



## Article

# Bird-Borne Samplers for Monitoring CO<sub>2</sub> and Atmospheric Physical Parameters

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**Abstract:** Air quality monitoring in cities is significant for both human health and environment. Here, an innovative miniaturized active air sampler wearable by free-flying birds is presented. The device integrates a GPS logger and atmospheric calibrated sensors allowing for high spatiotemporal resolution measurements of carbon dioxide (CO<sub>2</sub>) concentration, barometric pressure, air temperature, and relative humidity. A field campaign, carried out from January to June 2021, involved the repeated release of homing pigeons (*Columba livia*) from downtown Rome (Italy), to sample the air on their way back to the loft, located in a rural area out of the city. The measurements suggest the importance of green urban areas in decreasing CO<sub>2</sub> levels. Moreover, a positive relation between CO<sub>2</sub> levels, relative humidity, and air temperature was revealed. In contrast, a negative relation with distance from the point of release, month, and time of day was found. Flight speed and the altitude of flight were related to rising CO<sub>2</sub> levels. The easy use of such devices paves the way for the application of miniaturized air samplers to other synanthropic species (i.e., gulls), making birds convenient biomonitors for the urban environment.

**Keywords:** urban pollution; air quality; atmospheric monitoring; urban boundary layer; active air sampler; carbon dioxide concentration; homing pigeons



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## 1. Introduction

During the last decades, large rural regions have been converted into urbanized areas, with an increase in traffic volume and a deterioration in air quality. An ever-increasing fraction of the global population resides in cities, and it is estimated that, in 2050, this percentage will reach the record level of 68.4% [1].

The continuous building expansion leads to an escalation in the energy demand of cities, which contributed to 80% of the global primary energy demand [2]. Moreover, among greenhouse gases, carbon dioxide (CO<sub>2</sub>) is the most worrying, as it represents about 80% of the total emissions [3]. In 2019, atmospheric CO<sub>2</sub> levels were higher than at any time in at least two million years, with an increase of 47% since 1750 [4]. The significant growth in concentration, along with other greenhouse gases, has increased global average temperatures in the first two decades of 21st century by  $0.99 \pm 0.15$  °C [4], also contributing to severe climatic events [5,6].

Although CO<sub>2</sub> is one of the main gases responsible for climate change, surface/atmosphere fluxes are generally evaluated only above vegetative canopies (e.g., see the EUROFLUX [7] and AMERIFLUX [8] projects). CO<sub>2</sub> emissions can be indirectly estimated from emissions inventories, which are rarely validated in the presence of traffic and domestic heating, while the direct measurements of urban CO<sub>2</sub> concentration are still very rare. For example, Morikawi and Kanda [9] used micrometeorological sensors to assess the diurnal and seasonal variability of CO<sub>2</sub> in a suburban area of Tokyo (Japan). They found a daily trend closely related to anthropogenic emissions, and peaks associated with morning and

evening rush hours traffic and population density. Seasonal CO<sub>2</sub> fluxes show a minimum in summer, related to the greater absorption of vegetation, and a maximum during wintertime because of the increased fossil fuel consumption [9,10]). Although a few other studies have focused on CO<sub>2</sub> levels in cities [11–13], a lack of knowledge regarding the temporal and spatial variability of CO<sub>2</sub> concentration in urban environments remains. In particular, the campaigns carried out so far have considered in situ measurements at the ground level, not allowing for the investigation of the Planetary Boundary Layer (PBL), especially in built-up areas, where local scale topographic and meteorological features (e.g., sea/land breeze, valleys, plateaus) might produce variable conditions.

In this context, the atmospheric monitoring through small, dedicated air samplers, wearable by birds freely moving in the urban environment represents an innovative technique. In fact, the use of birds could be convenient, especially in urban areas, where numerous limitations must be observed (e.g., flight of drones and release of atmospheric probes are restricted) and where most of the anthropogenic CO<sub>2</sub> sources are located.

The development of miniaturized sensors is pushing the frontiers of animal ecology through biologging. Biologging refers to the use of devices (biologgers) attached to animals, that collect data about the wearers' movement, behaviour, physiology, and/or environment [14]. The use of wild animals to measure the state of the environment has been a topic of interest since the first International Biologging Symposium in 2003 [15]. Marine animals were used as 'oceanographers' in areas not easily reachable by standard monitoring systems. Equipped with wearable global positioning systems (GPS) loggers integrated with environmental sensors, the animals could measure the chemical and physical parameters of the water they moved in [16]. For many years, this application has mostly been limited to marine biology, primarily due to size constraints. Since then, technological improvements have been producing increasingly frequent calls to use terrestrial animals to measure environmental parameters [17,18].

The main objective of this paper was to introduce and test an innovative miniaturized set of sensors, integrated with a small GPS data logger for deployment on homing pigeons (*Columba livia*) and other birds. So far, the application of GPS loggers on birds has been used to collect qualitative information on the development of thermals in relation to orography and winds in soaring vultures [19], on wind intensity and flight direction in seabirds [20], and for the quantitative study of atmospheric variables [21].

Domestic pigeons, selected for their homing ability, have been used throughout human history to carry messages [22], medication, and even to smuggle drugs [23]. Famously, during the 20th century, an aerial photography technique based on pigeons carrying lightweight miniature cameras was invented by Julius Neubronner [24]. Homing pigeons have been at the forefront of biologging, with the first GPS tracking studies published in the early 2000s [25,26] and even neurophysiology studies using EEG-equipped GPS tags [27].

To the best of our knowledge, to this day, two attempts have been made to use homing pigeons as urban environmental monitors but neither yielded reproducible prototypes or any scientific output. The first one is "Pigeonblog", an artistic and political endeavour by Beatriz da Costa [28]. Da Costa, inspired by Neubronner's aerial photography and by the early scientific literature on pigeon tracks, collaborated with engineers and pigeon fanciers to develop a GPS unit with sensors for monitoring levels of carbon monoxide and nitrogen oxides. Within the project, three pigeon releases were carried out with data accessible during the project from a dedicated website. The other project, named "PigeonAirPatrol" [29], aimed to use feral pigeons to investigate air pollution. The project made headlines [30] and raised public awareness around the issue of urban air quality monitoring in London, where the implementation took place.

Here, we present a calibrated atmospheric sensor integrated on a GPS logger wearable by homing pigeons. The pigeons, repeatedly released in the urban area of Rome (Italy), recorded atmospheric pressure, air temperature, humidity, and CO<sub>2</sub> concentration at high spatial and temporal resolution. In particular, on their return flight to the home loft, the birds flew above buildings, large urban parks, and crops, allowing for the measurement

and the comparison of atmospheric variables within the PBL and in conditions typically difficult to investigate when employing fixed, ground-based instruments.

## 2. Materials and Methods

### 2.1. Development and Design of the Air Sampler

The development of the miniaturized air samplers used in the present study is carried out in collaboration with a private company specialized in devices for animal tracking (Technosmart Europe S.r.l., Rome, Italy). The device is based on the integration of a set of sensors on an existing GPS data logger with a wire antenna powered by a 200 mA LIPO battery (AxyTrek).

The boards and the battery are arranged in a flat and aerodynamic design (50 mm × 20 mm) to reduce possible drag to the birds in flight. The weight of the complete system (Figure 1) is 14.6 g, including the battery, and therefore at the limit of the recommended 3% of the bird's body mass, considering that the pigeons weighed around 450 g [31]. In any case, the units are used for short-term deployments of 1–2 h.



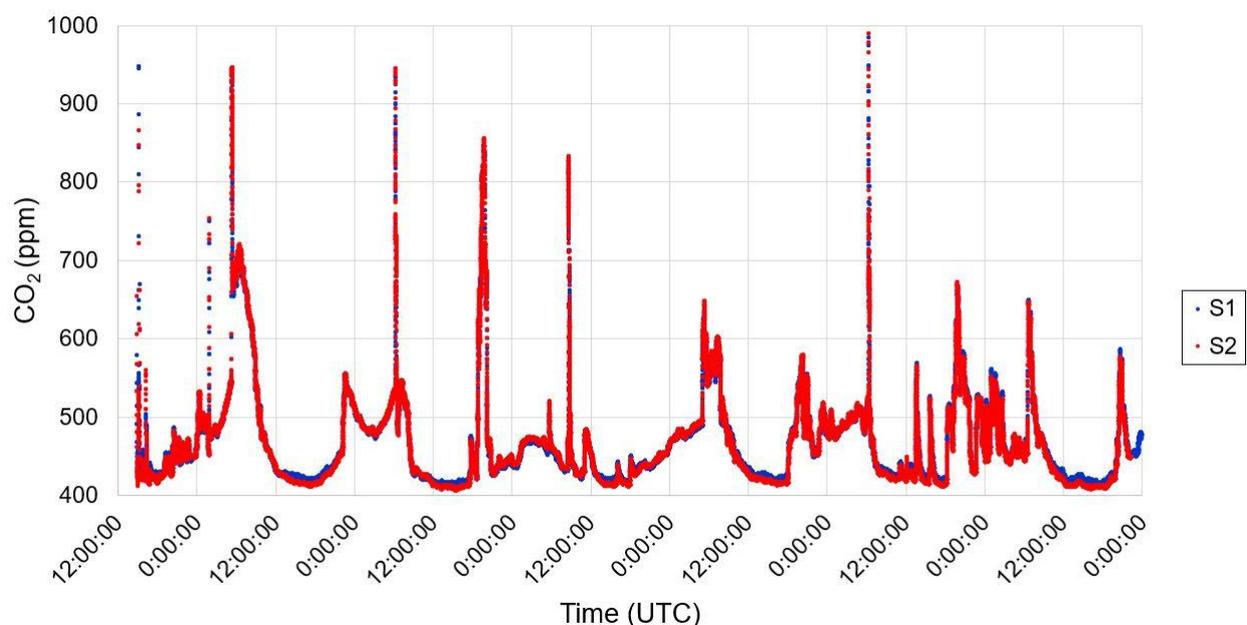
**Figure 1.** Photograph of the device attached on a homing pigeon (a), schematic view of printed circuit board with CO<sub>2</sub> and atmospheric sensors (b), CO<sub>2</sub> sensor (Senseair Sunrise, Senseair AB, Delsbo, Sweden) with the gas diffusion area represented by the white membrane (c), and AxyTrek data logger (d).

The atmospheric sensors (BME280 Environmental sensor, Bosch Sensortec, Reutlingen, Germany) measure temperature (operational range:  $-40$ – $85$  °C, absolute accuracy:  $\pm 1$  °C, resolution:  $0.01$  °C), barometric pressure (operational range:  $300$ – $1100$  hPa; absolute accuracy:  $\pm 1$  hPa, resolution:  $0.18$  hPa), and relative humidity (operational range:  $0$ – $100\%$ , absolute accuracy:  $\pm 3\%$ , resolution:  $0.008\%$ ). The CO<sub>2</sub> sensor (article n. 006-0-0007, Senseair Sunrise, Senseair AB, Delsbo, Sweden) is based on non-dispersive infrared (NDIR) technology and could measure CO<sub>2</sub> concentrations from  $400$  to  $5000$  ppm, with an accuracy of  $\pm 30$  ppm +  $3\%$  of reading (operating range temperature:  $0$ – $50$  °C, operating range relative humidity:  $0$ – $85\%$ ). The CO<sub>2</sub> sensor measures the light absorption emitted by a light-emitting diode (LED) into a dark chamber employing a photodiode (Figure 1). The number of CO<sub>2</sub> particles contained in the airflow is related to the light intensity detected by the photodiode at a specific wavelength, which for CO<sub>2</sub> is  $4.26$   $\mu\text{m}$ . The CO<sub>2</sub> sensor measures  $33.5(\text{L}) \times 19.7(\text{W}) \times 11.5(\text{H})$  mm<sup>3</sup> and weighs  $5$  g.

## 2.2. CO<sub>2</sub> Calibration and Configuration of Environmental Sensors

A variety of calibration options are made available for the CO<sub>2</sub> sensors by the producer, such as the Automatic Baseline Correction (ABC) algorithm [32] and manual calibration. The former works in the background over  $180$  h cycles, provided the sensor is exposed to “fresh air” (a customizable baseline concentration value,  $400$  ppm by default) at least once during the cycle. For each cycle, the sensor stores the lowest value recorded, which is assumed as the “fresh air” reference to calculate a correction factor for the data, ensuring data is reliable in the long run. The latter requires the use of a reference gas mixture and does not correct drift in long-term acquisitions.

For this study, being the data collected only for short periods during homing flights, we ensured different tags have similar sensitivity. Therefore, CO<sub>2</sub> sensors were placed indoors for a week-long test, close to an open window, in a room occasionally occupied, increasing CO<sub>2</sub> levels significantly. As shown in Figure 2, the two sensors tested show very similar sensitivity, with comparable responses to variations in CO<sub>2</sub> concentration in the test room. Granger tests were performed over all the combinations of sensors, testing whether one-time series predicted the other and vice versa. All tests were highly significant ( $p < 0.001$ ), meaning that every timeseries is predictable by the other.



**Figure 2.** Temporal trend of CO<sub>2</sub> concentration during a week-long test to compare the sensitivity of two sensors (different colours).

CO<sub>2</sub> concentrations were acquired with a sampling frequency of 0.42 Hz. This is the default value for the Senseair Sunrise, which collects eight samples per measurement. A sample takes less than 300 ms, producing a period of 2.4 s for each complete measurement. The response time is reported to be less than 30 s. This is the time needed for the sensor to read 90% (namely, T90%) of the true gas concentration, in an enclosure that changes from 8500 ppm to 400 ppm and the opposite with a gas flow rate of 1 L/min. Details about the ABC algorithm, pressure dependence, and measurement period can be found in the sensor user manual [33].

Atmospheric pressure has been found to produce a 1.6% change in CO<sub>2</sub> readings for each 10 hPa deviation from a mean sea-level pressure (MSLP) of 1013.25 hPa. During the field campaign, the atmospheric pressure varied between 986.0 hPa and 1021.0 hPa, producing a maximum deviation of 26.9 hPa from the MSLP. This corresponds to a maximum error of 4.3% in the CO<sub>2</sub> readings. Only 2.3% of data used in the present study have pressure with a deviation >10 hPa from the MSLP and CO<sub>2</sub> values exceeding 625 ppm, being therefore concerned by an error greater than 10 ppm.

Temperature, air pressure, and relative humidity were collected at 1 Hz. The response time for the humidity sensor is 1 s, calculated as the time needed for the sensor to reach 63% of the final value (namely, T63%) when going from 90% to 0% or vice versa [34]. During the field campaign, the maximum daily variation in relative humidity was 49%.

### 2.3. Study Area and Sample Collection

The experimental campaign, carried out from January to June 2021, involved the repeated release of homing pigeons equipped with the air sampler data logger near the campus of the University of Rome “La Sapienza” (41.90°N, 12.51°E) in downtown Rome (Italy), i.e., in a highly urbanized and moderately polluted area. The loft (41.58°N, 12.37°E) was located about 13 km northeast of the release site, about 3 km out of the “Grande Raccordo Anulare”, the highly trafficked ring road that encircles Rome.

During their flight back to the loft, homing pigeons passed through areas with different degrees of urbanization, land use, and pollution levels, i.e., they were expected to fly across a gradient of decreasing CO<sub>2</sub> concentration to reach their loft in the countryside. Moreover, sampling was expected to take place in the range of altitudes between 0 and 150 m above ground level (m a.g.l.), the typical flight range of pigeons. All the releases were carried out on working days and most of the releases took place early in the morning (at about 07:00 UTC) to capture the increase in CO<sub>2</sub> concentration associated with morning traffic rush hours. A few pigeons were released at later hours (from 09:00 UTC up to 13:00 UTC). A list of the release trials, together with details about distance flown, duration of flights, and statistics on measured CO<sub>2</sub> is shown in Table A1.

For the trials, twelve adult homing pigeons were equipped with the devices following the procedure described for GPS loggers by [35]. Briefly, the birds were habituated to carry the load using a plastic dummy (of the same weight and size as the logger) attached with a Velcro strip on their back, on a dorsal area between the wings. The hard side of the Velcro strip (30.0 × 20.0 mm) was attached using a neoprenic glue on an area of half-cut feathers—therefore without causing pain and discomfort to the birds—whereas the soft side of Velcro was attached at the base of both the dummy and the device. Pigeons carried the dummy for two weeks before the experimental release. In general, they already resumed their normal behaviour (feeding, daily flight, reproductive activities) on the day after the attachment of the dummy. On the day of the release, birds were taken from the loft and gently placed in a wooden box for transportation to the city centre by car (about a 30 min trip). Ten minutes before the release, the dummies were replaced with the air samplers, and then, pigeons were released in flocks of 3–4 birds with 1–2 birds equipped with the air sampler and the remaining wearing the dummies. Releases were carried out only with good meteorological conditions, avoiding rainy and windy days. On their arrival at the loft, birds were handled to recover the loggers and reposition the dummy. The data were

downloaded using dedicated software to obtain a CSV file containing timeseries of GPS position, temperature, air pressure, humidity, and CO<sub>2</sub> concentration.

#### 2.4. Data Processing

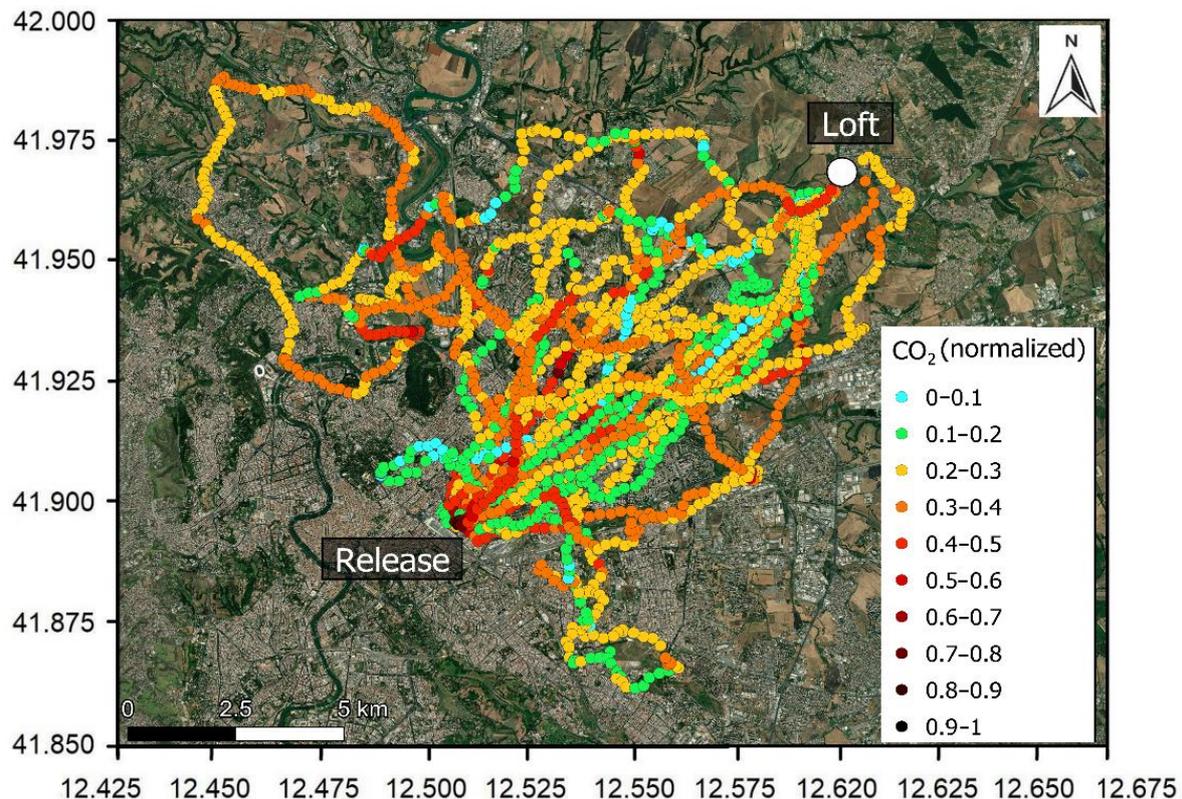
Firstly, data collected by the various sensors were visually inspected to ascertain whether the tracks were complete. The GPS locations were plotted on the city map using QGIS software (version 3.10), official project of the Open Source Geospatial Foundation (OSGeo, Beaverton, OR, USA), and the symbols corresponding to the GPS positions were colour-coded along a gradient according to the CO<sub>2</sub> level for each of the sensors used allowing for a first data inspection.

GPS data were pre-processed following the procedure proposed by [36]. In particular, GPS values related to velocities greater than 90 km/h and negative altitudes were discarded. Moreover, only active flight points (flight speed as measured by GPS  $\geq 0.1$  m/s) were considered for the analysis and pigeon tracks were cut on their arrival by excluding points in a circular area of 500 m centred on the loft. Pigeon tracks were merged with level 3 Corine Land Cover (CLC) data [37] and classified as “urban fabric” (CLC classes from 1.1.1 to 1.3.3), “green urban areas” (CLC classes 1.4.1 and 1.4.2), and “agricultural areas” (CLC classes from 2.1.1 to 2.4.4). Points along the Tiber River (CLC class 5.1.1) were assimilated into green urban areas. No other land cover classes were crossed during the flights. It is worth noticing that CLC classification probably does not offer the optimal resolution for land cover features. Nonetheless, to the best of our knowledge, this is the most recent and accurate dataset available for the area under investigation. To improve the representation of the examined area, a shapefile of roads [38] was used to classify points lying within 50 m of streets, considering both motorways and small urban roads.

Then, CO<sub>2</sub> values were regressed against land cover, intersection with streets, and time of release (i.e., the hour of day and month of the year). Whether to include in the analysis temperature, humidity, barometric pressure, flight speed, flight altitude, distance from the release point, and interaction between the latter two variables was decided via Akaike Information Criterion (AIC) [39] using stepwise elimination (function “buildmer” from package “buildmer” [40]). Multicollinearity was checked by calculating the variance inflation factor (VIF, function “vif” from package “car” [41]). Distance from the release point was included due to an increasing presence of agricultural areas moving out of the city centre towards the loft. From preliminary inspection (see colours in Figure 3), the distance from the release point seemed to be related to CO<sub>2</sub> values and, therefore, it was considered as a continuous variable, providing information on land cover. A linear mixed effect model (function “lme” from package “nlme” [42]) with pigeon ID as a random effect was run, testing the random effects by comparing the model to a generalized least squares regression only containing the fixed predictors.

Data were highly autocorrelated, and while this did not affect regression coefficient estimates, it might have produced biased standard errors, making the coefficient significance unreliable [43]. To account for this, data were first subsampled with 15 s temporal resolution (corresponding to approximately 200 m spatial resolution if birds were flying with an average velocity of about 13 m/s resulting from our GPS data). Then, the model was fitted with a continuous autoregressive structure of the first order (function “corCAR1” from package “nlme” [42]). All analyses were performed using R Statistical Software (version 4.0.3, [44]), developed by the R Foundation for Statistical Computing (Vienna, Austria).

Throughout the whole campaign, the atmospheric sensors collected about  $1.9 \times 10^4$  data points of active flight outside the 500 m loft buffer during 25 flights. Further subsampling to one fix per 15 s resulted in about  $3.3 \times 10^3$  points with CO<sub>2</sub> measurements, which constitute the filtered dataset used for the following analysis.



**Figure 3.** Map of data subsample used for the regression model (i.e., one GPS value every 15 s). The map shows the tracks of pigeons released on different days. The colour intensity depicts the CO<sub>2</sub> concentration measured after normalization to allow for better comparison between flights.

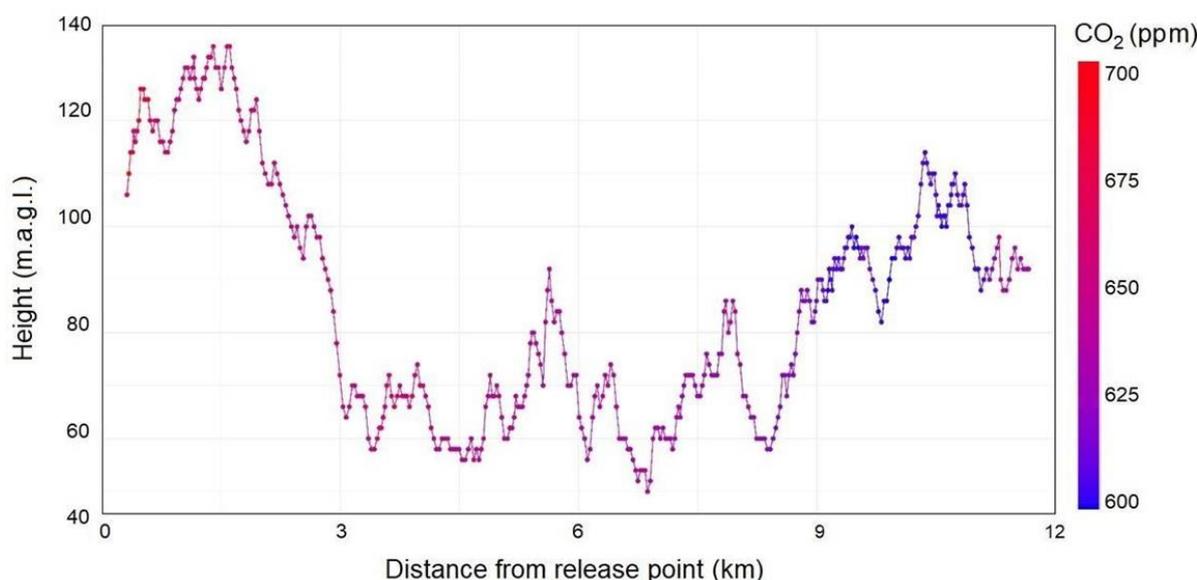
### 3. Results

Figure 3 shows the paths followed by the homing pigeons during the measurement campaign and the normalized CO<sub>2</sub> concentrations. In accordance with the details in Table A1, CO<sub>2</sub> levels range between 410 ppm and 993 ppm. The variation in concentrations between different flights are mainly due to different land use and flight time. In fact, the CO<sub>2</sub> peaks correspond to the measurements carried out close to major roads and during rush hour traffic. Clearly, during each flight, each pigeon can choose a different path to follow based on its experience and environmental conditions.

In Figure 4, an example of the CO<sub>2</sub> concentration time series collected during a pigeon flight expressed as a function of the distance from the release point and of the height is given. The concentration decreases with increasing flight altitude, as expected, moving away from surface emissions, but it is strongly influenced by land use, e.g., the decrease observable at about 10 km from the release point between 80 and 100 m.a.g.l. is due to the overflight of the agricultural landscape after the “Grande Raccordo Anulare” road, characterized by few roads and buildings.

The stepwise regression reveals the full model, i.e., the one comprising all terms (see Table 1), to have the lowest AIC score. VIF is below 2.7 for all the predictors, meaning only a low correlation is found among them. Including random effects, the regression lowered the model’s AIC by 116 points (Anova test,  $p < 0.001$ ), and the corCAR1 error structure lowered it by 3914 points (Anova test,  $p < 0.001$ ). The full model’s conditional coefficient of determination (hereinafter,  $R^2$ ) was 0.69, while the marginal  $R^2$  is 0.39 [45]. Our data shows that CO<sub>2</sub> concentrations are positively related to relative humidity (estimated: 40.71, standard error: 4.07,  $p < 0.001$ ) and air temperature (estimated: 12.30, standard error: 5.61,  $p = 0.03$ ) and negatively related to barometric pressure (estimates:  $-8.98$ , standard error: 5.61,  $p = 0.001$ ). Regarding the temporal trend, a negative relation with both months of the

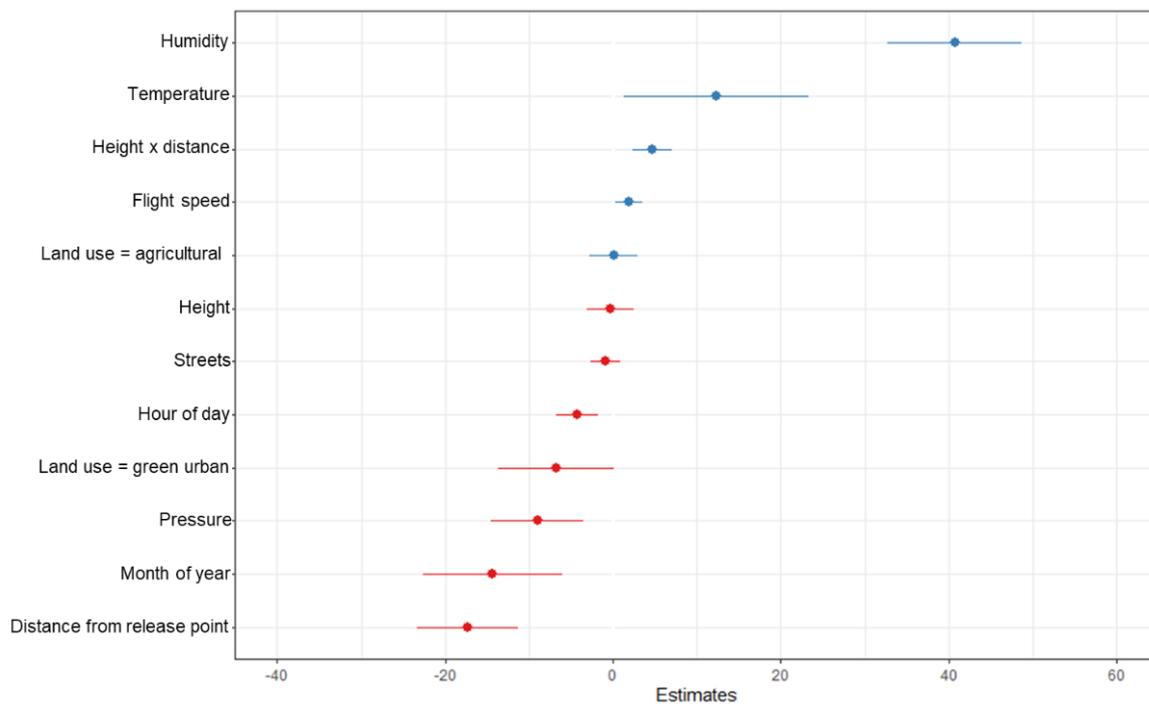
year (estimated:  $-14.34$ , standard error:  $4.24$ ,  $p < 0.001$ ) and time of the day (estimated:  $-4.32$ , standard error:  $1.27$ ,  $p < 0.001$ ) emerged. No evidence that the sections of flight over streets measured higher CO<sub>2</sub> levels than average was found. In some of the tracks, the overflight of the “Grande Raccordo Anulare” road determined the sudden and significant increase in the concentration of CO<sub>2</sub> but, in general, the dense urban road infrastructure did not reveal fine-scale differences within the urban environment itself (see Figure 3). However, lower CO<sub>2</sub> concentrations over urban green areas compared to pure urban fabric (estimated:  $-6.76$ , standard error:  $3.55$ ,  $p = 0.057$ ) were found. The interaction term between distance and height showed a positive relation with CO<sub>2</sub> (estimated:  $4.70$ , standard error:  $1.18$ ,  $p < 0.001$ ), unveiling a complex three-dimensional spatial pattern of diffusion. In fact, while for low heights lower CO<sub>2</sub> values moving outside of the city were observed, the relationship reversed for greater heights. It is worth noting the small but positive effect of flight speed on CO<sub>2</sub> (estimated:  $1.87$ , standard error:  $0.82$ ,  $p = 0.02$ ), which would need to be explained in a larger atmospheric dynamics context. A graphical summary of the model estimates is presented in Figure 5, while the full summary is given in Table 1.



**Figure 4.** Example of height profile of pigeon track (flight n. 16 in Table A1) with respect to distance from the release point. Colours refer to the CO<sub>2</sub> concentrations measured, which are higher (redder colours) close to the release point in the city centre, and lower (bluer colours) furthest from it.

**Table 1.** Summary of best linear mixed effects model. For each predictor, the coefficient estimate, standard error, and  $p$ -value are reported.  $p$ -values below 0.05 are reported in bold.

Predictor	Estimate	Standard Error	$p$ -Value
Intercept	645.37	26.25	<b>&lt;0.001</b>
Temperature	12.30	5.61	<b>0.029</b>
Pressure	$-8.98$	2.81	<b>0.002</b>
Relative humidity	40.71	4.07	<b>&lt;0.001</b>
Flight speed	1.88	0.82	<b>0.021</b>
Height	$-0.32$	1.43	0.822
Distance from release point	$-17.33$	3.09	<b>&lt;0.001</b>
Hour of day	$-4.33$	1.27	<b>&lt;0.001</b>
Month of year	$-14.34$	4.24	<b>&lt;0.001</b>
Over streets = true	$-0.86$	0.89	0.337
Land use = agricultural	0.07	1.49	0.961
Land use = green urban	$-6.76$	3.55	0.057
Height $\times$ distance	4.70	1.18	<b>&lt;0.001</b>



**Figure 5.** Graphical summary of the linear mixed effects regression model estimates. Dots are the coefficient estimates for each predictor (reported on the left), with lines representing the standard errors. Red and blue show negative and positive values, respectively.

#### 4. Discussion

The atmospheric CO<sub>2</sub> concentrations measured along the return journey of the pigeons released in the centre of Rome (see for example Figure 4) showed an evident negative gradient with increasing distance from the release point, as supported by the linear mixed effects regression model. The highest concentrations were recorded in downtown Rome, i.e., close to the release point, with a gradual decrease moving towards suburban and rural areas. This is consistent with the findings of other studies that have shown the presence of an “urban CO<sub>2</sub> dome” [46] and a close relationship between CO<sub>2</sub> levels and population density, in turn associated with the high traffic volume of urban centers [47]. The average concentration obtained considering only the samples measured over the “urban fabric” land cover class was 563 ppm, with average levels decreasing over “agricultural areas” (546 ppm) and even more so over “green urban areas” (538 ppm), highlighting the positive effect of urban greening on air quality [48]. All the average concentrations are well above the mean global atmospheric CO<sub>2</sub> concentration provided by [49] and referred to 2020, i.e., 412.5 ppm. In the urban environment, the concentration is comparable with results from [11], who carried out measurements in a highly urbanized and moderately polluted area in Rome, close to that investigated in this work, during the traffic rush hours. As expected, the values measured here for the “urban fabric” were slightly lower than the findings by [11], who carried out measurements along the road at the pedestrian level. A similar trend was also identified by [50], who examined the spatial variability of the near-surface CO<sub>2</sub> concentration in Shanghai (China). In the same time interval (from 9:00 to 11:00 AM) and the same season (spring), Liu et al. [50] identified a clear concentration decrease moving from the transportation area to crops, with a positive correlation with the percentage of impervious surface cover and a negative correlation with the percentage of vegetation cover. The positive relation between CO<sub>2</sub> level, air temperature, and relative humidity agrees with [11], who measured the highest CO<sub>2</sub> values with high temperatures and low wind speeds, and with [51], who identified a slight influence of temperature and relative humidity on the CO<sub>2</sub> concentration. Furthermore, the positive relation between

CO<sub>2</sub> levels and near-surface air temperatures was also highlighted by [51] at the remote sites of Mauna Loa (USA) and Point Barrow (USA). The slightly negative relationship between CO<sub>2</sub> and time of day, considering that most releases were carried out before 09:30 UTC, suggests that the convective mixing generated by the presence of solar radiation and the photosynthetic absorption of CO<sub>2</sub> by vegetation have a significant effect on the space-time redistribution of CO<sub>2</sub>, which, therefore, accumulates more in the layers of the atmosphere closest to the ground during the night and the early morning hours [46]. Moreover, this negative relation is presumably traceable also to factors we were unable to include in our model, such as the traffic rate, which is lower later in the morning after rush hours and shows a shift forward going towards the summer months. Of course, later morning hours are also associated with warmer air temperatures, therefore missing explanatory variables would confound the relationship between CO<sub>2</sub> and temperature, which we would expect to be significantly positive. To be able to verify this relation, it would be necessary to design measurement campaigns with continuous releases during the same day, allowing for the in-depth investigation of the PBL development and the photosynthetic absorption.

The negative relation with the month of the year is likely due to the CO<sub>2</sub> seasonal fluxes between the atmosphere and the land biosphere, which overlap with fossil fuel emissions, giving rise to large carbon dioxide seasonal variations [52].

Finally, the model shows a poor relation between CO<sub>2</sub> atmospheric concentration, flight speed, and altitude. This could be justifiable considering that the flight altitudes were mostly below 150 m a.g.l., i.e., widely within the PBL, where the presence of turbulent fluxes determines a high degree of pollutants mixing and a high space-time homogeneity.

The meteorological and air quality measurements, acquired with high spatial-temporal resolution in the vertical profiles of urban and non-urban environments can be integrated and compared with the data measured at the ground level, allowing for the detailed characterization of atmospheric parameters within the PBL. This involves hypotheses and assumptions that are not always truthful: during the flight, sensors are in continuous movement, while ground-based measurements are typically carried out by fixed stations, located in strategic points of the city. In the case of the homing pigeons used in our study it follows that, even if a release takes place near a ground station, the comparison could only be carried out for a few seconds after the release, i.e., when the bird is still close to the station itself. In addition, birds decide themselves both the route and the altitude during the entire journey, resulting in a non-predictable trajectory of their flight at a fine scale. Furthermore, if, as in the present study, homing pigeons are used, they require in-group releases and have to familiarize themselves with specific locations of release, so that they can learn the route back to their loft. In fact, if the same pigeon is released several times at the same point, it will tend to memorize the shortest, straightest path to the loft and will tend to follow it on each subsequent flight [53]. This means that the paths will tend to become similar with time, with the advantage of making the measurements gathered in different releases more comparable.

Another fundamental aspect of these measurements is the temporal constraint. Depending on the distance between the release point and the loft and on the individual's experience, the pigeon can take a shorter/longer time to go back home, i.e., it will acquire a shorter/longer dataset. On one hand, this ensures that during the flight, the environmental conditions can be assumed as constant (in terms of temperature, humidity, concentrations of pollutants, etc.), but, on the other hand, short flights do not allow the study of the daily PBL evolution. However, this apparent limitation can be overcome by carrying out subsequent releases at different moments of the day. Another non-negligible constraint is linked to the fact that pigeons—similar to many diurnal birds—do not fly at night and in the case of bad weather conditions.

Moreover, the sensor design might influence the relationship between CO<sub>2</sub> and other measured atmospheric variables. In our case, the absence of a casing and the direct exposure of the atmospheric sensors to solar radiation might produce higher temperatures than the actual environmental values. This could be an issue, especially if the pigeon stays still

under the sun for long periods. This does not concern the data presented in this study since our model only includes flying periods. Given the optical nature of the CO<sub>2</sub> sensor, we believe direct exposure to solar radiation does not affect gas readings. Indeed, there is no mention of exposure in the sensor documentation [33]. Nonetheless, further development of the tag will include a casing, which is essential for longer deployments. This will warrant an analysis of design-dependence on all sensor readings.

In conclusion, we demonstrated the feasibility of using birds as biomonitoring tools to sample the air quality in an urban environment. Previously, this had been attempted only in two short-lived projects that, without wishing to minimise their social impact, did not yield reproducible scientific outputs. The integrated approach used in this study has a limited disturbance to the birds. It also has a limited environmental impact, because does not require the release of probes, and the energy consumption is only related to recharging the sensors' batteries and for data storage. This study was based on the use of data loggers, which must be recovered to download the data. With homing pigeons, this was easily done but can be difficult with wild birds. Moreover, this technological limit is being overcome, thanks to already available transmitters that can be integrated with the devices and are capable of sending real-time measurements through the fourth generation (4G) technologies. In this manner, it will be possible to monitor cities and other environments through wild species with a high temporal resolution, reducing costs and bureaucratic limitations and, thus, allowing for intervening even in emergency phases. Technological improvements and sensor miniaturization are increasing the scope of animal ecology, making animal tracking a useful way to gather high-resolution data about the environment in which they live [54]. This kind of development in the science of biologging has already been applied to the marine environment, but studies in terrestrial or aerial habitats are scant [17]. Birds are a valuable animal group in biomonitoring studies [55]. In particular, species able to perform controlled flights, such as homing pigeons, have been often used for studies on air pollution. Such studies have been mostly carried out in an ecotoxicological framework, usually requiring animal sacrifice and collecting data for limited periods [56–59]. To the best of our belief, this is the first study on environmental pollution with virtually no impact on birds.

## 5. Conclusions and Future Perspectives

This study reported the development of a miniaturized air sampler, integrating a GPS logger and atmospheric calibrated sensors. The device allowed for the acquisition of measurements of physical and chemical parameters, such as CO<sub>2</sub> concentration, barometric pressure, air temperature, and relative humidity with high spatial/temporal resolution ensured by the GPS. The air samplers were applied to homing pigeons and the results demonstrated the potential of atmospheric and air quality monitoring using birds. The air sampler developed in this study represents a low-cost, environmentally friendly, easy-to-use tool for environmental monitoring, providing enhanced observation and interpretation opportunities, with minimal effects on the well-being of the birds.

The release of homing pigeons in the urban centre of Rome (Italy) and their flight to the loft highlighted the reduction of the CO<sub>2</sub> concentration in the layers of the atmosphere close to the ground, passing from anthropized to rural and agricultural areas.

The results show that the CO<sub>2</sub> concentration varies considerably according to the level of urbanization, underlining the positive impact of green urban areas on air quality. Furthermore, the application of a stepwise regression reveals a positive relationship between CO<sub>2</sub> levels, relative humidity, and air temperature. Conversely, a negative relation between CO<sub>2</sub> concentration and distance from the point of release, month, and hour of the day has been found.

This research can be considered as a starting point for further studies, aimed at developing miniaturized sensors for the study of other atmospheric gases, wearable by other bird species (i.e., gulls) with very limited impact on their well-being, capable of flying at higher altitudes and over greater distances than pigeons. Furthermore, as mentioned,

sensor technology is constantly improving, both on the side of the electronics' performance and on the possibility of transmitting data through 4G technology. This latter development, together with the progress of sensor miniaturization, which increases the possible application on wearable devices, could provide the potential for the integration of data gathered by real-time air samplers with those from other devices as part of the Internet Of Things (e.g., [60]).

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## Appendix A

**Table A1.** List of pigeon releases, including departure and loft arrival times, flight duration, and track length, with a summary of CO<sub>2</sub> values measured.

Flight Number	Pigeon ID	Sample Points	Time (UTC)		Flight		CO <sub>2</sub> Concentration (ppm)			
			Release	Arrival	Duration (min)	Distance (km)	Mean	Min	Median	Max
1	p701	512	21 January 2021 08:22	21 January 2021 08:39	17.0	16.7	614	533	610	812
2	p788	882	26 January 2021 08:29	26 January 2021 09:07	37.7	27.1	568	498	557	725
3	p788	480	28 January 2021 07:55	28 January 2021 08:44	49.1	14.7	670	544	661	978
4	p710	1686	29 January 2021 08:15	29 January 2021 11:37	202.2	47.9	564	457	565	938
5	p788	756	29 January 2021 08:15	29 January 2021 08:40	25.2	24.0	666	578	652	993
6	p701	604	5 February 2021 09:05	5 February 2021 09:39	34.0	17.6	639	572	635	985
7	p788	669	5 February 2021 09:01	5 February 2021 10:24	83.1	19.9	587	489	571	923
8	p788	436	8 February 2021 08:37	8 February 2021 10:37	120.4	14.3	571	452	547	903

Table A1. Cont.

Flight Number	Pigeon ID	Sample Points	Time (UTC)		Flight		CO <sub>2</sub> Concentration (ppm)			
			Release	Arrival	Duration (min)	Distance (km)	Mean	Min	Median	Max
9	p561	693	19 March 2021 09:02	19 March 2021 09:25	23.1	18.7	497	476	489	576
10	p778	1007	1 April 2021 07:04	1 April 2021 12:13	309.2	25.1	562	504	555	913
11	p47	968	7 April 2021 07:06	7 April 2021 12:53	347.6	26.1	466	410	465	558
12	p778	1249	7 April 2021 07:01	7 April 2021 09:37	156.0	38.3	531	481	527	696
13	p47	256	9 April 2021 08:57	9 April 2021 09:06	8.5	9.4	455	436	453	481
14	p561	428	9 April 2021 08:52	9 April 2021 09:06	14.2	15.2	577	550	577	963
15	p47	463	21 April 2021 07:19	21 April 2021 07:34	15.4	14.7	606	514	564	963
16	p47	391	23 April 2021 08:15	23 April 2021 08:28	13.0	12.8	614	566	617	702
17	pG	373	4 May 2021 07:26	4 May 2021 08:09	42.7	11.1	516	472	516	618
18	pG	379	6 May 2021 07:18	6 May 2021 07:30	12.6	13.8	566	516	558	688
19	p787	351	7 May 2021 07:16	7 May 2021 07:38	22.1	13.5	567	436	560	703
20	pG	360	7 May 2021 07:26	7 May 2021 07:38	12.0	13.7	507	493	502	547
21	p34	561	11 May 2021 07:18	11 May 2021 08:40	81.9	15.1	586	492	587	730
22	pG	410	11 May 2021 07:18	11 May 2021 07:35	17.0	12.8	572	478	567	699
23	p684	3486	15 June 2021 11:19	16 June 2021 16:20	1740.3	82.6	586	528	582	869
24	p701	1287	15 June 2021 06:18	15 June 2021 14:49	510.2	35.8	565	439	569	793
25	p701	592	18 June 2021 06:35	18 June 2021 07:55	80.5	16.7	634	538	633	969

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