

Article

# Agent-Based Modelling of a Coupled Water Demand and Supply System at the Catchment Scale

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**Abstract:** Water is of uttermost importance for human well-being and a central resource in sustainable development. Many simulation models for sustainable water management, however, lack explanatory and predictive power because the two-way dynamic feedbacks between human and water systems are neglected. With Agent-based Modelling of Resources (Aqua.MORE; here, of the resource water), we present a platform that can support understanding, interpretation and scenario development of resource flows in coupled human–water systems at the catchment scale. Aqua.MORE simulates the water resources in a demand and supply system, whereby water fluxes and socioeconomic actors are represented by individual agents that mutually interact and cause complex feedback loops. First, we describe the key steps for developing an agent-based model (ABM) of water demand and supply, using the platform Aqua.MORE. Second, we illustrate the modelling process by application in an idealized Alpine valley, characterized by touristic and agricultural water demand sectors. Here, the implementation and analysis of scenarios highlights the possibilities of Aqua.MORE (1) to easily deploy case study-specific agents and characterize them, (2) to evaluate feedbacks between water demand and supply and (3) to compare the effects of different agent behavior or water use strategies. Thereby, we corroborate the potential of Aqua.MORE as a decision-support tool for sustainable watershed management.

**Keywords:** agent-based modelling; resources demand and supply; socio-hydrology; NetLogo

## 1. Introduction

In order to sustain crucial ecosystem services for current and future generations, resources need to be managed sustainably. The resource water is linked to a range of ecosystem services that are central to human well-being, including extractive water supply, in-stream water supply, and the provision of water-related cultural and supporting services [1]. Sustainable water management has been claimed to require integrated approaches addressing all natural and social dimensions of human–water systems [2,3]. For this reason, the recent scientific decade of the International Association of Hydrological Sciences (IAHS), ‘Panta Rhei–Everything Flows’, initiates research activities to investigate changes in both hydrology and society with interdisciplinary approaches [4]. Traditional hydrology treats human activities as an external force on water cycle dynamics, and therefore attempts to understand and model water-related ecosystem services and existing water usage conflicts often fail [5,6]. Contrastingly, socio-hydrology investigates the mutual interactions and feedback loops between human and water systems in an explicit way that can support explaining the past, understanding the present and showing sustainable future trajectories of human–water systems [3,5,6]. Modelling approaches that cover the

complex interaction between humans and water as well as spontaneous and unexpected behavior improve the decision-making process for sustainable watershed management and can thereby support the conservation of essential ecosystem services [6,7].

Numerous studies have confirmed that agent-based modelling (ABM) is well suited for the investigation of dynamic coupled natural and social systems [8–11]. ABMs are used for the simulation of the behavior and mutual interaction of autonomous actors ('agents') to assess the response of a considered system as a whole [12]. Fifteen years after the first social ABM simulation had been developed by Thomas Schelling in 1978 [13], Lansing and Kremer [14] first applied ABM to water resource modelling for the example of a rice irrigation system in Bali. Since then, a large number of ABM frameworks and models have been developed and used to yield insights into water use for irrigation [15–17] or for urban households [18–22]. Apart from modelling individual water use sectors, broader ABM applications support water resource management across entire watersheds. Akhbari and Grigg [12] created an ABM framework where the state (decision-making agent), the farmers (demand agents) and the environment (demand agent, i.e., for good water quantity and quality in the river) are the relevant agents. Khan et al. [23] presented an ABM framework in which administrative units are the relevant agents in a watershed, making water management decisions for agricultural production, hydropower generation and ecological management. Originally developed for the upper Danube catchment in Germany and Austria, the decision-support system DANUBIA comprises 17 model components, whereby the six socioeconomic models (for households, tourism, economy, water supply, farming and demography) follow an agent-based approach [24–28]. DANUBIA is intended for use on a very high administrative level, i.e., governmental institutions on the state, country or river basin level.

All these ABM frameworks and model applications are limited to (1) one water use sector and/or to (2) a predefined set of agents and/or are (3) focusing on the water distribution by institutions and higher administrative units. They are hence of limited use for detailed investigations of the various small-scale variabilities in water demand and supply characteristics, e.g., in topographically complex areas like the European Alps. As the Alpine rivers supply a large share of Europe's population with fresh water, water management in the headwater regions affects the well-being of millions of people in the Alps and their surrounding lowlands [29]. Relevant water demand sectors can vary from one Alpine valley to the next, depending on the presence of hydropower plants, the necessity of farmland irrigation especially in the Southern Alpine belt, and the intensity of (winter) tourism including technical snowmaking [30]. Even though the European Alps exhibit a positive water supply–demand balance on a long-term (annual) basis [31], some regions (i.e., in the Southern belt and the inner zone) are repeatedly affected by short- to mid-term scarcity problems [32]. These are often triggered by (unexpected) droughts, but can mostly be attributed to management conflicts between competing user interests for industry, drinking water supply, hydropower production, agriculture, ecology, etc. [33]. Being intensified by climate change and increased drought risks [34], competition between local stakeholders will probably become stronger in the future and result in more frequent local and small-scale water management conflicts [30,32,33].

Hence, this study presents the new ABM approach Agent-based Modelling of Resources (Aqua.MORE; here, of the resource water) for detailed analyses of the resource water in interaction with local socioeconomic actors in smaller catchments, rather than concentrating on national regulatory frameworks and authorities. Following the concept of socio-hydrology, Aqua.MORE enables considering the co-evolution of decentralized human agents and water flows and can thereby complement an eco-hydrological model to simulate the complex and multifaceted nature of coupled human–water systems. We address the following objectives: (1) We present the structure, functionalities and features of the platform Aqua.MORE and provide step-by-step instructions for setting it up as a supply and demand water management model. (2) We illustrate the application of Aqua.MORE by an example of an idealized Alpine valley that is characterized by touristic and agricultural water demand sectors. Rather than for making specific policy recommendations, this example is intended to illustrate involved processes and actions in an agent-based water supply and demand model. (3) We analyze

scenarios representing various decision and behavioral rules of the agents to corroborate the platform's potential as a decision-support tool.

## 2. Platform Description

### 2.1. Agent-Based Modelling (ABM)

ABMs are generally used for the simulation of the behavior and mutual interactions of autonomous actors (micro-interactions) to assess the response of a considered system as a whole (macro-outcomes); ABM is therefore a 'bottom-up' approach. The actors in an ABM are called agents and represent entities of real-world systems, either individuals or collective entities such as organizations or groups [12]. In Aqua.MORE, the higher-level entities of the agents are *waters*, *managers* and *users*. *Waters* represent discrete quantities of the resource, *users* consume the resource, and *managers* regulate and manage resource flows to *users*. By representing both the available resource and the socioeconomic actors as agents, we treat them equally in their general abilities. This specialty would allow also the resource to change its properties or behavior, e.g., water could become contaminated and disseminate pollution. Within the higher-level entities, one can create and characterize an unlimited number of various lower-level entities.

### 2.2. The Programming Software

Aqua.MORE is implemented in NetLogo [35], a powerful and well-established open-source programming framework. NetLogo provides a code section as well as a graphical user interface (GUI) to inspect and alter the running model during runtime [36]. Visualization comprises a graphical representation of the model realm and its variables. The model user can interact through buttons, input fields and sliders. NetLogo is bundled with a series of extension packages and toolkits, e.g., 'BehaviorSpace' [37], which supports the performing of experiments and scenarios, or 'BehaviorSearch' [38], which uses genetic algorithms and other heuristic techniques to search the parameter space.

### 2.3. The Resource Component

Ideally, Aqua.MORE complements an eco-hydrological model which represents the biophysical processes of energy and water fluxes in the catchment (i.e., soil-plant-atmosphere continuum). In the current version, the amount of resource input is generated externally, i.e., it is read in from one or more csv-formatted data files. In a future version, Aqua.MORE could also be directly coupled with either classic rainfall-runoff models or with eco-hydrological models representing the interplay between vegetation and water. Literature already provides examples for ABM applications being coupled with hydrological models, e.g., with the 'Soil & Water Assessment Tool' (SWAT) [23,39,40].

Within Aqua.MORE, the available water is represented by agents named '*waters*'. *Waters* are created every time step at one side of the model environment and are moving all the way through it to the other side (i.e., downstream), thereby forming a one-dimensional downstream flow. They have the obligatory state variable 'amount\_water', representing the amount of water available at this point, the value of which is initially imported from the data file and is adapted whenever water is extracted by *managers* or *users*. Additional state variables could be attributed to the *waters*, e.g., 'quality\_water'.

### 2.4. The Socioeconomic Component

The socioeconomic actors are represented as agents in the same way as the resource water. *Users* consume the resource, while the *managers* regulate and manage resource flows to *users*. *Managers* and *users* are stationary in the model environment and must have the state variables 'residualwater', 'demand', 'scarcity' and 'excess'. The 'demand' represents the amount of water the *managers/users* wants to extract, but they eventually need to leave residual water in the resource flow ('residualwater') to satisfy the demands of downstream users and to fulfil ecological requirements. In contrast to the

*users*, the *managers* do not consume the water themselves, but redirect all or a part of it to the associated *users*. Within the *managers* and *users*, lower-level entities (i.e., types of water users) are created in form of various breeds according to the specific case study and additional individual state variables can be attributed to each breed.

### 2.5. Temporal and Spatial Resolution and Extent

The smallest time step for the computation, i.e., one iteration of the model code, is defined through discrete *ticks*. The model world in NetLogo is two dimensional and is divided up into a grid of *patches* [36]. Each *patch* is defined by coordinates and is a square piece of ‘ground’ in between the non-stationary *waters* can move. The stationary *managers* and *users* are fixed on distinct patches, whereby the respective location can be defined by the model user. By setting the order of the *users*, one can divide between upstream and downstream *users*. This schematic representation of the spatial dimension allows the investigation of upstream–downstream relationships, an important feature that is urgently called for in the literature [41,42].

### 2.6. Runtime Procedures

Aqua.MORE comprises eight runtime procedures, which are processed at each *tick* (Table 1). Most of the procedures are uniform for all applications, but with the possibility of slight adaptations for specific cases if necessary. The procedure ‘*to update-demands*’ includes the behavior and decision processes of the case-specific agents, and therefore needs to be individualized for each case study.

**Table 1.** Runtime Procedures of Agent-based Modelling of Resources (Aqua.MORE).

Procedure Name	Operation
<i>to create-waters</i>	one <i>water</i> is generated at the upper border of the model environment, its variable ‘amount_water’ is set according to the ‘inflow’ list (imported from data file)
<i>to move-waters</i>	all <i>waters</i> move one step forward
<i>to</i>	the variable ‘demand’ of the <i>managers</i> and the <i>users</i> is updated according to the specific submodel
<i>to managers-extract</i>	<i>managers</i> extract water (i.e., decrease the ‘amount_water’ of <i>waters</i> passing by according to their updated variable ‘demand’, under consideration of ‘residualwater’, resulting in values of ‘scarcity’ or ‘excess’) and send new <i>waters</i> with the respective ‘amount_water’ to associated <i>users</i>
<i>to users-extract</i>	<i>users</i> extract water (i.e., decrease the ‘amount_water’ of <i>waters</i> passing by, according to their updated variable ‘demand’, under consideration of ‘residualwater’, resulting in values of ‘scarcity’ or ‘excess’)
<i>to write-maxandmin</i>	maximum, mean and/or minimum values of ‘scarcity’ and ‘excess’ of <i>managers</i> and <i>users</i> are recorded
<i>to measure-runoff</i>	‘amount_water’ of the <i>waters</i> at the lower end of the model environment is recorded
<i>to kill-waters</i>	all <i>waters</i> that have reached the lower end of the model environment or that are exhausted (variable ‘amount_water’ = 0), are deleted

## 3. Modelling Methods

In conformity with the general steps in building an ABM [12,43,44], the key steps for setting up and applying Aqua.MORE for a coupled water supply and demand analysis are:

- Objective definition and stakeholder identification;
- Setting-up of temporal resolution and spatial extent;
- Agent identification and representation in the model;
- Specification of the agent behavior and interactions;
- Initialization and simulation;
- Verification, calibration, sensitivity analysis and validation.

### 3.1. Objective Definition and Stakeholder Identification

First, the model's objective, i.e., the research question, needs to be defined. Moreover, to determine the required level of detail for the model, the persons interested in the outcome of the simulation (stakeholders) should be identified.

### 3.2. Setting-Up of Temporal Resolution and Spatial Extent

The length of one *tick* is preferably defined such that the temporal dynamics and variability of the processes under investigation can be appropriately sampled. In the same way, the spatial extent of the model environment, i.e., the number and the size of patches, needs to be adapted by the model user according to the specific use case.

### 3.3. Agent Identification and Representation in the Model

As the resource of interest, water is represented by an agent type (*waters*) and can interact with socioeconomic actors (*managers* and *users*) in the catchment. The most important *managers* and *users* in the case study site need to be identified and are classified into lower-level entities (breeds) like 'inhabitants', 'hotels', 'snowmaking reservoirs' or 'hydropower users'. Several individuals can be represented by a single agent, as long as they are spatially close and are assumed to be comparable in their behavior; e.g., all inhabitants of one village. Care should be taken that the unifying of several individuals to one agent does not lead to the loss of micro-scale features that influences the macro-scale system behavior in an essential way.

### 3.4. Specification of the Agent Behavior and Interactions

The behavior and decision rules of the *users* and *managers* need to be designed, coded and included in the procedure 'to update-demands' (Table 1). We encourage the users of Aqua.MORE to design decision rules according to decision-making theories and/or empirical observations [8,9]. A recommendable way to estimate the behavior of agents in changing framing conditions is an intensive stakeholder and/or experts process [8]. Coding can be supported by model libraries or online repositories of published ABMs (e.g., [www.comses.net](http://www.comses.net)). Because NetLogo is widely used, various program codes for diverse applications can be downloaded and used as seed models for new agent implementations. This allows for easily deploying a wide variety of agent types, with traits and behavior grounded on manifold behavioral theories.

### 3.5. Initialization and Simulation

Before running a model, the initial situation should be defined. We recommend estimating historical or current water 'demand' of the *managers* and *users* by empirical and/or theoretical data, depending on their availability. It is often useful to establish several auxiliary state variables to calculate the final 'demand', e.g., 'demand' of farmers could be calculated from one variable representing the irrigated area in ha multiplied by another variable showing the specific irrigation demand of the vegetation per ha. Values for 'residualwater' are often available via official reports or publications (e.g., the 'Water Book' in Austria and Germany). Initial variables can either be defined directly in the code or can be imported via csv-formatted data files. For the simulation of scenarios, we suggest the use of the software tool 'BehaviorSpace' [37], which is integrated in NetLogo. 'BehaviorSpace' systematically varies the parameters, gives each run a specific run number and records the results of each model run. If the computer has multiple processor cores, then the model will be executed in parallel by default, one run per core.

### 3.6. Verification, Calibration, Sensitivity Analysis and Validation

Before application, the model is tested to find and fix errors in the implementation of the model in NetLogo (verification). After that, we recommend systematically testing the sensitivity of the

model to the high number of model parameters and stochastic elements [9], e.g., with the integrated software tool 'BehaviorSearch'. In a next step, the model can be calibrated and validated with these parameters, preferably also with 'BehaviorSearch'. The calibration and validation of ABMs to empirical data, however, brings numerous difficulties [9,10,45,46]. It is surely not sufficient to base calibration and validation only on empirical data of macro properties of the system. Even if one successfully reproduces an observed macro-scale phenomenon in the past, e.g., a particular water scarcity situation, it cannot be proven that the utilized set of behavioral rules at the micro-level is the correct one; as there are many of such sets that lead to the same macro-scale outcome. It is necessary to validate the model components and processes as well, especially when analyzing non-observable scenarios like in our idealized case study, e.g., the future impacts of management strategies or of the hypothetical activation of a hydropower plant. By their very nature, there are no real-world data available for future simulations. Therefore, the model also needs to show theoretical validity, agent behavioral validity, validity under extreme conditions, and structural validity [47].

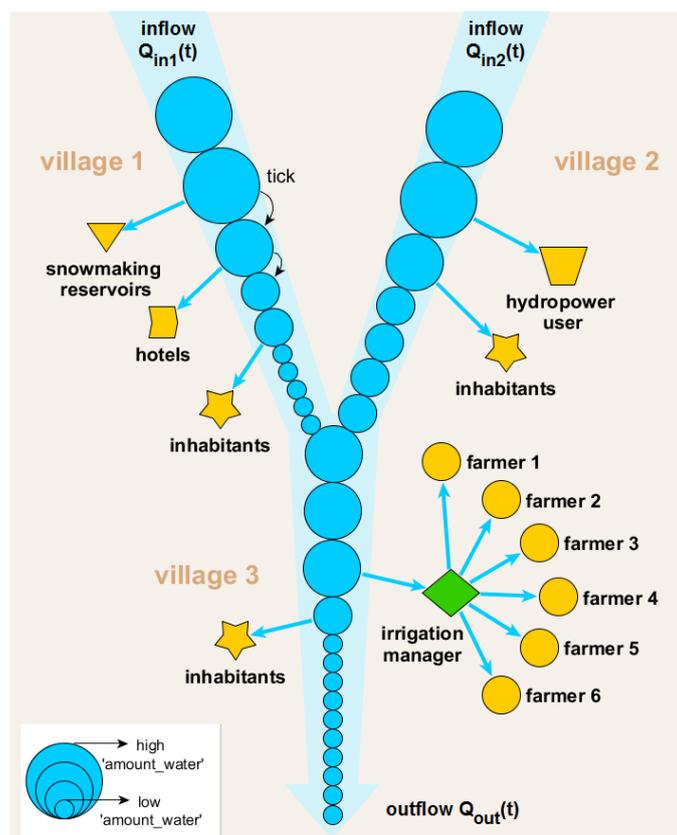
#### 4. Implementation Example

In this section, we present an idealized case study that illustrates the process of developing an application of the Aqua.MORE platform and the investigation of research questions by means of a scenario simulation. Although our example is inspired by existing water management as practiced in Alpine valleys in Tyrol (Austria) and South Tyrol (Italy), the simulation is neither based on real-world data, nor are the scenarios designed to make any specific management recommendation beyond our synthetic example. The idealized case study is rather intended for illustrating how coupled water supply and demand dynamics can be simulated and for describing how those processes are conceptualized in Aqua.MORE.

We assume the idealized case study site to be situated in a high mountain catchment, with the natural available water having a high seasonal variability, characterized inter alia by the melt of snow and glacier ice (nivo-glacial regime type). There are two mountain creeks passing one village each (village 1 and village 2), which then flow together to one river, passing village 3. Due to varying economic orientation of the villages, they differ in their main water demand sectors. Village 1 is very touristic, with hotels and technical snowmaking being the most important water users. Tourism has a generally increasing trend, but it is tied to a sufficient water supply. In village 2, tourism is negligible but, in future hydropower generation, it could become an important type of water use as the installation of a hydropower plant is envisaged. Village 3 is situated lower in the valley and is characterized by agriculture. During the vegetation period, an irrigation manager needs to supply six separate irrigation areas ('farmers') with irrigation water, with all farmers following the trend of an increasing irrigation demand. Moreover, households are also important water demand sectors in all three villages, but with static water demands. The aim of our exemplary, idealized model setup is the simulation of scenarios and the interpretation of their outputs (1) to assess the potential effects of a hydropower plant in village 2 and (2) to compare two different behavioral strategies of the irrigation manager in village 3.

##### 4.1. The Model Setup

Figure 1 shows the design of Aqua.MORE for the idealized case study site, including the realized agents and their order in the model environment. The NetLogo model environment consists of  $15 \times 17$  patches. One *tick* represents one hour.



**Figure 1.** Implementation of Aqua.MORE for the idealized case study site, with *waters* (blue circles), *managers* (green shapes) and *users* (yellow shapes). The varying shapes of *users* represent different ‘breeds’; the size of the *waters* represents the magnitude of their ‘amount\_water’.

1. The resource component. We intentionally use the same input for every simulation year—without any interannual variability or trends, e.g., due to climate change—to set the major focus on the various *users* and *managers*, with all their behavior and interactions.
2. The socioeconomic component. Behavior rules and interactions of the *managers* and *users* in the villages 1 to 3 (Figure 1) are the following:
3. Village 1. During years of sufficient water supply, tourism and therefore also the ‘demand’ of the *hotels* increases every year. Conversely, if there has been any severe scarcity situation for *hotels* or *snowmaking-reservoirs* in the previous year (i.e., the available water is satisfying less than 90% of their ‘demand’), the tourism is affected, therefore, leading to the decreasing ‘demand’ of *hotels*. Also, the ‘demand’ of the *snowmaking reservoirs* is directly coupled to the ‘demand’ changes of the *hotels*, but only being adapted every 5 years. Although these tourism trends are the result of an external (not simulated) process of tourists deciding their holiday destination, they can be coded with if/then rules as if they were the decisions of the *hotels* and *snowmaking reservoirs*.
4. Village 2. A ‘starting year’ for the water consumption of the *hydropower user* can be defined, allowing simulations either without *hydropower user* at all (‘demand’ = 0) or a start during the simulation (before start: ‘demand’ = 0, after start: ‘demand’ as imported from data file). The *hydropower user* does not make decisions and does not change ‘demand’ during the simulation run.
5. Village 3. The ‘demand’ of the *farmers* is calculated from auxiliary variables defining the irrigated area, the specific water demand per unit area and the traded water between the *farmers*. The *farmers* increase their irrigated area every 5 years. Moreover, the *farmers* trade water at the beginning of every year, i.e., every *farmer* with water scarcity in the previous year can ask every *farmer*

with excess in the previous year for water units. The decision for or against the trade is made randomly in our example, with a fifty percent chance. In the case of agreement between two *farmers*, the ‘demands’ of both are adapted for the following year as the ‘demands’ only represent the amount of water they need to obtain from *the irrigation manager*, i.e., a *farmer* who receives water units from another *farmer* decreases his/her ‘demand’ and vice versa. The *irrigation manager* can adapt the amount of irrigation water (i.e., his or her ‘demand’) every 10 years. If any of the *farmers* had experienced a severe scarcity situation (i.e., the available water is satisfying less than 90% of their ‘demand’) within the last 10 years—and if the *inhabitants* did not experience any scarcity—the *irrigation manager* increases the irrigation amount. Conversely, in the case of scarcity of *inhabitants* within the last 10 years, he or she decreases the irrigation amount.

Detailed model description following the updated ‘Overview, Design concepts and Details’ protocol (ODD) for describing individual-based and agent-based models [48] is included in the Supplementary Materials. The ODD protocol also includes lists of the model variables and their initialization values.

#### 4.2. Scenario Simulation and Output

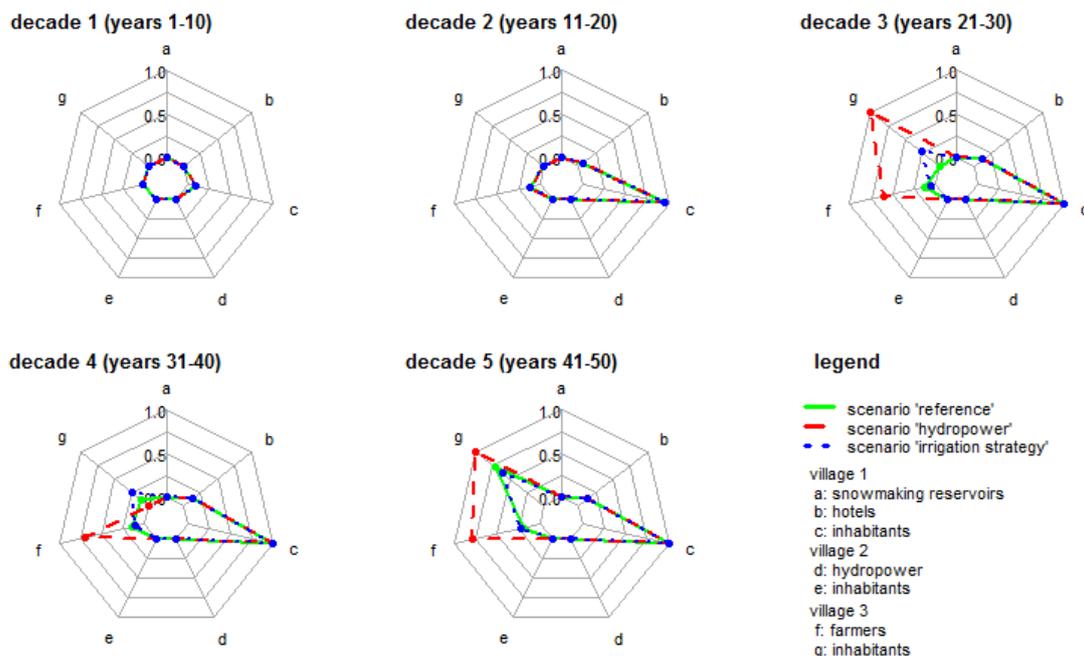
Three scenarios were implemented: The scenario ‘reference’ lacks a *hydropower user* in village 2, but includes a reactive strategy of the *irrigation manager* in village 3, meaning that if the criteria for an increase in irrigation water are fulfilled, the *irrigation manager* increases it by the maximum absolute amount of scarcity a farmer has experienced in the previous 10 years. The scenario ‘hydropower’ gives insights into the effects of taking in operation a hydropower plant in village 2. For this, the consumption of the *hydropower user* is started in simulation year 21 (beginning of the 3rd decade). Again, the *irrigation manager* follows a reactive strategy. The scenario ‘irrigation strategy’ uses a proactive strategy of the *irrigation manager*, i.e., he or she increases the irrigation water for the doubled amount of the maximum absolute value of scarcity a farmer has experienced in the previous 10 years, as he acts with foresight. In this scenario, no *hydropower user* is activated in village 2 (Table 2).

**Table 2.** Scenario Settings.

Scenario	Hydropower User in Village 2	Strategy of the Irrigation Manager
‘reference’	no	reactive (increase for single amount)
‘hydropower’	yes, starting in year 21 (beginning of the 3rd decade)	reactive (increase for single amount)
‘irrigation strategy’	no	proactive (increase for doubled amount)

All simulations were run for 50 years (five decades) to investigate the long-term effects of changing water user behavior (Figure 2). We conducted 100 model runs for each scenario to vary the random model components. By including the command ‘random-seed behaviorspace-run-number’ in the model setup, we ensured that all randomly created numbers within the model are based on an initial seed, which is again based on the run number of ‘BehaviorSpace’. This helps make the results reproducible and the scenarios comparable.

In the following, we will focus on the analysis of the mean results (the mean over all 100 runs per scenario) and on 10 year maxima of relative scarcities, i.e., the ratio between ‘scarcity’ and ‘demand’. Model outputs, however, would also allow (1) analyses with a higher temporal resolution up to 1 h, (2) analyses of several other monitoring variables and/or (3) a focus on extreme model runs representing worst cases. The data analyses and visualizations were performed using the R statistical programming language, version 3.6.1 [49].



**Figure 2.** Mean scenario results for the maximum of relative scarcity (relative scarcity = scarcity/demand) that occurred for the *managers* and *users* a–g within every decade of the 50 years model simulation. The value 0.0 indicates that the *user* has not experienced any scarcity within the decade; the value 0.5 indicates that the *user* could only satisfy half of his ‘demand’ in at least one time step within the decade; the value 1.0 indicates maximal scarcity in at least one time step within the decade.

Independent of the scenario, the occurrence of scarcity situations is limited to the *hotels* and *inhabitants* in village 1 and the *farmers* and *inhabitants* in village 3 (Figure 2)—the other three *users* are always sufficiently provided with water.

The *hotels* and *inhabitants* in village 1 are affected by scarcity to the same extent in all scenarios because the scenarios do not differ in tourism development. The mean relative scarcity of *hotels* is approximately 0.05 in decade 2 and approximately 0.12 in decades 3, 4 and 5. It hardly exceeds the threshold of 0.10, as tourism reacts every year, if the relative scarcity of *hotels* is higher than 0.10, tourism and therefore the ‘demand’ of the *hotels* decreases in the next year. But as tourism decreases only slightly and recovers quickly (as soon as the water supply of *hotels* recovers, tourism continues to increase immediately), the competing water user *inhabitants* cannot profit. Hence, the scarcity problem of the *inhabitants* in village 1 cannot be relieved by the self-regulation of tourism; but rather political interventions would be necessary.

The consequences of the scenario ‘hydropower’, in which the *hydropower* user in village 2 starts water abstraction in decade 3, do not manifest in village 2 itself. The water abstraction of the *hydropower* user rather affects the *farmers* and *inhabitants* of village 3, which show a significantly higher mean relative scarcity compared to the scenario ‘reference’. As the *inhabitants* in village 3 suffer from high relative scarcity in decade 3, the *irrigation manager* is politically forced to decrease the amount of irrigation water for the next decade. This leads to sufficient water supply of the *inhabitants* in decade 4, but at the same time to an exacerbated water scarcity situation of the *farmers*. As the *inhabitants* are not subject to any scarcity in decade 4, the *irrigation manager* is allowed to increase the amount of irrigation water again in decade 5 by diverting more water units to the *farmers*. He/she thereby improves the water supply of the *farmers*, but at the same time triggers again a severe scarcity of the *inhabitants*. These results show a typical upstream–downstream conflict, which cannot be solved by adaptations of water use behavior in village 3 alone. The *irrigation managers’* scope for action is limited in this case, as the total available water in village 3 is insufficient to satisfy the competing demands of *farmers* and *inhabitants*. A decision maker probably needs to address the problem at its source, namely in village 2.

When the criteria for an increase in irrigation water are fulfilled (i.e., severe water scarcity for the farmers and no water scarcity for the inhabitants within the last 10 years), the *irrigation manager* increases irrigation water for a higher amount in scenario ‘irrigation strategy’ (proactive strategy) than in scenario ‘reference’ (reactive strategy). Therefore, in decades 3 and 4, the proactive strategy leads to a significantly better water supply of the *farmers*, but also to significantly worse supply of the *inhabitants* in comparison to the reactive strategy. Surprisingly, in the fifth decade, the situation reverses: on average, the proactive strategy leads to a worse water supply of the *farmers* and to a better water supply of the *inhabitants*. This phenomenon can be explained as follows: in the scenario ‘irrigation strategy’, the *irrigation manager* is allowed to increase irrigation water in 60 of 100 model runs in decade 5 (condition: scarcity of farmers, but no scarcity of *inhabitants* in decade 4), thereby leading to a mean relative scarcity of 0.602 over all 100 model runs. In scenario ‘reference’, however, less model runs lead to scarcity of *inhabitants* in decade 4. Therefore, the *irrigation manager* is allowed to increase irrigation water in even 72 of 100 model runs in decade 5, thereby leading to a higher mean relative scarcity of 0.722 over all model runs. Thus, the scenario ‘reference’ could not prevent the scarcity of *inhabitant*’ in the long-term, but only delayed the scarcity situations to decade 5. These detailed insights on the consequences of the two irrigation strategies for the competing water demand sectors and for several periods can help a decision maker to make decisions for an irrigation strategy.

In addition to the three presented scenarios, further investigations could involve: (1) the effects of contrasting tourism developments and/or other (e.g., longer lasting) reactions of tourists to scarcity situations, (2) the effectiveness of additional irrigation strategies of the *irrigation manager* or (3) the effects of a dynamically adapted water abstraction of the *hydropower user*.

## 5. Discussion

### 5.1. Aqua.MORE as a Decision-Support Tool for Sustainable Water Management

Sivapalan et al. [5,6] have shown with several examples that the investigation of the fully coupled human–water system (i.e., the new concept of socio-hydrology) produces more explanatory and predictive power compared to conventional approaches, especially for long-term predictions. Therefore, a new kind of socio-hydrologic model and decision-support tool is required to formulate and test alternative hypotheses for sustainable watershed management [7]. With Agent-based Modelling of Resources (Aqua.MORE; here, of the resource water), we present such a platform to simulate the water resources in a coupled demand and supply system. Aqua.MORE is designed for covering the full coupled natural and social system, where people and nature mutually interact and cause feedback loops [50], i.e., higher water demand leads to lower water availability for competing users and, at the same time, the water availability also affects water demand because agents respond to water availability (i.e., to water scarcity) by modifying their water use behavior.

The overall aim of the study was to provide the key steps and a practicable workflow to implement a case study in the platform Aqua.MORE. The implementation of an ABM typically consists of two phases [51]. Our application of Aqua.MORE for the idealized case study is an example for the first phase, where model evaluations are qualitative and fitting to data is not a major issue. In this development stage, the model can already serve as a heuristic tool to explore ideas and gain system understanding, thereby matching one of the major application fields of an ABM [10,45]. We intentionally kept the agent behavior and interaction rules in the case study very simple to facilitate the interpretation of the model results, on the one hand, and to show that complex and unexpected macro-scale system behavior can emerge from the interactions of individual agents operating with such simple rules on the other hand. We have for example surprisingly found that a proactive strategy in comparison to a reactive strategy of the *irrigation manager* has negative impacts on the water supply of the inhabitants and positive impacts on the water supply of the farmers in years 21–40, but completely opposite effects in years 41–50. The second phase of the implementation would be parameterizing the model using

social surveys or decision-making theories and subsequently validating the model against empirical observations [51], which was obviously not realizable for the synthetic case study site.

Despite being an idealized example, our application scenarios powerfully demonstrate the necessity of including all upstream and downstream actors with their trends, behavior and interactions in the model. It is well documented that the competition for shared water resources has always been, and will continue to be in future, a frequent origin of international conflicts and even wars [52,53]. With our work, we contribute a new method to support the consideration of upstream–downstream conflicts at small scales for sustainable water resource management in a watershed [41].

### 5.2. Advantages and User Convenience of Aqua.MORE

In contrast to many socio-hydrological models in literature that are limited to a predefined set of agents and behavior, Aqua.MORE enables the model users to conceptualize and code agents case by case, thereby guaranteeing that no actor has to be neglected or underrepresented. We see it as a benefit of the Aqua.MORE workflow that the researchers need to implement the behavior and interactions of the *managers* and *users* ‘from scratch’, as this ensures that the researcher is forced to have a conceptual understanding of the agents being modelled rather than running with pre-existing, and perhaps inappropriate, agent behavior. Moreover, ecological water requirements are not realized as separate agents [12], but rather incorporated as variables of *users* and *managers*. In more detail, ‘demand’ and ‘residualwater’ directly and indirectly cover a wide range of water-related ecosystem services, such as extractive and in-stream water supply (e.g., household, agricultural and industrial use), water-related cultural services (e.g., tourism) or water-associated supporting services (e.g., habitat for aquatic organisms) [1].

The concept of Aqua.MORE promises to be adaptable for various kinds of interdisciplinary resource studies, addressing a wide range of researchers from different fields, e.g., hydrology, sociology, ecology or agricultural sciences. We would like to promote the application of the general concept of Aqua.MORE for many more resources. The concept could also be applied to, e.g., simulate competing demands for the resource ‘wood’ and the behavior and interactions of decentralized users of fuelwood and construction wood. Researchers benefit from Aqua.MORE being implemented in NetLogo code, as it is very easy to learn and implement applications in this user-friendly programming environment even for beginners with little programming background [54]. Moreover, many existing program codes for NetLogo are freely accessible (e.g., [www.comses.net](http://www.comses.net)) and can be used as seed models for agent implementations.

### 5.3. Embedding of Aqua.MORE

In contrast to classic rainfall–runoff models, eco-hydrological models enable considering the mutual feedbacks between vegetation and the hydrological cycle [5,6]. Eco-hydrology has therefore long been recognized as an appropriate way of assessing and sustainably managing freshwater resources [7,55]. However, eco-hydrological models alone do not allow for proper consideration of the mutual human feedbacks in the co-evolution of the human–water system. Aqua.MORE, on the other hand, is designed to incorporate the interactions between decentralized human agents and water flows. An eco-hydrological model and Aqua.MORE can perfectly complement each other; whereas the former generates the amount of water input for Aqua.MORE, the latter determines its partitioning and use. In future application, the two models could be directly coupled and exchange data in every time step of the simulation.

## 6. Conclusions

- Given the modular structure and integrated functionalities and features, Aqua.MORE is adaptable to various case study sites and water management problems, supporting detailed analyses of coupled human–water systems.

- The analyses of the presented scenarios show that modelling the behavior and mutual interactions of individual water-related actors can provide unexpected insights into human–water system dynamics. This illustrates the potential of Aqua.MORE as a decision-support tool for watershed management.
- Ideally, Aqua.MORE complements an eco-hydrological model to allow the overall assessment and management of the resource water including all biotic and mutual human interactions and feedback loops with the water component.
- By recognizing the relevance of the mutual interplay between the human and the natural components of water supply and demand, we make a contribution towards understanding and managing sustainability challenges.

**Supplementary Materials:** The following is available online at <http://www.mdpi.com/2071-1050/11/21/6178/s1>, ODD protocol for Aqua.MORE 1.0. The software can be downloaded from <https://www.comses.net/codebase-release/febcb4a9-1af7-447c-a4da-78ba6bbec86a/> (requires NetLogo 6.0.4 or higher).

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