate the differences in the effects of the built environment (e.g., land use, transportation, housing, and urban development density) on these two fractions of PM. This study

will help contribute to the development of eco-friendly environmental policies and interventions to support sustainability in Seoul.

2. Materials and Methods

2.1. Study Area

This study was conducted using air quality monitoring data from Seoul collected in 2014. Seoul is the capital city of the Republic of Korea and is the center of politics, economy, culture, and transportation in Korea. The total area of Seoul is 605.21 km². Seoul has a population of about 9.8 million, with 25 autonomous districts and 424 neighborhood-level administrative units. The total area of Seoul comprises 20% residential area, 30% green area, 13% residential and commercial mixed-use area, and 6% commercial and official area.

2.2. Data

2.2.1. Air Pollution

This study used air quality monitoring data from 2014 provided by the Korea Environment Corporation (KECO) [33]. These data included six types of air pollutants measured and collected at 40 monitoring stations, including 25 municipal air monitoring stations located in each of Seoul's autonomous districts and 15 main road air monitoring stations (Figure 1). This study focused on PM_{2.5} and PM₁₀ among the six air pollutants to examine the differences in the effects of the built environment on PM_{2.5} and PM₁₀. The 40 air pollution monitoring stations were geocoded and displayed on a geographic information system (GIS) map for considering the spatial attributes of the air pollutants. The concentrations of PM_{2.5} and PM₁₀ collected from the 40 monitoring stations were converted into a GIS-compatible database.

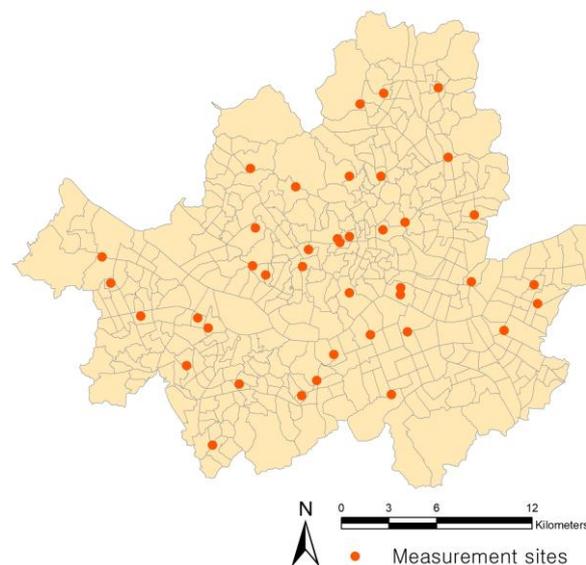


Figure 1. Air monitoring stations in Seoul.

2.2.2. Land Use

Land-use data from 2014 provided by the National Geographic Information Institute (NGII) were used [34]. In Korea, land use in urban areas is classified into four categories: residential, commercial, industrial, and green areas. Data on river areas were obtained from the Seoul Open Data Plaza [35].

2.2.3. Transportation

Transportation-related data were provided from the Korea Transportation Database (KTDB), the road name address guidance system, and the Korea Local Information Research & Development Institute (KLID) [36]. These transportation data were based on GIS maps, and include spatial coordinate information, bus routes, bus stops, subway station, road type, and road length.

2.2.4. Housing Type and Development Density

The characteristics of housing type were obtained from data on building use from the road name address guidance system, as well as the transportation data [36]. These data include the number of ground floors, underground floors, and spatial coordinate information for individual buildings, such as single-family housing, multi-family housing, sports facilities, business facilities, etc. In this study, we extracted housing data related to single-family and multi-family housing to identify the effects of neighborhood development density in residential areas on $PM_{2.5}$ and PM_{10} .

In addition, we examined the differences in the influence of urban development density in residential and commercial areas. Development density was obtained from the National Spatial Data Infrastructure Portal, which provides an OPEN API service [37]. These data include information on the building structure, land area, height, coverage ratio, and building area (e.g., gross floor area and floor area ratio). Using these data, we measured the average of gross floor area in residential and commercial areas by neighborhood.

2.3. Measurement of Variables

2.3.1. Spatial Unit of Analysis

We set the neighborhood-level administrative unit (*dong*, in Korean) as the spatial unit of analysis to identify the characteristics of the built environment related to $PM_{2.5}$ and PM_{10} at the neighborhood level. We analyzed the overall characteristics of air pollution in Seoul by measuring $PM_{2.5}$ and PM_{10} at the neighborhood level, rather than analyzing only the built environment around the 40 air pollution monitoring stations. The accuracy of analyses can be improved by using the smallest unit of administration as a spatial unit. Therefore, we expected our analysis to offer an accurate description of the current status of PM pollution in Seoul to help establish efficient air pollution reduction policies and strategies at the neighborhood level.

2.3.2. Dependent Variables

The $PM_{2.5}$ and PM_{10} concentrations at the neighborhood level were used as the dependent variables. $PM_{2.5}$ and PM_{10} concentrations, which were geocoded and built into the geodatabase, were determined using the inverse distance weighted (IDW) interpolation method for GIS. The size of the raster used in the IDW interpolation method was $25\text{ m} \times 25\text{ m}$, representing the smallest area used in previous studies [38]. Finally, the $PM_{2.5}$ and PM_{10} values from the IDW interpolation method were determined at the neighborhood level through the GIS zonal statistics tool.

Figures 2 and 3 show the processes used to determine the $PM_{2.5}$ and PM_{10} concentrations at the neighborhood level. $PM_{2.5}$ and PM_{10} at the neighborhood level in Seoul were in the range of $21\text{--}37\ \mu\text{g}/\text{m}^3$ and $43\text{--}62\ \mu\text{g}/\text{m}^3$, respectively. In Korea, the air-quality standard for $PM_{2.5}$ is an annual average of $15\ \mu\text{g}/\text{m}^3$; therefore, the $PM_{2.5}$ concentrations in all neighborhoods in Seoul exceeded the standard, indicative of potentially detrimental atmospheric conditions in Seoul. Meanwhile, the annual air-quality standard of PM_{10} is $50\ \mu\text{g}/\text{m}^3$; therefore, only some neighborhoods in Seoul had PM_{10} concentrations of concern. Moreover, the GIS maps in Figures 2 and 3 showed markedly different spatial distributions of $PM_{2.5}$ and PM_{10} . This suggests that the neighborhood-level built environment may have had different effects on $PM_{2.5}$ and PM_{10} ; therefore, it was necessary to examine the differences between the effects of the built environment on $PM_{2.5}$ and PM_{10} .

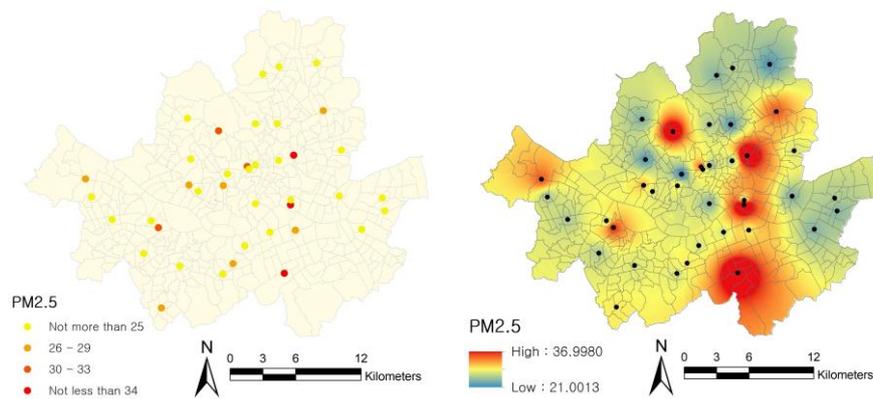


Figure 2. Spatial distribution of particulate matter ($PM_{2.5}$) in Seoul.

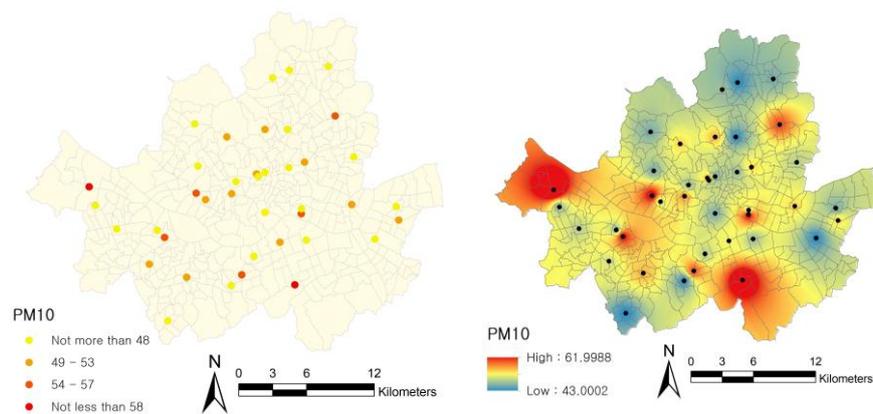


Figure 3. Spatial distribution of PM_{10} in Seoul.

2.3.3. Independent Variables

The independent variables were classified into four categories: land use, transportation, housing, and development density. Regarding the characteristics of land use, the proportions of commercial area, industrial area, green area, and river area were considered as features of the built environment. The land-use mix index, a measure differentiating between single or mixed use in a neighborhood, was also set as a variable where the index value ranged from 0 (single use) to 1 (perfectly mixed use) [39].

For the transportation attributes, the number of bus routes per hectare, number of bus stops per hectare, proportion of neighborhood streets, proportion of main roads, number of intersections per hectare, and presence of subway stations (dummy) were included in the analysis. In relevance to development density, this study considered the difference in residential density according to the type of housing and the distinction between residential density and commercial density. Development density can be represented as the factor for the degree of land use. High development density may significantly increase the consumption of energy and the operations of urban facilities. The number of single-family houses per hectare and the number of multi-family houses per hectare were included in the model to examine the degree of development density by housing type in residential areas. In addition, the average gross floor area of residential areas and the average gross floor area of commercial areas were measured to distinguish the differences in development density between residential and commercial areas. Finally, all variables were reported in terms of density, proportion, and average to minimize analysis errors caused by differences in neighborhood area.

3. Results and Discussion

3.1. Spatial Autocorrelation Analysis

Air pollutants naturally have spatial attributes, given the transport of dust in the atmosphere; therefore, it was necessary to assess the spatial autocorrelation of the PM fractions using Moran's I test. Spatial autocorrelation is defined as a measure of spatial similarity between nearby observations. Spatially, parameters that are nearby have a similar tendency as parameters that are far away. If spatial autocorrelation is found, it should be analyzed using a spatial econometrics model that can control for spatial autocorrelation. If spatial autocorrelation does not appear, it is appropriate to employ the ordinary least squares (OLS) regression model.

Table 1 shows the Moran's I test results for PM_{2.5} and PM₁₀ using ArcGIS. PM_{2.5} (Moran's I = -0.22 , $p = 0.18$) and PM₁₀ (Moran's I = -0.21 , $p = 0.22$) failed to show spatial autocorrelation, indicating that they were influenced more by the neighborhood-level built environment than by each other. Therefore, this study used the OLS regression model, as described previously [10,24,40].

Table 1. Moran's I test for spatial autocorrelation.

| Air Pollutants | Moran's I | <i>p</i> -Value |
|-------------------|-----------|-----------------|
| PM _{2.5} | -0.224 | 0.181 |
| PM ₁₀ | -0.210 | 0.220 |

3.2. Descriptive Statistics of Variables

Table 2 shows the descriptive statistics of variables considered in this study. The neighborhood mean of annual average concentrations of PM_{2.5} and PM₁₀ were 26 $\mu\text{g}/\text{m}^3$ and 48 $\mu\text{g}/\text{m}^3$, respectively. Among the built environment characteristics, land-use-related attributes (e.g., commercial, industrial, green, and river areas) showed large variations among neighborhoods. The mean land-use mix index was 0.21, indicative of relatively low levels of mixed land use in neighborhoods in Seoul. Regarding the transportation attributes, the average densities of bus routes and bus stops were 0.3 counts/ha and 0.2 counts/ha, respectively, and about 50% of the neighborhoods in Seoul had at least one subway station. Meanwhile, there was an average of 0.21 intersections/ha in the neighborhoods. In terms of roads, 57% were neighborhood streets and 25% were main roads. Residential areas had an average of 8.73 single-family houses/ha and 3.28 multi-family houses/ha. The average gross floor area in residential areas and commercial areas were 4714 m² and 2903 m², respectively. These results revealed a relatively low degree of urban development in commercial areas compared to residential areas, despite the presence of commercial land use.

Table 2. Descriptive statistics of variables.

| Variables | | Definition (Measures of Neighborhood) | Unit | Mean | Std. Deviation |
|---------------------|------------------------------|--|------------------------|---------|----------------|
| Particulate matter | PM _{2.5} | Annual average concentration of PM _{2.5} | µg/m ³ | 25.54 | 1.54 |
| | PM ₁₀ | Annual average concentration of PM ₁₀ | µg/m ³ | 48.44 | 2.16 |
| Land use | Commercial area | Proportion of commercial area | % | 5.25 | 13.18 |
| | Industrial area | Proportion of industrial area | % | 3.70 | 13.34 |
| | Green area | Proportion of green area | % | 19.99 | 24.82 |
| | Water area | Proportion of water area | % | 4.65 | 10.10 |
| | Mixed use | Land-use mix index | $0 \leq x \leq 1$ | 0.21 | 0.18 |
| Transportation | Bus route | Number of bus routes per hectare | counts/ha | 0.30 | 0.26 |
| | Bus stop | Number of bus stops per hectare | counts/ha | 0.23 | 0.13 |
| | Intersection | Number of intersections per hectare | counts/ha | 0.21 | 0.12 |
| | Neighborhood road | Proportion of neighborhood roads | % | 57.21 | 17.46 |
| | Main road | Proportion of main roads | % | 25.11 | 15.41 |
| | Subway station | Presence of subway station (0 = no presence, 1 = presence) | dummy | 0.50 | 0.50 |
| Housing type | Single-family housing | Number of single-family houses per hectare | counts/ha | 8.78 | 7.60 |
| | Multi-family housing | Number of multi-family houses per hectare | counts/ha | 3.28 | 2.55 |
| Development density | Gross commercial floor area | Average gross floor area of commercial areas | Avg. (m ²) | 2903.64 | 21,370.91 |
| | Gross residential floor area | Average gross floor area of residential areas | Avg. (m ²) | 4714.19 | 54,272.23 |

3.3. Regression Analysis

Table 3 shows the results of the regression models. The adjusted R^2 values of PM_{2.5} and PM₁₀ were 10.7% and 13.1%, respectively, indicating that the PM₁₀ model was slightly more explanatory than the PM_{2.5} model. However, the adjusted R^2 values of both models were relatively low so that the models did not seem to explain very much of this variability. This could suggest the limitation of the variables considered in this model that does not effectively take into account the direct pollutant sources of PM_{2.5} and PM₁₀. For more precise and effective pollution management and measurement, it also suggests that it is necessary to build more air monitoring stations in addition to the existing 40 monitoring stations. Overall, the PM_{2.5} and PM₁₀ results were similar, although some built environment characteristics had different effects on PM_{2.5} and PM₁₀. This suggests that it may be necessary to consider individual PM fractions to support more efficient targeted air pollution reductions, rather than general, comprehensive air pollution reduction plans.

Among the land-use characteristics, the proportion of commercial area had a significant effect in both the PM_{2.5} and PM₁₀ models. However, interestingly, both PM_{2.5} and PM₁₀ were likely to be lower in neighborhoods with a greater proportion of commercial area. These results are inconsistent with a previous study [18] but may represent the unique characteristics of land use in Seoul, Korea. In Seoul, even when land use is zoned as commercial land, the actual level of development of commercial land tends to be low. However, when considering urban development density characteristics indicative of the actual degree of development density, the average gross floor area of commercial areas showed a positive correlation with both PM_{2.5} and PM₁₀ concentrations. Therefore, it may be necessary to consider inconsistencies between the development status of land use and typical land use [10,30].

Table 3. Results of the regression models.

| | | PM _{2.5} | | PM ₁₀ | |
|---------------------|---|-------------------|-----------------|------------------|-----------------|
| Variable | | β | <i>t</i> -Value | β | <i>t</i> -Value |
| Constant | | 25.094 | 28.126*** | 46.992 | 38.192 *** |
| Land use | Commercial area | −0.025 | −3.055 *** | −0.033 | −2.936 *** |
| | Industrial area | 0.003 | 0.427 | 0 | −0.027 |
| | Green area | −0.007 | −1.212 | 0.009 | 1.204 |
| | Water area | 0.002 | 0.264 | 0.022 | 1.908 * |
| | Mixed use | 0.206 | 0.376 | 0.072 | 0.095 |
| Transportation | Bus route | −0.128 | −0.372 | 1.249 | 2.628 *** |
| | Bus stop | −0.184 | −0.277 | 1.563 | 1.705 * |
| | Intersection ^a | 0.014 | 0.104 | 0.244 | 1.349 |
| | Neighborhood road | 0.007 | 0.904 | 0.003 | 0.315 |
| | Main road | 0.025 | 3.104 *** | 0.032 | 2.890 *** |
| | Subway station (dummy) | −0.037 | −0.238 | 0.193 | 0.895 |
| Housing type | Single-family housing ^a | 0.18 | 2.875 *** | 0.162 | 1.876 * |
| | Multi-family housing ^a | −0.583 | −4.793 *** | −0.344 | −2.052 ** |
| Development density | Gross commercial floor area ^a | 0.089 | 3.868 *** | 0.12 | 3.769 *** |
| | Gross residential floor area ^a | −0.052 | −0.645 | −0.075 | −0.674 |
| F | | 4.395 *** | | 5.237 *** | |
| R-Square | | 0.139 | | 0.161 | |
| Adj. R-Square | | 0.107 | | 0.131 | |

* $p < 0.1$, ** $p < 0.05$, *** $p < 0.01$, ^a, log transformation.

The PM_{2.5} and PM₁₀ models failed to show a statistical significance with the industrial area, in contrast to the findings of previous literature [12,13,16–18,20,21,23–25]. This finding may be related to urban policies that led to the reduction in air pollution through factory relocations or changes that were made to factories in existing industrial areas located in the city center, such as by their replacement with eco-friendly commercial complexes and residential complexes. Green areas also failed to yield statistically significant associations with PM_{2.5} or PM₁₀ in agreement with literature [26–29]. Interestingly, higher PM₁₀ concentrations were associated with neighborhoods with more river areas, although the significance of this relationship was low. This counterintuitive result may be due to the fact that the main traffic corridors within Seoul are located along the north and south banks of the Han River. Finally, the land-use mix index was not significantly associated with PM_{2.5} or PM₁₀.

Of the transportation attributes, the proportion of neighborhood streets was not significantly associated with PM_{2.5} or PM₁₀. However, the proportion of main roads, which tend to have high traffic volumes, showed a strong positive relationship with both PM_{2.5} and PM₁₀, similar to previous studies [12,16,17,21]. The presence of subway stations in a neighborhood, which is considered to reflect the association between PM and subways as a form of public transportation that helps mitigate road traffic, showed no significant association with PM_{2.5} or PM₁₀. Finally, no significant relationship between PM_{2.5} or PM₁₀ and the number of intersections was found, which was considered to represent the influence of air pollution due to idling vehicles in the city center. This is in contrast to the positive correlation shown in previous studies [10,16]. Interestingly, bus-related variables showed different results for PM_{2.5} and PM₁₀. The densities of bus routes and bus stations were positively associated with higher levels of PM₁₀ but showed no association with PM_{2.5}. This is consistent with the literature [10,16] and suggests that pollutants emitted by buses may be associated more with PM₁₀ than PM_{2.5}. The distinction observed for bus-related factors in PM_{2.5} and PM₁₀ could suggest that

buses have a larger emission in the coarse fraction ($2.5 \mu\text{m} < \text{particles} < 10 \mu\text{m}$ in diameter). This is compatible with a larger contribution of non-exhaust emissions of buses compared to cars [41].

Regarding housing type and urban development density, higher $\text{PM}_{2.5}$ and PM_{10} concentrations tended to be associated with neighborhoods with higher average gross commercial floor area, similar to previous research [30]. However, $\text{PM}_{2.5}$ or PM_{10} was not associated with the average gross residential floor area, unlike previous studies [16,22] that showed higher $\text{PM}_{2.5}$ levels in high-density residential areas. Regardless, the more detailed built environmental characteristics of residential areas showed interesting findings. Contrary to the results of Rivera's study [22], lower $\text{PM}_{2.5}$ and PM_{10} levels were associated with neighborhoods with more multi-family houses, whereas both $\text{PM}_{2.5}$ and PM_{10} tended to be higher in areas with more single-family houses. This distinction between multi-family and single-family housing may be due to the Korean-specific housing characteristic where community complexes of high-rise multi-family housing, which is a typical type of multi-family housing in Korea, tend to be well-organized and include parks and green areas.

4. Conclusions

Air pollution has a major impact on human health and quality of life, necessitating the identification of the determinants of air pollutants to support air pollution reductions. The purpose of this study was to examine the effects of the built environment on urban air pollution in Seoul, Korea, with a focus on the distinction between the effects on $\text{PM}_{2.5}$ and PM_{10} as the most serious atmospheric pollutants in Korea.

The $\text{PM}_{2.5}$ and PM_{10} models showed similar associations with many of the investigated neighborhood-level built environment characteristics. For example, $\text{PM}_{2.5}$ and PM_{10} tended to be low in neighborhoods with greater proportions of commercial areas and multi-family housing. By contrast, neighborhoods with greater proportions of main roads, single-family housing, and higher average gross residential floor area were associated with higher $\text{PM}_{2.5}$ and PM_{10} concentrations. However, clear discrepancies between $\text{PM}_{2.5}$ and PM_{10} were also observed. For instance, bus-related features (e.g., bus stops and bus routes) and the proportion of river area were associated only with higher PM_{10} levels.

These findings suggest the need to establish targeted policies and strategies that acknowledge the differences between $\text{PM}_{2.5}$ and PM_{10} . In particular, for urban sustainability in Seoul, efficient planning and management of the built environment—such as land use, transportation, housing, and development density considered in this study—is needed beyond the management of direct air pollutant sources. Finally, the results of this study highlight the need for additional research to obtain more detailed data, especially time series data of air pollution and built environments, to inform policy interventions and provide specific policy recommendations.

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References

1. Kim, H.J.; Jun, M.J. Analysis on Relationship between Urban Development Characteristics and Air Pollution level. *J. Korea Plan. Assoc.* **2014**, *49*, 151–167. [CrossRef]
2. World Health Organization (WHO). Available online: <http://www.who.int/airpollution/ambient/en/> (accessed on 22 August 2018).
3. Bell, M.L.; Dominici, F. Effect Modification by Community Characteristics on the Short-Term Effects of Ozone Exposure and Mortality in 98 US Communities. *Am. J. Epidemiol.* **2008**, *167*, 986–997. [CrossRef] [PubMed]

4. Lin, S.; Liu, X.; Le, L.H.; Hwang, S.A. Chronic Exposure to Ambient Ozone and Asthma Hospital Admissions among Children. *Environ. Health Perspect.* **2008**, *116*, 1725–1730. [[CrossRef](#)] [[PubMed](#)]
5. Kramer, U.; Herder, C.; Sugiri, D.; Strassburger, K.; Schikowski, T.; Ranft, U.; Rathmann, W. Traffic-Related Air Pollution and Incident Type 2 Diabetes: Results from the SALIA Cohort Study. *Environ. Health Perspect.* **2010**, *118*, 1273–1279. [[CrossRef](#)] [[PubMed](#)]
6. Puett, R.C.; Hart, J.E.; Schwartz, J.; Hu, F.B.; Liese, A.D.; Laden, F. Are Particulate Matter Exposures Associated with Risk of Type 2 Diabetes? *Environ. Health Perspect.* **2011**, *119*, 384–389. [[CrossRef](#)] [[PubMed](#)]
7. Hong, Y.C.; Lee, J.T.; Kim, H.; Ha, E.H.; Schwartz, J.; Christiani, D.C. Effects of Air Pollutants on Acute Stroke Mortality. *Environ. Health Perspect.* **2002**, *110*, 187–191. [[CrossRef](#)] [[PubMed](#)]
8. International Agency for Research on Cancer (IARC). Available online: https://www.iarc.fr/en/media-centre/iarcnews/pdf/pr221_E.pdf (accessed on 23 August 2018).
9. Organisation for Economic Co-Operation and Development (OECD). Available online: <http://www.oecd.org/officialdocuments/publicdisplaydocumentpdf/?cote=ENV/EPOC/WPEI> (accessed on 19 August 2018).
10. Weichenthal, S.; Van Ryswyk, K.; Goldstein, A.; Shekarrizfard, M.; Hatzopoulou, M. Characterizing the Spatial Distribution of Ambient Ultrafine Particles in Toronto, Canada: A Land use Regression Model. *Environ. Pollut.* **2016**, *208*, 241–248. [[CrossRef](#)]
11. Wolf, K.; Cyrys, J.; Harciniková, T.; Gu, J.; Kusch, T.; Hampel, R.; Schneider, A.; Peters, A. Land use Regression Modeling of Ultrafine Particles, Ozone, Nitrogen Oxides and Markers of Particulate Matter Pollution in Augsburg, Germany. *Sci. Total Environ.* **2017**, *579*, 1531–1540. [[CrossRef](#)]
12. Ho, C.; Chan, C.; Cho, C.; Lin, H.; Lee, J.; Wu, C. Land use Regression Modeling with Vertical Distribution Measurements for Fine Particulate Matter and Elements in an Urban Area. *Atmos. Environ.* **2015**, *104*, 256–263. [[CrossRef](#)]
13. Tang, R.; Blangiardo, M.; Gulliver, J. Using Building Heights and Street Configuration to Enhance Intraurban PM₁₀, NO_x, and NO₂ Land use Regression Models. *Environ. Sci. Technol.* **2013**, *47*, 11643–11650. [[CrossRef](#)]
14. Kim, Y.K. An Analysis of Traffic Flows and Land-Use on Urban Air Pollution Concentrations Using Geographic Information System. *Korea Transp. Inst.* **2017**, *24*, 67–81. [[CrossRef](#)]
15. Oh, I.B.; Bang, J.H.; Kim, S.T.; Kim, E.H.; Hwang, M.K.; Kim, Y.H. Spatial Distribution of Air Pollution in the Ulsan Metropolitan Region. *J. Korea Soc. Atmos. Environ.* **2016**, *32*, 394–407. [[CrossRef](#)]
16. Hankey, S.; Marshall, J.D. Land use Regression Models of on-Road Particulate Air Pollution (Particle Number, Black Carbon, PM_{2.5}, Particle Size) using Mobile Monitoring. *Environ. Sci. Technol.* **2015**, *49*, 9194–9202. [[CrossRef](#)]
17. Sabaliauskas, K.; Jeong, C.; Yao, X.; Reali, C.; Sun, T.; Evans, G.J. Development of a Land-use Regression Model for Ultrafine Particles in Toronto, Canada. *Atmos. Environ.* **2015**, *110*, 84–92. [[CrossRef](#)]
18. Weichenthal, S.; Farrell, W.; Goldberg, M.; Joseph, L.; Hatzopoulou, M. Characterizing the Impact of Traffic and the Built Environment on Near-Road Ultrafine Particle and Black Carbon Concentrations. *Environ. Res.* **2014**, *132*, 305–310. [[CrossRef](#)] [[PubMed](#)]
19. Hong, J.S.; Kim, H.Y.; Lee, S.J. An Analysis of the Effects of Urban Characteristics on NO₂ Concentrations in Seoul. *Seoul Inst.* **2007**, *8*, 117–130.
20. Weichenthal, S.; Van Ryswyk, K.; Kulka, R.; Sun, L.; Wallace, L.; Joseph, L. In-Vehicle Exposures to Particulate Air Pollution in Canadian Metropolitan Areas: The Urban Transportation Exposure Study. *Environ. Sci. Technol.* **2015**, *49*, 597–605. [[CrossRef](#)]
21. Chen, L.; Bai, Z.; Kong, S.; Han, B.; You, Y.; Ding, X.; Du, S.; Liu, A. A Land use Regression for Predicting NO₂ and PM₁₀ Concentrations in Different Seasons in Tianjin Region, China. *J. Environ. Sci.* **2010**, *22*, 1364–1373. [[CrossRef](#)]
22. Rivera, M.; Basagaña, X.; Aguilera, I.; Agis, D.; Bouso, L.; Foraster, M.; Medina-Ramón, M.; Pey, J.; Künzli, N.; Hoek, G. Spatial Distribution of Ultrafine Particles in Urban Settings: A Land use Regression Model. *Atmos. Environ.* **2012**, *54*, 657–666. [[CrossRef](#)]
23. Farrell, W.; Weichenthal, S.; Goldberg, M.; Valois, M.; Shekarrizfard, M.; Hatzopoulou, M. Near Roadway Air Pollution Across a Spatially Extensive Road and Cycling Network. *Environ. Pollut.* **2016**, *212*, 498–507. [[CrossRef](#)]
24. Liu, C.; Henderson, B.H.; Wang, D.; Yang, X.; Peng, Z. A Land use Regression Application into Assessing Spatial Variation of Intra-Urban Fine Particulate Matter (PM_{2.5}) and Nitrogen Dioxide (NO₂) Concentrations in City of Shanghai, China. *Sci. Total Environ.* **2016**, *565*, 607–615. [[CrossRef](#)] [[PubMed](#)]

25. Rodríguez, M.C.; Dupont-Courtade, L.; Oueslati, W. Air Pollution and Urban Structure Linkages: Evidence from European Cities. *Renew. Sustain. Energy Rev.* **2016**, *53*, 1–9. [CrossRef]
26. Cho, H.; Choi, M.J. Effects of Compact Urban Development on Air Pollution: Empirical Evidence from Korea. *Sustainability* **2014**, *6*, 5968–5982. [CrossRef]
27. Hoek, G.; Beelen, R.; Kos, G.; Dijkema, M.; Zee, S.C.; Fischer, P.H.; Brunekreef, B. Land use Regression Model for Ultrafine Particles in Amsterdam. *Environ. Sci. Technol.* **2010**, *45*, 622–628. [CrossRef] [PubMed]
28. McCarty, J.; Kaza, N. Urban Form and Air Quality in the United States. *Landsc. Urban Plan.* **2015**, *139*, 168–179. [CrossRef]
29. Contini, D.; Cesari, D.; Donato, A.; Chirizzi, D.; Belosi, F. Characterization of PM10 and PM2.5 and their metals content in different typologies of sites in South-Eastern Italy. *Atmosphere* **2014**, *5*, 435–453. [CrossRef]
30. Oh, K.S.; Chung, H.B. The Influence of Urban Development Density on Air Pollution. *J. Korea Plan. Assoc.* **2007**, *42*, 197–210.
31. Clougherty, J.E.; Kheirbek, I.; Eisl, H.M.; Ross, Z.; Pezeshki, G.; Gorczynski, J.E.; Johnson, S.; Markowitz, S.; Kass, D.; Matte, T. Intra-Urban Spatial Variability in Wintertime Street-Level Concentrations of Multiple Combustion-Related Air Pollutants: The New York City Community Air Survey (NYCCAS). *J. Exp. Sci. Environ. Epidemiol.* **2013**, *23*, 232–240. [CrossRef]
32. Abernethy, R.C.; Allen, R.W.; McKendry, I.G.; Brauer, M. A Land use Regression Model for Ultrafine Particles in Vancouver, Canada. *Environ. Sci. Technol.* **2013**, *47*, 5217–5225. [CrossRef]
33. Air Korea. Available online: http://www.airkorea.or.kr/last_amb_hour_data (accessed on 27 July 2018).
34. National Geographic Information Institute. Available online: <http://www.ngii.go.kr/kor/main/main.do?rbsIdx=1> (accessed on 30 July 2018).
35. Seoul Open Data Plaza. Available online: <http://data.seoul.go.kr/search/newSearch.jsp> (accessed on 30 July 2018).
36. Korea Local Information Research & Development Institute. Available online: <http://www.juso.go.kr/addrlink/main.do> (accessed on 3 August 2018).
37. National Spatial Data Infrastructure Portal. Available online: <http://www.nsdi.go.kr/lxportal/?menu=2679> (accessed on 5 August 2018).
38. Wu, H.; Reis, S.; Lin, C.; Heal, M.R. Effect of Monitoring Network Design on Land use Regression Models for Estimating Residential NO2 Concentration. *Atmos. Environ.* **2017**, *149*, 24–33. [CrossRef]
39. Rajamani, J.; Bhat, C.; Handy, S.; Knaap, G.; Song, Y. Assessing Impact of Urban Form Measures on Nonwork Trip Mode Choice After Controlling for Demographic and Level-of-Service Effects. *Transp. Res. Rec. J. Transp. Res. Board* **2003**, *1831*, 158–165. [CrossRef]
40. Habermann, M.; Billger, M.; Haeger-Eugensson, M. Land use Regression as Method to Model Air Pollution. Previous Results for Gothenburg/Sweden. *Procedia Eng.* **2015**, *115*, 21–28. [CrossRef]
41. Nagpure, A.S.; Gurjar, B.R.; Kumar, V.; Kumar, P. Estimation of exhaust and non-exhaust gaseous, particulate matter and air toxics emissions from on-road vehicles in Delhi. *Atmos. Environ.* **2016**, *127*, 118–124. [CrossRef]



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