



Article

Modeling the Compaction Characteristics of Fine-Grained Soils Blended with Tire-Derived Aggregates

Amin Soltani 1,* , Mahdieh Azimi 2 and Brendan C. O'Kelly 3

- School of Engineering, IT and Physical Sciences, Federation University, Churchill, VIC 3842, Australia
- School of Engineering and Technology, Central Queensland University, Melbourne, VIC 3000, Australia; m.azimiadehmortezapasha@cqu.edu.au
- Department of Civil, Structural and Environmental Engineering, Trinity College Dublin, D02 PN40 Dublin, Ireland; bokelly@tcd.ie
- * Correspondence: a.soltani@federation.edu.au

Abstract: This study aims at modeling the compaction characteristics of fine-grained soils blended with sand-sized (0.075-4.75 mm) recycled tire-derived aggregates (TDAs). Model development and calibration were performed using a large and diverse database of 100 soil-TDA compaction tests (with the TDA-to-soil dry mass ratio ≤ 30%) assembled from the literature. Following a comprehensive statistical analysis, it is demonstrated that the optimum moisture content (OMC) and maximum dry unit weight (MDUW) for soil-TDA blends (across different soil types, TDA particle sizes and compaction energy levels) can be expressed as universal power functions of the OMC and MDUW of the unamended soil, along with the soil to soil-TDA specific gravity ratio. Employing the Bland-Altman analysis, the 95% upper and lower (water content) agreement limits between the predicted and measured OMC values were, respectively, obtained as +1.09% and -1.23%, both of which can be considered negligible for practical applications. For the MDUW predictions, these limits were calculated as +0.67 and -0.71 kN/m³, which (like the OMC) can be deemed acceptable for prediction purposes. Having established the OMC and MDUW of the unamended fine-grained soil, the empirical models proposed in this study offer a practical procedure towards predicting the compaction characteristics of the soil-TDA blends without the hurdles of performing separate laboratory compaction tests, and thus can be employed in practice for preliminary design assessments and/or soil-TDA optimization studies.

Keywords: fine-grained soil; tire-derived aggregate; optimum moisture content; maximum dry unit weight; Bland–Altman analysis

check for updates

Citation: Soltani, A.; Azimi, M.; O'Kelly, B.C. Modeling the Compaction Characteristics of Fine-Grained Soils Blended with Tire-Derived Aggregates. Sustainability 2021, 13, 7737. https://doi.org/10.3390/su13147737

Academic Editor: Castorina Silva Vieira

Received: 9 June 2021 Accepted: 8 July 2021 Published: 11 July 2021

Publisher's Note: MDPI stays neutral with regard to jurisdictional claims in published maps and institutional affiliations.



Copyright: © 2021 by the authors. Licensee MDPI, Basel, Switzerland. This article is an open access article distributed under the terms and conditions of the Creative Commons Attribution (CC BY) license (https://creativecommons.org/licenses/by/4.0/).

1. Introduction

Lately, many developed and developing countries have initiated the transition to 'sustainable infrastructure', a concept that (among other things) encourages the replacement of natural quarry-based aggregates with recycled solid waste materials. End-of-life tires (ELTs) from the automotive industry are among the largest and most problematic global waste streams, prompting recycled tire-derived aggregates (TDAs) to become one of the most targeted materials for civil engineering applications. Because of their physical and mechanical attributes, particularly in terms of their relatively low density, high energy absorption capacity, resilience and low water adsorption–retention potential, granulated TDA-based products (e.g., crumbs, buffings and fibers) have been well established as effective soil-blending agents for the development of high-performance (and sustainable) geomaterials for a variety of practical geotechnical applications, such as soil stabilization, highway embankment and pavement constructions, as well as for bridge abutment and retaining wall backfills [1–5]. Further, Shahrokhi-Shahraki et al. [6] investigated the use of pulverized waste tire, either on its own or mixed with soil (well-graded sand), to act as an adsorptive fill material, demonstrating adsorption of organic/inorganic contaminants,

Sustainability **2021**, 13, 7737 2 of 21

namely benzene, toluene, ethylbenzene and xylene (BTEX) components, and two heavy metal ions (Pb^{2+} and Cu^{2+}).

Research on soil-TDA mixtures dates back to the early 1990s, where theoretical concepts governing the mechanical performance of this (then-emerging) geomaterial were first put into perspective. Earlier investigations were mainly focused on coarse-grained soils (mainly sands), demonstrating that the granular soil-TDA blend, resembling a rigid-soft matrix, can be optimized in terms of the TDA content and its particle geometry (i.e., its mean particle size and shape) to achieve any desired balance between the strength/stiffness and deformability parameters of the TDA-based blend [7–13]. These early investigations unanimously concluded that the addition of TDA to coarse-grained soils leads to notable reductions in the soils' mobilized strength and stiffness while enhancing their ductility characteristics, which was mainly ascribed to the lower stiffness (and higher deformability) of the soft TDA particles compared with that of the rigid soil grains. Furthermore, depending on the TDA content and its mean particle size (in relation to the rigid soil grains), the stress-strain response of a granular soil-TDA blend can fall into one of three behavioral categories [12,13]: (i) rigid-dominant; (ii) rigid-soft transitional; and (iii) soft-dominant. The transitional behavior (by definition) resembles a perfect balance between the blend's strength/stiffness and ductility/toughness—a review of the research literature indicates that the transitional behavior is often encountered at a volumetric TDA content (commonly defined as the 'TDA-to-granular soil + TDA' volume ratio) of 30–50% [14].

Later studies followed suit, confirming the suitability of TDA-based products, particularly when paired with chemical binders, as effective blending agents for compacted finegrained soils (including expansive clays) capable of promoting improved shear strength performance, reduced swell-shrink (and hence desiccation-induced cracking) potential and improved damping [15–22]. In terms of shear strength, for instance, these studies concluded that the addition of TDA at low TDA-to-fine-grained soil mass ratios (mainly less than 10%) often produces relatively small improvements, attributed to 'arching' between the TDA particles within the soil-TDA agglomerations [19,22] and induced 'inter-particle friction' generated at the soil-TDA interfaces [16]. These studies also demonstrate that higher TDA contents tend to cause serious concerns for undrained strength and stiffness, largely due to the relatively lower stiffness (and higher deformability) of the soil-TDA agglomerations compared to individual soil agglomerations containing no TDA [17,19]. Accordingly, for projects where the strength and stiffness are of primary importance, the compacted fine-grained soil-TDA blend requires stabilization by means of conventional cementitious (e.g., Portland cement and quick lime [15,18]) or polymer (e.g., polyacrylamide and sodium alginate [22,23]) binders.

Like natural (unamended) fine-grained soils, an essential step towards the production and placement of suitable soil-TDA earth fills is compaction. The governing variables which control the compactability of TDA-blended fine-grained soils have been well documented in the research literature. It is generally accepted that the addition of TDA to fine-grained soils leads to notable reductions in the optimum moisture content (OMC) and the corresponding maximum dry unit weight (MDUW), mainly attributed to the TDA material's hydrophobic character (water adsorption being mainly less than 4%), relatively lower density and higher energy absorption capacity compared with that of the soil solids [24–26]. The OMC and MDUW parameters are commonly measured by standardized laboratory compaction tests which, though straightforward in terms of execution, are fairly labor-intensive and highly time-consuming. Accordingly, several attempts have been made to devise empirical-type correlations for indirect estimation of the OMC and MDUW of unamended fine-grained soils, all of which employ the soil consistency (Atterberg) limits as the primary predictors [27-33]. Common TDA-based products (e.g., crumbs, buffings and fibers) used in conjunction with fine-grained soils are mainly similar in size to predominantly medium-coarse sand (0.425-4.75 mm); as such, the soil consistency limit tests would not be applicable to most soil-TDA blends. This implies that the many well-established empirical correlations reported for indirect estimation of the OMC and

Sustainability **2021**, 13, 7737 3 of 21

MDUW of unamended fine-grained soils cannot be extended to soil–TDA blends; this limitation highlighting the need to develop an entirely new modeling framework.

A review of the research literature indicates that no modeling framework exists for the compaction characteristics of fine-grained soil–TDA blends. Accordingly, this study aims at establishing practical empirical models for indirect estimation of the OMC and MDUW of TDA-blended fine-grained soils. Model development and calibration are carried out using a large and diverse database of 100 soil–TDA compaction tests assembled from the research literature. The empirical models proposed in this study offer a practical procedure towards predicting the compaction characteristics of soil–TDA blends without the hurdles of performing separate laboratory compaction tests, and thus can be used for preliminary design assessments and/or soil–TDA optimization studies.

2. Database of Soil-TDA Compaction Tests

Given that empirical models/correlations are purely data-driven, their predictive capability is highly dependent on the database from which they are developed. Accordingly, to establish practical empirical models for the OMC and MDUW of soil–TDA blends, a large and diverse database of 100 soil–TDA compaction tests was gathered from the authors' previous publications [19,22,23,34,35] as well as other recent literature sources [15–17,36–39]. Detailed descriptions of the assembled database are presented in Tables 1 and 2. The compiled database consisted of 21 datasets (designated as D1–D21), each defined as a collection of standard or modified (heavy) Proctor compaction tests for a given fine-grained soil mixed with a particular TDA material (constant particle size/shape) at varying TDA contents (denoted as $f_{\rm T}$ and defined as the TDA-to-soil dry mass ratio, here expressed as a percentage value).

CH

CH

CV

CI

CI

CI

| Dataset | Source | Compaction Energy Level | N | Soil ID | $G_{\mathrm{s}}^{\mathrm{S}}$ | LL (%) | PI (%) | $f_{\rm fines}$ (%) | $f_{ m clay}$ (%) | \boldsymbol{A} | Soil USCS |
|---------|--------|----------------------------|---|---------|-------------------------------|--------|--------|---------------------|-------------------|------------------|-----------|
| D1 | [15] | Modified Proctor | 4 | S1 | 2.61 | 49.5 | 26.2 | - | - | - | CI |
| D2 | [36] | Standard Proctor | 5 | S2 | 2.65 | 74.0 | 38.0 | 97.0 | 70.0 | 0.54 | MV |
| D3 | [36] | Standard Proctor | 5 | S3 | 2.65 | 56.0 | 36.0 | 94.0 | 59.0 | 0.61 | CH |
| D4 | [37] | Standard Proctor | 4 | S4 | 2.72 | 53.0 | 14.0 | 63.6 | 52.8 | 0.27 | MH |
| D5 | [16] | Standard Proctor | 7 | S5 | 2.69 | 52.2 | 28.1 | 96.4 | 47.2 | 0.60 | CH |
| D6 | [16] | Modified Proctor | 7 | S5 | 2.69 | 52.2 | 28.1 | 96.4 | 47.2 | 0.60 | CH |
| D7 | [38] | Modified Proctor | 5 | S6 | 2.69 | 34.2 | 9.4 | 92.2 | 60.7 | 0.15 | ML |
| D8 | [17] | Modified Proctor | 5 | S6 | 2.69 | 34.2 | 9.4 | 92.2 | 60.7 | 0.15 | ML |
| D9 | [39] | Standard Proctor | 5 | S7 | 2.65 | 52.0 | 31.0 | 62.0 | - | - | CH |
| D10 | [39] | Standard Proctor | 5 | S8 | 2.65 | 60.0 | 38.0 | 54.0 | - | - | CH |
| D11 | [34] | Standard Proctor | 5 | S9 | 2.69 | 44.2 | 21.9 | 99.0 | 49.0 | 0.45 | CI |
| D12 | [34] | Standard Proctor | 5 | S10 | 2.67 | 47.2 | 29.2 | 69.0 | 37.0 | 0.79 | CI |
| D13 | [34] | Standard Proctor | 5 | S11 | 2.71 | 59.5 | 31.6 | 99.0 | 53.0 | 0.60 | CH |
| D14 | [34] | Standard Proctor | 5 | S12 | 2.72 | 77.6 | 57.0 | 80.0 | 44.0 | 1.30 | CV |

Table 1. Soil properties for the compiled database of soil-TDA compaction tests.

Note: N = number of compaction tests; G_s^S = specific gravity of soil solids (in the absence of a reliable value, the typical representative specific gravity of 2.65 was considered); LL = liquid limit; PI = plasticity index; f_{fines} = fines content (<75 μ m); f_{clay} = clay content (<2 μ m); A = activity index (calculated as A = PI/ f_{clay}); and USCS = Unified Soil Classification System, as per BS 5930 [40].

2.73

2.73

2.76

2.77

2.77

2.77

2.73

59.6

59.6

78.4

43.6

43.6

43.6

32.3

32.3

54.2

21.5

21.5

21.5

99.1

99.1

78.0

80.0

80.0

80.0

51.7

51.7

43.0

43.0

43.0

43.0

0.62

0.62

1.26

0.50

0.50

0.50

S13

S13

S14

S15

S15

S15

5

5

4

4

D15

D16

D17

D18

D19

D20

D21

[35]

[35]

[23]

[19]

[19]

[19]

[22]

Standard Proctor

Sustainability **2021**, 13, 7737 4 of 21

| Table 2. TDA | properties for | the compiled da | tabase of soil-1DA | A compaction tests. | |
|--------------|----------------|-----------------|--------------------|---------------------|--|
| | | | | | |

| Dataset | Source | N | Soil ID | TDA Type | $G_{\mathrm{s}}^{\mathrm{T}}$ | D ₅₀ (mm) | C_{U} | C_{C} | TDA USCS | TDA Content f _T (%) |
|---------|--------|---|---------|------------------------------|-------------------------------|----------------------|------------------|------------------|----------|--------------------------------|
| D1 | [15] | 4 | S1 | Tire buffings | 1.08 | 2.36 | 2.39 | 0.94 | SP | 0, 5.3, 11.1, 17.6 |
| D2 | [36] | 5 | S2 | Crumb rubber (425–600 μm) | 0.85 | - | - | - | SP | 0, 5, 10, 15, 20 |
| D3 | [36] | 5 | S3 | Crumb rubber (425–600 μm) | 0.85 | - | - | - | SP | 0, 5, 10, 15, 20 |
| D4 | [37] | 4 | S4 | Granular rubber | 1.12 | 1.04 | 2.55 | 1.08 | SP | 0, 5, 10, 20 |
| D5 | [16] | 7 | S5 | Rubber powder (0.075–2 mm) | 1.14 | - | - | - | SP | 0, 2.5, 5, 10, 15, 20, 25 |
| D6 | [16] | 7 | S5 | Rubber powder (0.075–2 mm) | 1.14 | - | - | - | SP | 0, 2.5, 5, 10, 15, 20, 25 |
| D7 | [38] | 5 | S6 | Crumb rubber | 1.13 | 1.54 | 2.20 | 1.31 | SP | 0, 2.5, 5, 7.5, 10 |
| D8 | [17] | 5 | S6 | Rubber fiber | 1.07 | 1.64 | 2.95 | 1.27 | SP | 0, 2.5, 5, 7.5, 10 |
| D9 | [39] | 5 | S7 | Crumb rubber (75–425 μm) | 0.85 | - | - | - | SP | 0, 5, 10, 15, 20 |
| D10 | [39] | 5 | S8 | Crumb rubber (75–425 µm) | 0.85 | - | - | - | SP | 0, 5, 10, 15, 20 |
| D11 | [34] | 5 | S9 | Ground rubber | 1.09 | 0.48 | 2.83 | 1.19 | SP | 0, 5, 10, 20, 30 |
| D12 | [34] | 5 | S10 | Ground rubber | 1.09 | 0.48 | 2.83 | 1.19 | SP | 0, 5, 10, 20, 30 |
| D13 | [34] | 5 | S11 | Ground rubber | 1.09 | 0.48 | 2.83 | 1.19 | SP | 0, 5, 10, 20, 30 |
| D14 | [34] | 5 | S12 | Ground rubber | 1.09 | 0.48 | 2.83 | 1.19 | SP | 0, 5, 10, 20, 30 |
| D15 | [35] | 5 | S13 | Rubber crumbs | 1.09 | 0.48 | 2.83 | 1.19 | SP | 0, 5, 10, 20, 30 |
| D16 | [35] | 5 | S13 | Rubber buffings | 1.09 | 1.58 | 1.56 | 1.03 | SP | 0, 5, 10, 20, 30 |
| D17 | [23] | 5 | S14 | Ground rubber | 1.09 | 0.48 | 2.83 | 1.19 | SP | 0, 5, 10, 20, 30 |
| D18 | [19] | 4 | S15 | TDA-Fine | 1.08 | 0.46 | 3.06 | 1.23 | SP | 0, 5, 10, 20 |
| D19 | [19] | 4 | S15 | TDA-Medium | 1.10 | 1.67 | 1.85 | 1.09 | SP | 0, 5, 10, 20 |
| D20 | [19] | 4 | S15 | TDA-Coarse | 1.11 | 3.34 | 1.89 | 1.10 | SP | 0, 5, 10, 20 |
| D21 | [22] | 5 | S16 | Ground rubber | 1.09 | 0.46 | 3.06 | 1.23 | SP | 0, 5, 10, 20, 30 |

Note: N = number of compaction tests; G_s^T = specific gravity of TDA particles; D_{50} = mean TDA particle size; C_U and C_C = coefficients of uniformity and curvature, respectively; f_T = TDA-to-soil dry mass ratio, here expressed as a percentage value; and USCS = Unified Soil Classification System, as per BS 5930 [40].

As demonstrated in Table 1, the 21 datasets included a total of sixteen fine-grained soils (designated as S1–S16), covering reasonably wide ranges of surface texture, plasticity and mineralogical properties—that is, $f_{\text{clav}} = 37-70\%$, LL = 34.2 – 84.3% and $A = \text{PI}/f_{\text{clav}} = 0.15-1.30$ (where f_{clay} , LL, PI and A denote clay content, liquid limit, plasticity index and activity index, respectively). In terms of classification, the database soils consisted of three silts and thirteen clays with the following USCS frequencies, as per BS 5930 [40]: ML = 1; MH = 1; MV = 1; CI = 4; CH = 6; and CV = 3. Referring to Table 2; the compiled database covers all major types of commercially available poorly-graded sand-sized (i.e., SP classification) TDA products (e.g., powder, crumbs and buffings), with the TDA mean particle size (or D_{50}) ranging between 0.46 and 3.34 mm. Furthermore, in terms of mix design, the TDA content varied between 2.5% and 30% (the latter considered to be the highest possible whilst still maintaining mixture homogeneity), and each of the 21 datasets, in addition to the unamended soil ($f_T = 0$), included a minimum of three compaction test data for three different TDA contents. It should be mentioned that the experimental OMC values ranged between 12.4% and 28.0% water content, with the complete results of the compaction tests presented in Figure A1 of the Appendix A.

3. Governing Mechanisms Controlling the Compactability of Soil-TDA Blends

It is generally accepted that the addition of (and content increase in) TDA, with constant particle size/shape, leads to a 'leftward–downward' translation of the soil compaction curve (for a given compactive effort), causing notable reductions in the OMC and MDUW parameters [24–26]. The TDA material's lower specific gravity (or density) compared with that of the soil solids has been reported to be the primary factor responsible for decreasing the MDUW [15]. For the compiled soil–TDA database used in this investigation (see Tables 1 and 2), the TDA-to-soil specific gravity ratio (i.e., $G_{\rm s}^{\rm T}/G_{\rm s}^{\rm S}$) was found to range between 0.31 and 0.44. Moreover, some researchers have postulated that, because of their high energy absorption capacity (attributed to their high elasticity), the compacted TDA particles may progressively recover their initial uncompacted shapes through a so-called

Sustainability **2021**, 13, 7737 5 of 21

'elastic-rebound' recovery mechanism, thereby offsetting the efficiency of the imparted compactive effort, also contributing towards decreasing the MDUW [17,34,41]. The reductions reported for the OMC have been mainly ascribed to the TDA material's hydrophobic character and hence its lower water adsorption–retention capacity (being mainly less than 4%) compared with that of the fine soil particles, particularly clays [16]. It should be mentioned that, while the reported results for the MDUW are fairly consistent, a limited number of studies have reported either negligible or increasing trends for the OMC with respect to increasing the TDA content [42–44]. These unexpected trends may be attributed to TDA segregation (and hence TDA clustering) effects caused by inadequate soil–TDA mixing during sample preparation for the compaction test [19,39]. As such, in compiling our database for the present investigation, it was decided to include only those datasets which are consistent with the more unanimous 'OMC-decreasing' trend.

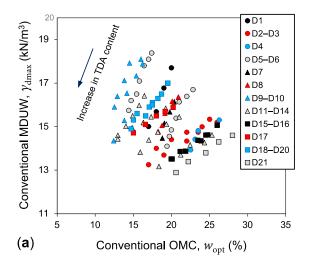
It is well accepted that the MDUW and OMC for unamended fine-grained soils (irrespective of compactive effort) are strongly correlated, following a unique 'path of optimums' somewhat parallel to the standard zero-air-voids (ZAV) saturation line (commonly obtained for a typical specific gravity of 2.65 for the soil solids) [27,29,31]. For instance, in their investigation, Gurtug and Sridharan [27] reported the following 'path of optimums' relationships based on a database of 181 compaction tests (with OMC water contents ranging $w_{\text{opt}} = 7.4\text{--}49.0\%$) involving a variety of 'unamended' fine-grained soils tested at four different compaction energy levels (i.e., reduced, standard, reduced modified and modified Proctor): (i) $\gamma_{\rm dmax} = -0.28~w_{\rm opt} + 22.26$ (with R² = 0.941); and (ii) $\gamma_{\rm dmax} = 23.68~exp[-0.018~w_{\rm opt}]$ (with R² = 0.960), where $\gamma_{\rm dmax}$ is the deduced MDUW value. Figure 1a illustrates the variations of MDUW against OMC for the compiled database of 100 soil-TDA compaction tests (data values presented in Figure A1). As is evident from this figure, the data points are significantly scattered, indicating that the MDUW and OMC are poorly correlated for the investigated soil-TDA blends. For unamended finegrained soils, an increase in the coarse fraction (>75 µm) leads to an 'upward–leftward' translation of the compaction curve, with the optimum (peak) point translating along the previously described universal 'path of optimums'. For soil-TDA blends, however, an increase in TDA content (0.075–4.75 mm particle size range), which is essentially similar to increasing the soil coarse fraction, results in a 'downward-leftward' translation of the optimum point, implying that soil-TDA blends do not conform to the general 'path of optimums' correlation framework. This discrepancy can be attributed to the significant mismatch in density (and hence specific gravity) between the soil solids and TDA particles, allowing one to postulate that the lower MDUW values obtained for soil-TDA blends may not necessarily reflect their lower compactability potential. To achieve a more familiar visualization of the compaction characteristics of soil-TDA blends, consistent with the traditional soil compaction framework (for unamended soils), the conventional OMC and MDUW parameters, w_{opt} and γ_{dmax} , should be 'normalized' as follows [19]:

$$w_{\text{opt}}^* = w_{\text{opt}} \left(\frac{G_{\text{s}}^{\text{ST}}}{G_{\text{s}}^{\text{S}}} \right) \tag{1}$$

$$\gamma_{\rm dmax}^* = \gamma_{\rm dmax} \left(\frac{G_{\rm s}^{\rm S}}{G_{\rm s}^{\rm ST}} \right) \tag{2}$$

where w_{opt}^* and γ_{dmax}^* = normalized OMC and MDUW parameters, respectively; and G_s^{S} and G_s^{ST} = specific gravity of soil solids (values presented in Table 1) and soil–TDA mixture (values presented in Table A1 of the Appendix A), respectively.

Sustainability **2021**, 13, 7737 6 of 21



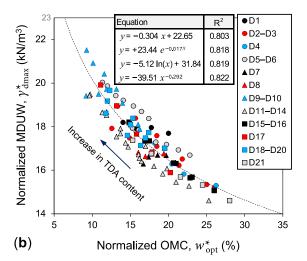


Figure 1. Variations of MDUW against OMC for the compiled database of 100 soil–TDA compaction tests: (a) Conventional definition; and (b) Normalized definition according to Equations (1) and (2).

It should be mentioned that the soil–TDA mixture specific gravity $G_{\rm s}^{\rm ST}$, which decreases with increasing the TDA content (i.e., $G_{\rm s}^{\rm ST} \sim f_{\rm T}^{-1}$, as demonstrated in Table A1), was obtained as follows [34,45]:

$$G_{\rm s}^{\rm ST} = \frac{G_{\rm s}^{\rm S} G_{\rm s}^{\rm T} (M_{\rm S} + M_{\rm T})}{G_{\rm s}^{\rm S} M_{\rm T} + G_{\rm s}^{\rm T} M_{\rm S}} = \frac{G_{\rm s}^{\rm S} G_{\rm s}^{\rm T} (1 + f_{\rm T})}{G_{\rm s}^{\rm S} f_{\rm T} + G_{\rm s}^{\rm T}}$$
(3)

where G_s^T = specific gravity of TDA particles (values presented in Table 2); M_S and M_T = mass of oven-dried soil and TDA, respectively; and f_T = TDA content, defined as the TDA-to-soil dry mass ratio (or M_T/M_S).

Figure 1b illustrates the variations of $\gamma_{\rm dmax}^*$ against $w_{\rm opt}^*$ for the compiled database. As is evident from this figure, the normalized MDUW and OMC, in addition to showcasing a strong correlation, conform to the general 'path of optimums' framework described earlier for unamended fine-grained soils. As a typical example, exponential-fitting of the normalized compaction data resulted in $\gamma_{\rm dmax}^* = 23.44~exp[-0.017~w_{\rm opt}^*]$ (with $R^2 = 0.818$) for $f_{\rm T} \leq 30\%$, which is essentially identical to $\gamma_{\rm dmax}$ – $w_{\rm opt}$ relationships previously reported for unamended fine-grained soils by Gurtug and Sridharan [27] and Sivrikaya et al. [29].

4. Results and Discussion

4.1. Modeling Premise

Following a comprehensive trial-and-error investigation employing the 21 soil–TDA compaction datasets (see Table A1 and Figure A1), it was observed that, for a given fine-grained soil mixed with a particular TDA material (constant particle size/shape), the conventional OMC and MDUW parameters (for the standard or modified Proctor energy level) can be expressed as follows:

$$w_{\text{opt}}^{\text{ST}} = w_{\text{opt}}^{\text{S}} \left(\frac{G_{\text{s}}^{\text{S}}}{G_{\text{s}}^{\text{ST}}}\right)^{\beta_{\text{M}}} \tag{4}$$

$$\gamma_{\rm dmax}^{\rm ST} = \gamma_{\rm dmax}^{\rm S} \left(\frac{G_{\rm s}^{\rm S}}{G_{\rm s}^{\rm ST}}\right)^{\beta_{\rm D}} \tag{5}$$

where $w_{\rm opt}^{\rm ST}$ and $\gamma_{\rm dmax}^{\rm ST}$ = conventional OMC and MDUW for the soil–TDA mixture, respectively; $w_{\rm opt}^{\rm S}$ and $\gamma_{\rm dmax}^{\rm S}$ = intercept parameters for $f_{\rm T}$ = 0 (in % and kN/m³, respectively); and $\beta_{\rm M}$ and $\beta_{\rm D}$ = reduction rate parameters (both < 0).

Sustainability **2021**, 13, 7737 7 of 21

The dependent/input variable G_s^S/G_s^{ST} selected for model development captures the combined effects of TDA content and TDA specific gravity; the latter well-established to vary (i.e., 0.85–1.14 for the present investigation, as outlined in Table 2) depending on the source-tire composition, the adopted tire recycling process, and the TDA particle size/shape [24,26]. The intercept parameters $w_{\rm opt}^S$ and $\gamma_{\rm dmax}^S$ represent the OMC and MDUW of the unamended soil, since setting $f_{\rm T}=0$ in Equation (3) results in $G_s^{\rm ST}=G_s^{\rm S}$. Note that these intercept parameters can be either fixed based on measured values or set as independent fitting parameters. Further, since the soil–TDA mixture specific gravity decreases with increasing the TDA content (i.e., $G_s^{\rm ST}\sim f_{\rm T}^{-1}$ and hence $G_s^{\rm S}/G_s^{\rm ST}\sim f_{\rm T}$; see Table A1), the parameters $\beta_{\rm M}$ and $\beta_{\rm D}$ (which are both negative) represent the rates of reduction in the OMC and MDUW, respectively, in relation to increasing the TDA content.

It should be mentioned that the trial-and-error investigation performed by the authors and leading to the proposal of Equations (4) and (5) involved applying various functional expressions (i.e., linear, logarithmic, polynomial, exponential and power) to the 21 soil—TDA compaction datasets and then cross-checking their predictive performances using routine fit-measure indices; namely, the mean absolute percentage error (MAPE) and the normalized root-mean-squared error (NRMSE), which were calculated as follows [46]:

MAPE =
$$\frac{1}{N} \sum_{n=1}^{N} \left| \frac{y_n - \hat{y}_n}{y_n} \right| \times 100\%$$
 (6)

$$NRMSE = \frac{RMSE}{\overline{y_n}} \times 100\% \tag{7}$$

RMSE =
$$\sqrt{\frac{1}{N} \sum_{n=1}^{N} (y_n - \hat{y}_n)^2}$$
 (8)

where RMSE = root-mean-squared error (in the same unit as OMC or MDUW); y_n = measured variable (OMC or MDUW); \hat{y}_n = predicted variable (OMC or MDUW); \overline{y}_n = arithmetic mean of y_n data; and N = number of observations (or compaction tests) in each dataset, as reported in Table 1.

The regression analysis outputs with respect to Equations (4) and (5) (with $w_{\rm opt}^{\rm S}$ and $\gamma_{\rm dmax}^{\rm S}$ set as independent fitting parameters) are presented in Tables 3 and 4, respectively. Judging by the high R² (with median values of 0.989 and 0.987 for the OMC and MDUW predictions, respectively) and the low MAPE or NRMSE (unanimously less than 4%) values, the functional expressions proposed in Equations (4) and (5) can be deemed acceptable. Quite clearly, to employ Equations (4) and (5) for routine prediction purposes, the fitting parameters $\beta_{\rm M}$ and $\beta_{\rm D}$ should be calibrated. In view of their definitions, the intercept parameters $w_{\rm opt}^{\rm S}$ and $\gamma_{\rm dmax}^{\rm S}$ can be simply fixed based on the measured OMC and MDUW of the unamended soil ($f_{\rm T}$ = 0). Provided that the reduction rate parameters $\beta_{\rm M}$ and $\beta_{\rm D}$ can be practically calibrated without the need for obtaining any specific soil–TDA compaction test data, it would follow that, having established the OMC and MDUW of an unamended fine-grained soil (along with the soil and TDA specific gravities; the latter often provided by the TDA manufacturer), one can predict the OMC and MDUW of the same soil mixed with any specified TDA content. The following sections describe practical calibration frameworks for obtaining $\beta_{\rm M}$ and $\beta_{\rm D}$.

Sustainability **2021**, 13, 7737 8 of 21

Table 3. Summary of the regression analysis outputs with respect to Equation (4) (OMC model).

| Dataset | N | $w_{\mathrm{opt}}^{\mathrm{S}}(\%)$ | $eta_{	ext{M}}$ | \mathbb{R}^2 | MAPE (%) | NRMSE (%) |
|---------|---|-------------------------------------|-----------------|----------------|----------|-----------|
| D1 | 4 | 20.06 | -0.842 | 0.997 | 0.32 | 0.32 |
| D2 | 5 | 25.99 | -1.275 | 0.931 | 3.19 | 3.77 |
| D3 | 5 | 26.56 | -1.202 | 0.977 | 1.88 | 1.97 |
| D4 | 4 | 25.45 | -0.668 | 0.904 | 2.54 | 2.72 |
| D5 | 7 | 22.37 | -0.639 | 0.944 | 1.10 | 1.33 |
| D6 | 7 | 17.37 | -0.788 | 0.993 | 0.47 | 0.53 |
| D7 | 5 | 20.93 | -0.944 | 0.989 | 0.38 | 0.42 |
| D8 | 5 | 20.83 | -1.057 | 0.985 | 0.48 | 0.59 |
| D9 | 5 | 15.87 | -0.817 | 0.991 | 0.80 | 0.85 |
| D10 | 5 | 16.44 | -0.839 | 0.952 | 1.74 | 1.93 |
| D11 | 5 | 20.74 | -0.925 | 0.990 | 0.87 | 1.02 |
| D12 | 5 | 16.53 | -0.813 | 0.992 | 0.66 | 0.70 |
| D13 | 5 | 25.74 | -1.046 | 0.956 | 2.05 | 2.22 |
| D14 | 5 | 19.17 | -1.263 | 0.987 | 1.21 | 1.40 |
| D15 | 5 | 26.28 | -0.862 | 0.987 | 0.93 | 1.04 |
| D16 | 5 | 26.08 | -0.905 | 0.996 | 0.45 | 0.57 |
| D17 | 5 | 20.41 | -0.976 | 0.992 | 0.84 | 0.90 |
| D18 | 4 | 19.80 | -1.009 | 0.977 | 1.31 | 1.36 |
| D19 | 4 | 19.65 | -1.163 | 0.993 | 0.78 | 0.81 |
| D20 | 4 | 19.34 | -1.241 | 0.992 | 0.74 | 0.93 |
| D21 | 5 | 28.04 | -1.038 | 0.999 | 0.19 | 0.23 |

Note: N = number of compaction tests; R^2 = coefficient of determination; MAPE = mean absolute percentage error (Equation (6)); NRMSE = normalized root-mean-squared error (Equation (7)); $w_{\rm opt}^{\rm S}$ = OMC intercept parameter; and $\beta_{\rm M}$ = OMC reduction rate parameter for increasing TDA content.

Table 4. Summary of the regression analysis outputs with respect to Equation (5) (MDUW model).

| Dataset | N | $\gamma_{\rm dmax}^{\rm S}$ (kN/m ³) | β_{D} | R ² | MAPE (%) | NRMSE (%) |
|---------|---|--|----------------------|----------------|----------|-----------|
| | 4 | 17.72 | -0.874 | 0.996 | 0.35 | 0.39 |
| D2 | 5 | 15.42 | -0.481 | 0.984 | 0.51 | 0.65 |
| D3 | 5 | 15.32 | -0.265 | 0.937 | 0.69 | 0.73 |
| D4 | 4 | 15.27 | -0.440 | 0.995 | 0.23 | 0.24 |
| D5 | 7 | 16.76 | -0.714 | 0.996 | 0.34 | 0.40 |
| D6 | 7 | 18.46 | -0.655 | 0.996 | 0.29 | 0.35 |
| D7 | 5 | 16.56 | -1.030 | 0.944 | 1.01 | 1.05 |
| D8 | 5 | 16.49 | -0.846 | 0.966 | 0.62 | 0.74 |
| D9 | 5 | 18.38 | -0.454 | 0.942 | 1.08 | 1.24 |
| D10 | 5 | 18.29 | -0.734 | 0.938 | 1.79 | 2.05 |
| D11 | 5 | 15.48 | -0.368 | 0.987 | 0.36 | 0.42 |
| D12 | 5 | 16.38 | -0.404 | 0.996 | 0.21 | 0.27 |
| D13 | 5 | 14.66 | -0.336 | 0.968 | 0.61 | 0.64 |
| D14 | 5 | 15.81 | -0.290 | 0.976 | 0.42 | 0.48 |
| D15 | 5 | 15.04 | -0.360 | 0.998 | 0.13 | 0.16 |
| D16 | 5 | 15.03 | -0.359 | 0.997 | 0.18 | 0.22 |
| D17 | 5 | 15.96 | -0.254 | 0.984 | 0.33 | 0.35 |
| D18 | 4 | 16.96 | -0.371 | 0.995 | 0.21 | 0.22 |
| D19 | 4 | 16.92 | -0.460 | 0.989 | 0.39 | 0.41 |
| D20 | 4 | 16.82 | -0.580 | 0.964 | 0.91 | 0.93 |
| D21 | 5 | 14.63 | -0.415 | 0.990 | 0.41 | 0.44 |

Note: N = number of compaction tests; R^2 = coefficient of determination; MAPE = mean absolute percentage error (Equation (6)); NRMSE = normalized root-mean-squared error (Equation (7)); $\gamma_{\rm dmax}^{\rm S}$ = MDUW intercept parameter; and $\beta_{\rm D}$ = MDUW reduction rate parameter for increasing TDA content.

4.2. Predictive Models Employing Mean Reduction Rate Parameters

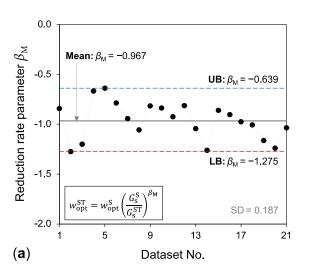
Figure 2a,b illustrates the variations of β_M and β_D for the 21 soil–TDA compaction datasets, respectively. In terms of absolute magnitude, β_M was found to be consistently greater than its β_D counterpart, indicating that the rate of OMC reduction with respect to

Sustainability **2021**, 13, 7737 9 of 21

increasing TDA content is greater than that of the MDUW. Judging by the low standard deviation (SD) for β_M and β_D (computed as 0.187 and 0.218, respectively), as well as the relatively small vertical distance between their upper and lower variation boundaries (see 'UB' and 'LB' in Figure 2), it may be possible to achieve reliable OMC and MDUW predictions (across different fine-grained soil types, TDA particle sizes/shapes and compaction energy levels) by adopting mean values for the β_M and β_D parameters. To examine this hypothesis, the arithmetic means for the 21 β_M and β_D values were calculated (i.e., $\overline{\beta_M} = -0.967$ and $\overline{\beta_D} = -0.509$, as outlined in Figure 2), and appointed to Equations (4) and (5), resulting in the following new relationships:

$$w_{\text{opt}}^{\text{ST}} = w_{\text{opt}}^{\text{S}} \left(\frac{G_{\text{s}}^{\text{S}}}{G_{\text{s}}^{\text{ST}}}\right)^{-0.967} \tag{9}$$

$$\gamma_{\rm dmax}^{\rm ST} = \gamma_{\rm dmax}^{\rm S} \left(\frac{G_{\rm s}^{\rm S}}{G_{\rm s}^{\rm ST}}\right)^{-0.509} \tag{10}$$



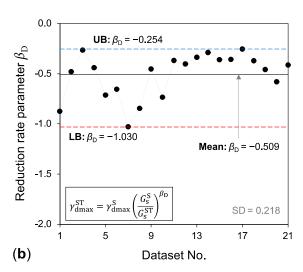
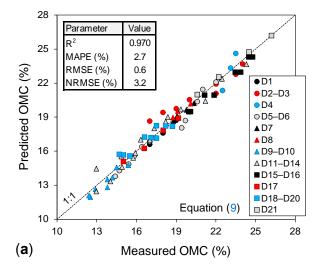


Figure 2. Variations of (a) $\beta_{\rm M}$ and (b) $\beta_{\rm D}$ for the 21 soil–TDA compaction datasets investigated. Note: UB and LB denote the upper and lower variation boundaries, respectively.

Scatter plots illustrating the variations of predicted (by Equations (9) and (10)) against measured OMC and MDUW values are presented in Figure 3a,b, respectively. As is evident from these figures, the predicted and measured values, particularly for the OMC, are strongly correlated with each other. The R², MAPE and NRMSE associated with these predictions were, respectively, calculated as 0.970, 2.7% and 3.2% for the OMC, and 0.908, 2.6% and 3.2% for the MDUW.

The excellent graphical correlation (high R^2), together with the low MAPE or NRMSE values, obtained for Equations (9) and (10) would normally lead to corroborating their predictive capability. However, a critical examination of the prediction residuals should also be performed to better perceive the true implications of these predictions for routine geotechnical engineering applications [47]. This can be achieved by quantifying and critically examining the statistical 'limits of agreement' between the predicted and measured values, which was conducted using the Bland–Altman (BA) analysis [48]. The BA analysis involves developing a scatter plot with the y-axis representing the difference between the two compared variables (e.g., $OMC_P - OMC_M$, where the subscripts 'P' and 'M' denote predicted and measured variables, respectively) and the x-axis showing the average of these variables (e.g., $OMC_P + OMC_M$)/2). The 95% upper and lower agreement limits with respect to the BA plot can be, respectively, quantified as UAL = Mean + 1.96 × SD and LAL = Mean $- 1.96 \times SD$ (where 'Mean' and 'SD' are the arithmetic mean and standard deviation of the y-axis data, respectively).

Sustainability **2021**, 13, 7737 10 of 21



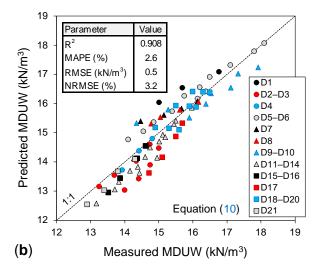
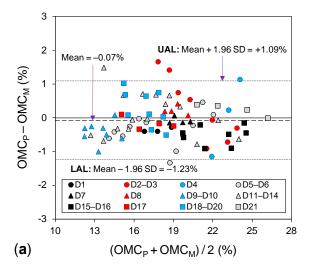


Figure 3. Variations of predicted against measured compaction parameters for the 21 soil–TDA compaction datasets: (a) OMC (Equation (9)); and (b) MDUW (Equation (10)).

BA plots for the OMC and MDUW predictions (Equations (9) and (10)) are presented in Figure 4a,b, respectively. The mean of differences between OMC_P and OMC_M was found to be -0.07%, indicating that the OMC predictions were on average 0.07% (water content) lower than their measured counterparts. The 95% agreement limits between OMC_P and OMC_M were calculated as UAL = +1.09% and LAL = -1.23%, implying that 95% of the predictions made by Equation (9) are associated with errors ranging between these two water content limits, both of which can be considered negligible for practical applications. As for the MDUW (see Figure 4b), the mean of differences, UAL and LAL were obtained as -0.12, +0.80 and -1.04 kN/m³, respectively. Taking into account the nature of the MDUW parameter and its variations across different fine-grained soil types and also with standard and modified compaction energy levels (these variations being relatively smaller compared with that of the OMC [27,31,33]), the errors associated with Equation (10), though practically acceptable, may require further improvement. Alternatively, having predicted the OMC by Equation (9), the corresponding MDUW can be estimated with more accuracy through a practical single-point compaction test (performed at the predicted OMC).



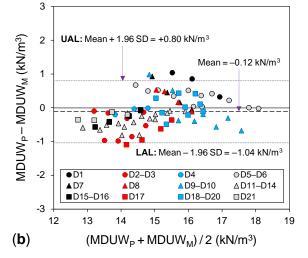


Figure 4. Bland–Altman plots for the (a) OMC and (b) MDUW predictions (made by Equations (9) and (10), respectively). Note: UAL and LAL denote the 95% upper and lower agreement limits, respectively.

Sustainability **2021**, 13, 7737 11 of 21

4.3. Prediction Models Employing Empirical Reduction Rate Parameters

The authors postulate that the reduction rate parameters $\beta_{\rm M}$ and $\beta_{\rm D}$ may be systematically related to basic soil properties (namely those listed in Table 1). Accordingly, attempts were made to explore the existence of potential links/correlations between these fitting parameters and other parameters reflective of the soil gradation, plasticity and mineralogy. Following a comprehensive statistical analysis of the data, no meaningful correlation was found for $\beta_{\rm M}$. However, it was observed that $|\beta_{\rm D}|$ systematically decreases with increasing the soil activity index (i.e., $|\beta_{\rm D}| \sim A^{-1}$). In other words, as the soil's principal clay mineral becomes more active (e.g., kaolinite to montmorillonite), the rate of reduction in the MDUW (with respect to increasing TDA content) decreases. Figure 5 illustrates the variations of $|\beta_{\rm D}|$ against the activity index for the compiled database (excluding datasets D1, D9 and D10 for which the clay contents were not reported). As demonstrated in this figure, $\beta_{\rm D}$ can be expressed as follows (for $f_{\rm T} \leq 30\%$):

$$\beta_{\rm D} = 0.269 \ln \left(\frac{\rm PI}{f_{\rm clay}} \right) - 0.311 \tag{11}$$

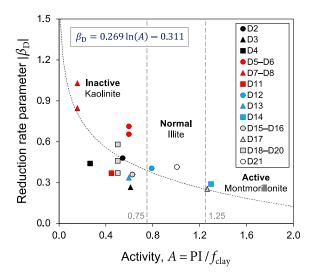


Figure 5. Variations of $|\beta_D|$ against the activity index for the compiled database of soil–TDA compaction tests (excluding datasets D1, D9 and D10 for which the soil activity index could not be calculated due to non-reporting of their clay contents).

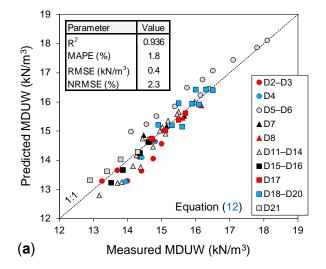
Accordingly, substituting Equation (11) into Equation (5) leads to the following new relationship for the MDUW:

$$\gamma_{\rm dmax}^{\rm ST} = \gamma_{\rm dmax}^{\rm S} \left(\frac{G_{\rm s}^{\rm S}}{G_{\rm s}^{\rm ST}}\right)^{[0.269 \ln{(\frac{\rm PI}{f_{\rm clay}})} - 0.311]} \tag{12}$$

Figure 6a illustrates the variations of predicted (by Equation (12)) against measured MDUW values for the compiled database. The R^2 , MAPE and NRMSE for these new predictions were calculated as 0.936, 1.8% and 2.3%, respectively; corroborating the predictive capability of the newly proposed Equation (12). The 95% upper and lower agreement limits, as shown in Figure 6b, were obtained as UAL = $+0.67 \text{ kN/m}^3$ and LAL = -0.71 kN/m^3 , indicating that 95% of the MDUW predictions are associated with errors ranging between these two small unit weight limits. Note that Equations (10) and (12) were developed based on different dataset sizes (i.e., 21 and 18 datasets, respectively); as such, their predictive performances cannot be directly compared. However, a reliable comparison can be performed if the R^2 , MAPE, NRMSE, UAL and LAL parameters for Equation (10) are recalculated based on the same 18 datasets (i.e., D2–D8 and D11–D21) used for the development of Equation (12). The outcome of this recalculation was $R^2 = 0.920$, MAPE = 2.5%,

Sustainability **2021**, 13, 7737 12 of 21

NRMSE = 3.1%, UAL = +0.69 kN/m³ and LAL = -1.0 kN/m³, which appear to work (slightly) in favor of the more elaborate Equation (12). Even so, for prediction purposes, this performance improvement may not be sufficient to justify the use of Equation (12) over the more practical Equation (10); the latter making MDUW predictions without the need for PI and $f_{\rm clav}$ measurements.



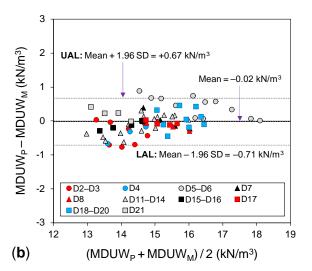


Figure 6. MDUW predictions made by Equation (12) (excluding datasets D1, D9 and D10): (a) Variations against measured MDUW; and (b) Bland–Altman plot. Note: UAL and LAL denote the 95% upper and lower agreement limits, respectively.

In addition to fundamental soil properties, the authors speculated that $\beta_{\rm M}$ and $\beta_{\rm D}$ may also be related to other variables, such as the TDA mean particle size (or D_{50}) and the imparted compaction energy level. Even though the compiled database did not permit a critical investigation of these variables to be performed (i.e., since only a small number of the database soils included compaction results for varying D_{50} and/or compaction energy levels), it is considered that both D_{50} and compaction energy would likely have minor effects on $\beta_{\rm M}$ and $\beta_{\rm D}$. As mentioned in Section 4.1, for predominantly sand-sized TDA materials (0.075–4.75 mm), changes in the TDA mean particle size is normally reflected in the TDA specific gravity [24,26]. In other words, the dependent/input variable $G_{\rm S}^{\rm S}/G_{\rm S}^{\rm ST}$ not only captures the combined effects of TDA content and TDA density, but it is also expected to account, at least in part, for changes in the TDA mean particle size. Moreover, a review of the admittedly limited literature (including those listed in Tables 1 and 2) indicates that, for TDA materials (i.e., powder, crumbs and buffings) having the same specific gravity but different D_{50} values, the variations in OMC and MDUW across the two TDA sizes are relatively small.

The elastic-rebound recovery exhibited by TDA particles in compacted soil–TDA mixtures has been reported to increase with increasing the compactive effort. In other words, the higher the imparted compaction energy level (from standard to modified Proctor), the lower the compaction efficiency of the soil–TDA matrix [25,41]. This may explain the limited soil–TDA compaction data reported for the modified (heavy) Proctor energy level (accounting for only four of the twenty-one cases listed in Tables 1 and 2). In view of this mechanism, it is speculated that the beneficial effects of compaction energy increase (from standard to modified Proctor) would likely be offset by the TDA material's increased energy dissipation potential, allowing one to postulate that the reduction rate parameters, particularly β_D , may not be significantly influenced by compactive effort. As such, the modeling framework proposed in this investigation allocates similar OMC and MDUW reduction rates for standard and modified compaction energy levels. Given that the bulk of the compiled database used for model development consisted of standard Proctor compaction data (17 datasets out of 21 examined), the predictions made for modified Proctor should be taken with some caution. Nevertheless, a systematically controlled test

Sustainability **2021**, 13, 7737 13 of 21

program involving a variety of TDA particle sizes and a range of compaction energy levels should be performed with the dual aims of checking the above postulations and potentially developing improved empirical correlations for the reduction rate parameters $\beta_{\rm M}$ and $\beta_{\rm D}$.

5. Summary and Conclusions

This study aimed at modeling the compaction characteristics of fine-grained soils blended with sand-sized TDA products (e.g., powder, crumbs and buffings). Model development and calibration were carried out using a large and diverse database of 100 soil–TDA compaction tests (with $f_{\rm T} \leq 30\%$) assembled from the literature. Following a comprehensive statistical analysis of the data, the following general and fundamental conclusions can be drawn from this study:

- Irrespective of the imparted compaction energy level (from standard to modified Proctor), the addition of (and content increase in) TDA leads to notable reductions in the OMC and MDUW parameters, both following exponentially decreasing trends with respect to increasing TDA content.
- The OMC and MDUW for soil-TDA blends (across different fine-grained soil types, TDA particle sizes and compaction energy levels) can be expressed as universal power functions of the OMC and MDUW of the unamended soil, together with the soil to soil-TDA specific gravity ratio; the latter capable of capturing the combined effects of TDA content and its lower density.
- Making use of the Bland–Altman analysis, the 95% upper and lower (water content) agreement limits between the predicted and measured OMC values were, respectively, obtained as +1.09% and −1.23%, both of which can be considered negligible for practical applications. For the MDUW predictions, these limits were calculated as +0.80 and −1.04 kN/m³ and, employing a more elaborate correlation that also considers the soil activity, as +0.67 and −0.71 kN/m³, which (like the OMC) can be deemed acceptable for prediction purposes. Accordingly, having established the OMC and MDUW of the unamended fine-grained soil, the various empirical models proposed in this study offer a practical procedure towards predicting the compaction characteristics of the soil–TDA blends without the hurdles of performing separate laboratory compaction tests, and thus can be used for preliminary design assessments and/or soil–TDA optimization studies.

Further investigations are warranted regarding the possible application of the new modeling framework developed for fine-grained soil–TDA blends in predicting the compaction characteristics of fine-grained soils when mixed with other recycled solid waste and/or virgin materials.

Author Contributions: Conceptualization, A.S. and M.A.; methodology, A.S. and M.A.; validation, A.S., M.A. and B.C.O.; formal analysis, A.S. and M.A.; investigation, A.S. and M.A.; writing—original draft preparation, A.S. and M.A.; writing—review and editing, A.S. and B.C.O.; visualization, A.S. and M.A.; supervision, A.S.; funding acquisition, B.C.O. All authors have read and agreed to the published version of the manuscript.

Funding: This research received no external funding.

Data Availability Statement: This study has not generated new experimental data.

Acknowledgments: Special thanks go to Mark B. Jaksa of the University of Adelaide for valuable suggestions to the authors.

Conflicts of Interest: The authors declare no conflict of interest.

Sustainability **2021**, 13, 7737 14 of 21

Abbreviations

BA Bland-Altman (analysis/plot)

BS British Standard

BTEX Benzene, Toluene, Ethylbenzene and Xylene

CH Clay with high plasticity

CI Clay with intermediate plasticity
CV Clay with very high plasticity

ELT End-of-life tire

LB Lower (variation) boundary **MDUW** Maximum dry unit weight MH Silt with high plasticity MLSilt with low plasticity MVSilt with very high plasticity OMC Optimum moisture content SP Poorly-graded (sand) TDA Tire-derived aggregate UB Upper (variation) boundary **USCS** Unified Soil Classification System

ZAV Zero-air-voids

Notations

 $\begin{array}{ll} A & \text{Soil activity index} \\ C_{\text{C}} & \text{Coefficient of curvature} \\ C_{\text{U}} & \text{Coefficient of uniformity} \\ D_{50} & \text{TDA mean particle size (mm)} \\ f_{\text{clay}} & \text{Clay (<2 } \mu\text{m) content (\%)} \\ f_{\text{fines}} & \text{Fines (<75 } \mu\text{m) content (\%)} \end{array}$

 $f_{\rm T}$ TDA content (i.e., TDA-to-soil dry mass ratio) (%)

 G_s^S Specific gravity of soil solids G_s^T Specific gravity of TDA particles G_s^{ST} Specific gravity of soil–TDA mixture

LAL Lower (statistical) agreement limit (same unit as OMC or MDUW)

LL Liquid limit (%)

 $\begin{array}{ll} \text{MAPE} & \text{Mean absolute percentage error (\%)} \\ \text{MDUW}_{M} & \text{Measured MDUW (kN/m}^{3}) \\ \text{MDUW}_{P} & \text{Predicted MDUW (kN/m}^{3}) \\ M_{S} & \text{Mass of oven-dried soil (g)} \end{array}$

 $M_{\rm T}$ Mass of TDA (g) n Index of summation

N Number of observations (or compaction tests) NRMSE Normalized root-mean-squared error (%)

OMC_M Measured OMC (%)
OMC_P Predicted OMC (%)
PI Plasticity index (%)

R² Coefficient of determination

RMSE Root-mean-squared error (same unit as OMC or MDUW)
SD Standard deviation (same unit as OMC or MDUW)

UAL Upper (statistical) agreement limit (same unit as OMC or MDUW)

 w_{opt} Conventional OMC (%) w_{opt}^{*} Normalized OMC (%) $w_{\mathrm{opt}}^{\mathrm{S}}$ OMC of unamended soil (%) $w_{\mathrm{opt}}^{\mathrm{ST}}$ OMC of soil–TDA mixture (%) y_{n} Measured variable (OMC or MDUW)

 $\overline{y_n}$ Arithmetic mean of y_n data

 \hat{y}_n Predicted variable (OMC or MDUW)

 $\beta_{\rm D}$ Reduction rate parameter for increasing TDA content (MDUW model)

Sustainability **2021**, 13, 7737 15 of 21

 $\begin{array}{ll} \beta_{M} & \text{Reduction rate parameter for increasing TDA content (OMC model)} \\ \gamma_{dmax} & \text{Conventional MDUW (kN/m}^{3}) \\ \gamma_{dmax}^{*} & \text{Normalized MDUW (kN/m}^{3}) \\ \gamma_{dmax}^{ST} & \text{MDUW of unamended soil (kN/m}^{3})} \\ \gamma_{dmax}^{ST} & \text{MDUW of soil-TDA mixture (kN/m}^{3})} \end{array}$

Appendix A

The specific gravity values of the fine-grained soil–TDA mixtures—that is, $G_{\rm s}^{\rm ST}$ calculated by Equation (3)—for the compiled database are presented in Table A1.

Table A1. Specific gravity values of the fine-grained soil–TDA mixtures for the compiled database.

| Test ID | Dataset | Source | Soil ID | $G_{\mathrm{s}}^{\mathrm{S}}$ | $G_{ m s}^{ m T}$ | f _T (%) | $G_{ m s}^{ m ST}$ | $G_{\rm s}^{\rm S}/G_{\rm s}^{\rm ST}$ |
|---------|---------|-----------------------|---------|-------------------------------|-------------------|--------------------|--------------------|--|
| T1 | | | | | | 0 | 2.61 | 1 |
| T2 | D1 | Cabalar et al. [15] | S1 | 2.61 | 1.08 | 5.3 | 2.44 | 1.070 |
| T3 | Dī | | 01 | 2.01 | 1.00 | 11.1 | 2.29 | 1.140 |
| T4 | | | | | | 17.6 | 2.15 | 1.214 |
| T5 | | | | | | 0 | 2.65 | 1 |
| Т6 | | | | | | 5.0 | 2.41 | 1.100 |
| T7 | D2 | Prasad et al. [36] | S2 | 2.65 | 0.85 | 10.0 | 2.22 | 1.194 |
| T8 | | | | | | 15.0 | 2.08 | 1.274 |
| T9 | | | | | | 20.0 | 1.96 | 1.352 |
| T10 | | | | | | 0 | 2.65 | 1 |
| T11 | | | | | | 5.0 | 2.41 | 1.100 |
| T12 | D3 | Prasad et al. [36] | S3 | 2.65 | 0.85 | 10.0 | 2.22 | 1.194 |
| T13 | | | | | | 15.0 | 2.08 | 1.274 |
| T14 | | | | | | 20.0 | 1.96 | 1.352 |
| T15 | | | | | | 0 | 2.72 | 1 |
| T16 | D.1 | D : (1 [07] | 0.4 | | 4.40 | 5.0 | 2.55 | 1.067 |
| T17 | D4 | Ramirez et al. [37] | S4 | 2.72 | 1.12 | 10.0 | 2.41 | 1.129 |
| T18 | | | | | | 20.0 | 2.20 | 1.236 |
| T19 | | | | | | 0 | 2.69 | 1 |
| T20 | | | | | | 2.5 | 2.60 | 1.035 |
| T21 | | | | | | 5.0 | 2.53 | 1.063 |
| T22 | D5 | Signes et al. [16] | S5 | 2.69 | 1.14 | 10.0 | 2.39 | 1.126 |
| T23 | | | | | | 15.0 | 2.28 | 1.180 |
| T24 | | | | | | 20.0 | 2.19 | 1.228 |
| T25 | | | | | | 25.0 | 2.11 | 1.275 |
| T26 | | | | | | 0 | 2.69 | 1 |
| T27 | | | | | | 2.5 | 2.60 | 1.035 |
| T28 | | | | | | 5.0 | 2.53 | 1.063 |
| T29 | D6 | Signes et al. [16] | S5 | 2.69 | 1.14 | 10.0 | 2.39 | 1.126 |
| T30 | 20 | 8 | 30 | , | 2.2.2 | 15.0 | 2.28 | 1.180 |
| T31 | | | | | | 20.0 | 2.19 | 1.228 |
| T32 | | | | | | 25.0 | 2.11 | 1.275 |
| T33 | | | | | | 0 | 2.69 | 1 |
| T34 | | | | | | 2.5 | 2.60 | 1.035 |
| T35 | D7 | Yadav and Tiwari [43] | S6 | 2.69 | 1.13 | 5.0 | 2.52 | 1.067 |
| T36 | 51 | | 20 | | 1.10 | 7.5 | 2.45 | 1.098 |
| T37 | | | | | | 10.0 | 2.39 | 1.126 |
| T33 | | | | | | 0 | 2.69 | 1 |
| T38 | | | | | | 2.5 | 2.59 | 1.039 |
| T39 | D8 | Yadav and Tiwari [38] | S6 | 2.69 | 1.07 | 5.0 | 2.51 | 1.072 |
| T40 | Do | rada, and man [00] | 50 | 2.07 | 1.07 | 7.5 | 2.43 | 1.107 |
| T41 | | | | | | 10.0 | 2.36 | 1.140 |
| | | | | | | 10.0 | | 1.110 |

Sustainability **2021**, 13, 7737 16 of 21

 Table A1. Cont.

| Test ID | Dataset | Source | Soil ID | $G_{ m s}^{ m S}$ | $G_{ m s}^{ m T}$ | f _T (%) | $G_{ m s}^{ m ST}$ | $G_{\rm s}^{\rm S}/G_{\rm s}^{\rm ST}$ |
|---------------------------------|---------|--------------------------|---------|-------------------|-------------------|----------------------------------|--------------------------------------|--|
| T42 T43 T44 T45 T46 | D9 | Ravichandran et al. [39] | S7 | 2.65 | 0.85 | 0 5.0 10.0 15.0 20.0 | 2.65 2.41 2.22 2.08 1.96 | 1 1.100 1.194 1.274 1.352 |
| T47 T48 T49 T50 T51 | D10 | Ravichandran et al. [39] | S8 | 2.65 | 0.85 | 0 5.0 10.0 15.0 20.0 | 2.65 2.41 2.22 2.08 1.96 | 1 1.100 1.194 1.274 1.352 |
| T52 T53 T54 T55 T56 | D11 | Soltani et al. [34] | S9 | 2.69 | 1.09 | 0 5.0 10.0 20.0 30.0 | 2.69 2.51 2.37 2.16 2.01 | 1 1.072 1.135 1.245 1.338 |
| T57 T58 T59 T60 T61 | D12 | Soltani et al. [34] | S10 | 2.67 | 1.09 | 0 5.0 10.0 20.0 30.0 | 2.67 2.50 2.36 2.15 2.00 | 1 1.068 1.131 1.242 1.335 |
| T62 T63 T64 T65 T66 | D13 | Soltani et al. [34] | S11 | 2.71 | 1.09 | 0 5.0 10.0 20.0 30.0 | 2.71 2.53 2.39 2.17 2.02 | 1 1.071 1.134 1.249 1.342 |
| T67 T68 T69 T70 T71 | D14 | Soltani et al. [34] | S12 | 2.72 | 1.09 | 0 5.0 10.0 20.0 30.0 | 2.72 2.54 2.39 2.18 2.02 | 1 1.071 1.138 1.248 1.347 |
| T72 T73 T74 T75 T76 | D15 | Soltani et al. [35] | S13 | 2.73 | 1.09 | 0 5.0 10.0 20.0 30.0 | 2.73 2.55 2.40 2.18 2.03 | 1 1.071 1.138 1.252 1.345 |
| T72 T77 T78 T79 T80 | D16 | Soltani et al. [35] | S13 | 2.73 | 1.09 | 0 5.0 10.0 20.0 30.0 | 2.73 2.55 2.40 2.18 2.03 | 1 1.071 1.138 1.252 1.345 |
| T81 T82 T83 T84 T85 | D17 | Soltani et al. [23] | S14 | 2.76 | 1.09 | 0 5.0 10.0 20.0 30.0 | 2.76 2.57 2.42 2.20 2.04 | 1 1.074 1.140 1.255 1.353 |
| T86 T87 T88 T89 | D18 | Soltani et al. [19] | S15 | 2.77 | 1.08 | 0 5.0 10.0 20.0 | 2.77 2.58 2.43 2.20 | 1 1.074 1.140 1.259 |
| T86 T90 T91 T92 | D19 | Soltani et al. [19] | S15 | 2.77 | 1.10 | 0 5.0 10.0 20.0 | 2.77 2.58 2.43 2.21 | 1 1.074 1.140 1.253 |

Sustainability **2021**, 13, 7737 17 of 21

| T 1 1 | 4 4 | 0 1 |
|-------|-----|-------|
| Table | Δ Ι | (Out |
| | | |

| Test ID | Dataset | Source | Soil ID | $G_{ m s}^{ m S}$ | $G_{\mathrm{s}}^{\mathrm{T}}$ | f _T (%) | $G_{ m s}^{ m ST}$ | $G_{\rm s}^{\rm S}/G_{\rm s}^{\rm ST}$ |
|---------|---------|---------------------|---------|-------------------|-------------------------------|--------------------|--------------------|--|
| T86 | | | | | | 0 | 2.77 | 1 |
| T93 | D20 | Caltani at al [10] | C1 F | 2.77 | 2.77 1.11 | 5.0 | 2.59 | 1.069 |
| T94 | D20 | Soltani et al. [19] | S15 | | | 10.0 | 2.44 | 1.135 |
| T95 | | | | | | 20.0 | 2.22 | 1.248 |
| T96 | | | | | | 0 | 2.73 | 1 |
| T97 | | | | | | 5.0 | 2.55 | 1.071 |
| T98 | D21 | Soltani et al. [22] | S16 | 2.73 | 1.09 | 10.0 | 2.40 | 1.138 |
| T99 | | | | | | 20.0 | 2.18 | 1.252 |
| T100 | | | | | | 30.0 | 2.03 | 1.345 |

Note: f_T = TDA content (i.e., TDA-to-soil dry mass ratio, here expressed as a percentage value); G_s^S and G_s^T = specific gravity of soil solids and TDA particles, respectively; and G_s^{ST} = specific gravity of soil–TDA mixture (Equation (3)).

Figure A1 illustrates the variations of the conventional OMC and MDUW parameters, $w_{\rm opt}$ and $\gamma_{\rm dmax}$, against TDA content $f_{\rm T}$ for the 21 fine-grained soil–TDA compaction datasets.

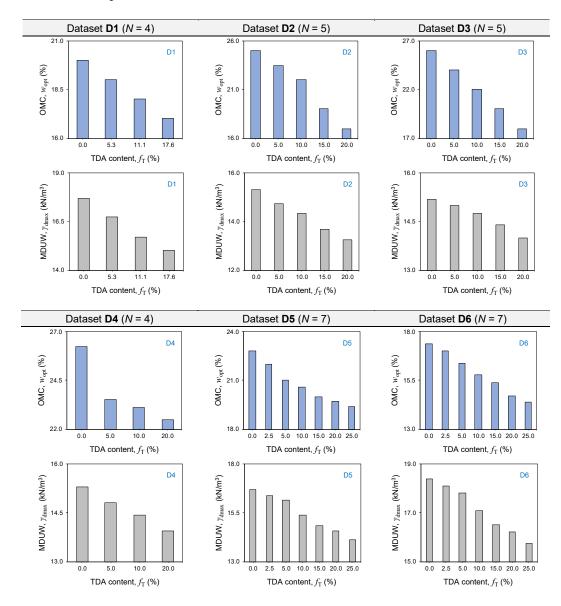


Figure A1. Cont.

Sustainability **2021**, 13, 7737 18 of 21

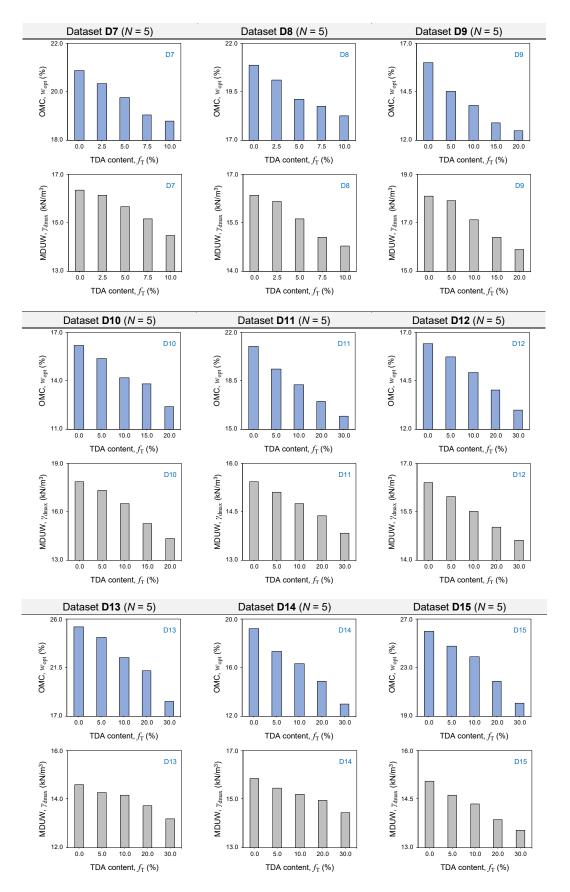


Figure A1. Cont.

Sustainability **2021**, 13, 7737 19 of 21

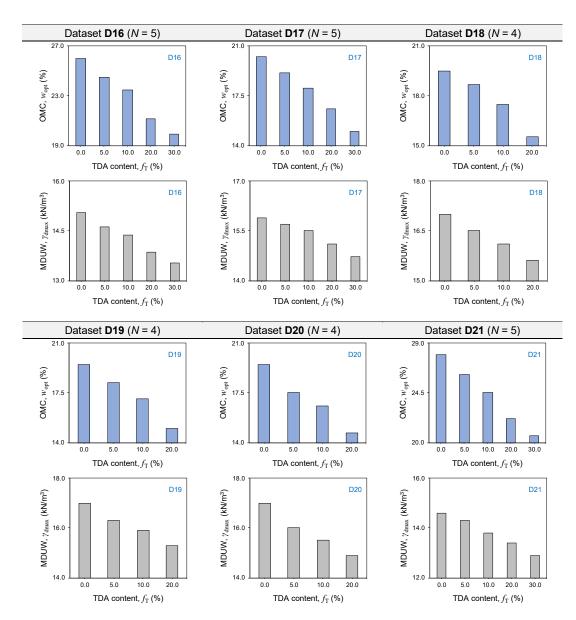


Figure A1. Variations of the conventional OMC and MDUW against TDA content for the 21 soil-TDA compaction datasets.

References

- 1. Humphrey, D.; Blumenthal, M. The use of tire-derived aggregate in road construction applications. In *Green Streets and Highways* 2010: An Interactive Conference on the State of the Art and How to Achieve Sustainable Outcomes; Weinstein, N., Ed.; ASCE: Reston, VA, USA, 2010; pp. 299–313. [CrossRef]
- 2. Reddy, S.B.; Krishna, A.M. Recycled tire chips mixed with sand as lightweight backfill material in retaining wall applications: An experimental investigation. *Int. J. Geosynth. Gr. Eng.* **2015**, *1*, 31. [CrossRef]
- 3. Li, L.; Xiao, H.; Ferreira, P.; Cui, X. Study of a small scale tyre-reinforced embankment. *Geotext. Geomembr.* **2016**, *44*, 201–208.
- 4. Khatami, H.; Deng, A.; Jaksa, M. Passive arching in rubberized sand backfills. Can. Geotech. J. 2020, 57, 549–567. [CrossRef]
- 5. Raeesi, R.; Soltani, A.; King, R.; Disfani, M.M. Field performance monitoring of waste tire-based permeable pavements. *Transp. Geotech.* **2020**, *24*, 100384. [CrossRef]
- 6. Shahrokhi-Shahraki, R.; Kwon, P.S.; Park, J.; O'Kelly, B.C.; Rezania, S. BTEX and heavy metals removal using pulverized waste tires in engineered fill materials. *Chemosphere* **2020**, 242, 125281. [CrossRef] [PubMed]
- 7. Edil, T.; Bosscher, P. Engineering properties of tire chips and soil mixtures. Geotech. Test. J. 1994, 17, 453–464. [CrossRef]
- 8. Foose, G.J.; Benson, C.H.; Bosscher, P.J. Sand reinforced with shredded waste tires. J. Geotech. Eng. 1996, 122, 760–767. [CrossRef]
- 9. Al-Tabbaa, A.; Blackwell, O.; Porter, S.A. An investigation into the geotechnical properties of soil–tyre mixtures. *Environ. Technol.* **1997**, *18*, 855–860. [CrossRef]

Sustainability **2021**, 13, 7737 20 of 21

10. Lee, J.H.; Salgado, R.; Bernal, A.; Lovell, C.W. Shredded tires and rubber–sand as lightweight backfill. *J. Geotech. Geoenviron. Eng.* **1999**, 125, 132–141. [CrossRef]

- 11. Zornberg, J.G.; Cabral, A.R.; Viratjandr, C. Behaviour of tire shred–sand mixtures. Can. Geotech. J. 2004, 41, 227–241. [CrossRef]
- 12. Lee, J.S.; Dodds, J.; Santamarina, J.C. Behavior of rigid-soft particle mixtures. J. Mater. Civ. Eng. 2007, 19, 179-184. [CrossRef]
- 13. Kim, H.K.; Santamarina, J.C. Sand-rubber mixtures (large rubber chips). Can. Geotech. J. 2008, 45, 1457–1466. [CrossRef]
- Mohammadinia, A.; Disfani, M.M.; Narsilio, G.A.; Aye, L. Mechanical behaviour and load bearing mechanism of high porosity permeable pavements utilizing recycled tire aggregates. Constr. Build. Mater. 2018, 168, 794

 –804. [CrossRef]
- 15. Cabalar, A.F.; Karabash, Z.; Mustafa, W.S. Stabilising a clay using tyre buffings and lime. *Road Mater. Pavement Des.* **2014**, *15*, 872–891. [CrossRef]
- 16. Signes, C.H.; Garzón-Roca, J.; Fernández, P.M.; de la Torre, M.E.G.; Franco, R.I. Swelling potential reduction of Spanish argillaceous marlstone Facies Tap soil through the addition of crumb rubber particles from scrap tyres. *Appl. Clay Sci.* **2016**, 132–133, 768–773. [CrossRef]
- 17. Yadav, J.S.; Tiwari, S.K. Effect of waste rubber fibres on the geotechnical properties of clay stabilized with cement. *Appl. Clay Sci.* **2017**, 149, 97–110. [CrossRef]
- 18. Bekhiti, M.; Trouzine, H.; Rabehi, M. Influence of waste tire rubber fibers on swelling behavior, unconfined compressive strength and ductility of cement stabilized bentonite clay soil. *Constr. Build. Mater.* **2019**, 208, 304–313. [CrossRef]
- 19. Soltani, A.; Taheri, A.; Deng, A.; O'Kelly, B.C. Improved geotechnical behavior of an expansive soil amended with tire-derived aggregates having different gradations. *Minerals* **2020**, *10*, 923. [CrossRef]
- 20. Akbarimehr, D.; Fakharian, K. Dynamic shear modulus and damping ratio of clay mixed with waste rubber using cyclic triaxial apparatus. *Soil Dyn. Earthq. Eng.* **2021**, *140*, 106435. [CrossRef]
- 21. Ghadr, S.; Samadzadeh, A.; Bahadori, H.; O'Kelly, B.C.; Assadi-Langroudi, A. Liquefaction resistance of silty sand with ground rubber additive. *Int. J. Geomech.* **2021**, 21, 04021076. [CrossRef]
- 22. Soltani, A.; Raeesi, R.; Taheri, A.; Deng, A.; Mirzababaei, M. Improved shear strength performance of compacted rubberized clays treated with sodium alginate biopolymer. *Polymers* **2021**, *13*, 764. [CrossRef]
- 23. Soltani, A.; Deng, A.; Taheri, A.; O'Kelly, B.C. Engineering reactive clay systems by ground rubber replacement and polyacry-lamide treatment. *Polymers* **2019**, *11*, 1675. [CrossRef]
- 24. Yadav, J.S.; Tiwari, S.K. The impact of end-of-life tires on the mechanical properties of fine-grained soil: A review. *Environ. Dev. Sustain.* **2019**, *21*, 485–568. [CrossRef]
- 25. Yang, Z.; Zhang, Q.; Shi, W.; Lv, J.; Lu, Z.; Ling, X. Advances in properties of rubber reinforced soil. *Adv. Civ. Eng.* 2020, 2020, 6629757. [CrossRef]
- 26. Mistry, M.K.; Shukla, S.J.; Solanki, C.H. Reuse of waste tyre products as a soil reinforcing material: A critical review. *Environ. Sci. Pollut. Res.* **2021**, *28*, 24940–24971. [CrossRef]
- 27. Gurtug, Y.; Sridharan, A. Compaction behaviour and prediction of its characteristics of fine grained soils with particular reference to compaction energy. *Soils Found.* **2004**, *44*, 27–36. [CrossRef]
- 28. Horpibulsuk, S.; Katkan, W.; Apichatvullop, A. An approach for assessment of compaction curves of fine grained soils at various energies using a one point test. *Soils Found.* **2008**, *48*, 115–125. [CrossRef]
- 29. Sivrikaya, O.; Togrol, E.; Kayadelen, C. Estimating compaction behavior of fine-grained soils based on compaction energy. *Can. Geotech. J.* **2008**, 45, 877–887. [CrossRef]
- 30. Pillai, G.A.S.; Vinod, P.P. Re-examination of compaction parameters of fine-grained soils. *Proc. Inst. Civ. Eng. Gr. Improv.* **2016**, 169, 157–166. [CrossRef]
- 31. Spagnoli, G.; Shimobe, S. An overview on the compaction characteristics of soils by laboratory tests. *Eng. Geol.* **2020**, 278, 105830. [CrossRef]
- 32. Di Matteo, L.; Spagnoli, G. Predicting compaction properties of soils at different compaction efforts. *Proc. Inst. Civ. Eng. Geotech. Eng.* **2021**, in press. [CrossRef]
- 33. Shivaprakash, S.H.; Sridharan, A. Correlation of compaction characteristics of standard and reduced Proctor tests. *Proc. Inst. Civ. Eng. Geotech. Eng.* **2021**, *174*, 170–180. [CrossRef]
- 34. Soltani, A.; Deng, A.; Taheri, A.; Sridharan, A. Consistency limits and compaction characteristics of clay soils containing rubber waste. *Proc. Inst. Civ. Eng. Geotech. Eng.* **2019**, 172, 174–188. [CrossRef]
- 35. Soltani, A.; Deng, A.; Taheri, A.; Sridharan, A. Swell–shrink–consolidation behavior of rubber-reinforced expansive soils. *Geotech. Test. J.* **2019**, 42, 761–788. [CrossRef]
- 36. Prasad, A.S.; Ravichandran, P.T.; Annadurai, R.; Rajkumar, P.R.K. Study on effect of crumb rubber on behavior of soil. *Int. J. Geomatics Geosci.* **2014**, *4*, 579–584.
- 37. Ramirez, G.G.D.; Casagrande, M.D.T.; Folle, D.; Pereira, A.; Paulon, V.A. Behavior of granular rubber waste tire reinforced soil for application in geosynthetic reinforced soil wall. *Rev. IBRACON Estrut. Mater.* **2015**, *8*, 567–576. [CrossRef]
- 38. Yadav, J.S.; Tiwari, S.K. A study on the potential utilization of crumb rubber in cement treated soft clay. *J. Build. Eng.* **2017**, *9*, 177–191. [CrossRef]

Sustainability **2021**, 13, 7737 21 of 21

39. Ravichandran, P.T.; Priyanga, G.; Krishnan, K.D.; Rajkumar, P.R.K. Compaction behavior of rubberized soil. In *Proceedings of the International Conference on Intelligent Computing and Applications—Advances in Intelligent Systems and Computing, Velammal Engineering College, Chennai, India, 2–3 February 2018*; Bhaskar, M., Dash, S., Das, S., Panigrahi, B., Eds.; Springer: Singapore, 2019; Volume 846, pp. 129–134. [CrossRef]

- 40. BS 5930. Code of Practice for Ground Investigations; British Standards Institution (BSI): London, UK, 2015; ISBN 9780539081350.
- 41. Özkul, Z.H.; Baykal, G. The influence of energy level on the effectiveness of compaction of clays with rubber fiber added. In *Proceedings of the Transportation Research Board 86th Annual Meeting, Washington, DC, USA, 21–25 January 2007*; Transportation Research Board (TRB): Washington, DC, USA, 2007; Paper No. 07-1685.
- 42. Al-Tabbaa, A.; Aravinthan, T. Natural clay–shredded tire mixtures as landfill barrier materials. *Waste Manag.* **1998**, *18*, 9–16. [CrossRef]
- 43. Priyadarshee, A.; Gupta, D.; Kumar, V.; Sharma, V. Comparative study on performance of tire crumbles with fly ash and kaolin clay. *Int. J. Geosynth. Gr. Eng.* **2015**, *1*, 38. [CrossRef]
- 44. Akbarimehr, D.; Eslami, A.; Aflaki, E. Geotechnical behaviour of clay soil mixed with rubber waste. *J. Clean. Prod.* **2020**, 271, 122632. [CrossRef]
- 45. Trouzine, H.; Bekhiti, M.; Asroun, A. Effects of scrap tyre rubber fibre on swelling behaviour of two clayey soils in Algeria. *Geosynth. Int.* **2012**, *19*, 124–132. [CrossRef]
- 46. Soltani, A.; O'Kelly, B.C. Discussion of "The flow index of clays and its relationship with some basic geotechnical properties" by G. Spagnoli, M. Feinendegen, L. Di Matteo, and D. A. Rubinos, published in Geotechnical Testing Journal 42, no. 6 (2019): 1685–1700. *Geotech. Test. J.* **2021**, 44, 216–219. [CrossRef]
- 47. Soltani, A.; O'Kelly, B.C. Reappraisal of the ASTM/AASHTO standard rolling device method for plastic limit determination of fine-grained soils. *Geosciences* **2021**, *11*, 247. [CrossRef]
- 48. Bland, J.M.; Altman, D.G. Measuring agreement in method comparison studies. *Stat. Methods Med. Res.* **1999**, *8*, 135–160. [CrossRef]