

Article

## Remote Sensing Estimates of Grassland Aboveground Biomass Based on MODIS Net Primary Productivity (NPP): A Case Study in the Xilingol Grassland of Northern China

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**Abstract:** The precise and rapid estimation of grassland biomass is an important scientific issue in grassland ecosystem research. In this study, based on a field survey of 1205 sites together with biomass data of the Xilingol grassland for the years 2005–2012 and the “accumulated” MODIS productivity starting from the beginning of growing season, we built regression models to estimate the aboveground biomass of the Xilingol grassland during the growing season, then further analyzed the overall condition of the grassland and the spatial and temporal distribution of the aboveground biomass. The results are summarized as follows: (1) The unitary linear model based on the field survey data and “accumulated” MODIS productivity data is the optimum model for estimating the aboveground biomass of the Xilingol grassland during the growing period, with the model accuracy reaching 69%; (2) The average aboveground biomass in the Xilingol grassland for the years 2005–2012 was estimated to be 14.35 Tg, and the average aboveground biomass density was estimated to be 71.32 g·m<sup>-2</sup>; (3) The overall variation in the aboveground biomass showed a decreasing trend from the eastern meadow grassland to the western

desert grassland; (4) There were obvious fluctuations in the aboveground biomass of the Xilingol grassland for the years 2005–2012, ranging from 10.56–17.54 Tg. Additionally, several differences in the interannual changes in aboveground biomass were observed among the various types of grassland. Large variations occurred in the temperate meadow-steppe and the typical grassland; whereas there was little change in the temperate desert-steppe and temperate steppe-desert.

**Keywords:** Xilingol; grassland; remote sensing; MOD17A2; aboveground biomass

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## 1. Introduction

Grassland ecosystems are one of the most important types of terrestrial ecosystems on the planet. They provide the ecosystem functions of soil and water conservation, wind erosion prevention, sand fixation and air purification. They also play an important role in the global terrestrial circulation among ecosystems. In addition, grassland ecosystems are important for “livestock production”. China’s grasslands encompass an area of approximately 400 million hectares, accounting for 41.7% of the country’s total land area [1], and constituting the single largest type of terrestrial ecosystem in China. Grassland biomass is the most direct indicator of grassland’s ecological status [2]. Thus, a precise and rapid method for the estimation of grassland biomass is of great importance for both basic science and the management and protection of grasslands.

Grassland biomass directly reflects the level of grassland productivity. Productivity refers to the amount of organic matter produced by autotrophic organisms in a given unit of area over a given unit of time. Grassland net primary productivity (NPP) is obtained by subtracting autotrophic respiration [3] from the total amount of organic matter fixed by grassland vegetation. Grassland biomass is defined as the yield of fresh grass or hay that can be harvested at a certain time, and is close to the net primary productivity without regard to disturbances. Productivity is the basis for the formation of biomass, and biomass is the manifestation of productivity. The currently available methods for the estimation of grassland productivity and biomass primarily include field surveys, statistical modeling [4–7], process modeling [8–11] and parameter modeling [12–16].

With the development of remote sensing technology, more and more fields have begun to involve remote sensing technology [17,18], and the development of remote sensing technology has stimulated studies on vegetation productivity and biomass. Due to the simple calculations involved and the high accuracy of the approach, statistical regression models using remote sensing data have been widely applied for the estimation of grassland biomass. The essence of the method is the establishment of a regression model between biomass measurements and either single-band remote sensing or vegetation index data, using satellite remote sensing data as the input parameters to estimate biomass. Initially, as the use of single-band information was simple, data of this type were used for estimating grassland biomass. However, due to strong interference from a number of factors (e.g., air, soil, sensor performance and the angle of the sun), the resulting estimation accuracy was found to be poor [19]. As such, the vegetation index is now commonly used as the form of input data for building empirical regression models of vegetation biomass. Piao, *et al.* [20] used the normalized difference vegetation

index (NDVI) calculated by NOAA/AVHRR to establish a grassland vegetation biomass estimation model for China, and used the model to study the spatial distribution characteristics of China's grassland vegetation biomass. Advances in remote sensing technology have generated higher-resolution images that are now used to estimate grassland productivity and biomass through methods, such as the application of MODIS data. Xu, *et al.* [21] performed a systematic estimation of China's grassland productivity by region using a combination of MODIS data and ground survey data for the same time period. For different types of grassland areas, they established relational models between NDVI and field survey biomass data that allowed them to estimate the distribution of grass production in China. Yang, *et al.* [22] employed enhanced vegetation index (EVI) data from MODIS to estimate aboveground biomass in Tibet and analyze the relationship between the grassland's aboveground biomass and climatic factors. Gao, *et al.* [23] used the MODIS vegetation index to conduct in-depth research on the spatial distribution of the aboveground and underground biomass of the Xilingol grassland.

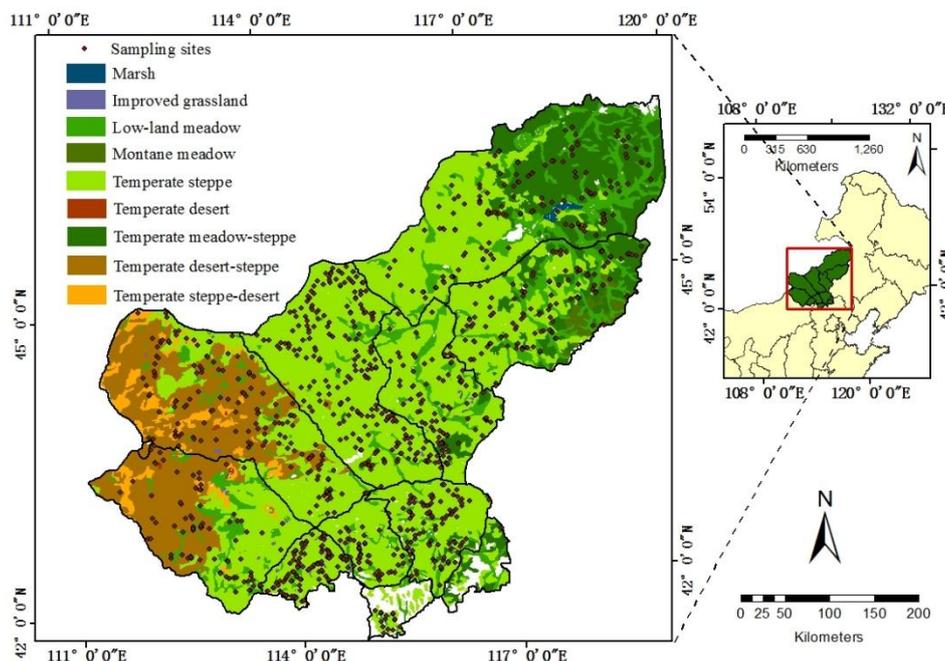
Using a remote sensing vegetation index to estimate aboveground biomass provides accurate and rapid results, but the limitations of the vegetation index itself may affect the obtained grassland biomass estimates. Specifically, in low-coverage grasslands, due to the significant influence of the soil background and grassland vegetation types, the estimation results exhibit high error rates. In addition, in high-coverage grasslands, the NDVI shows decreased sensitivity. A “saturation” phenomenon appears when the NDVI is higher, resulting in a decline in the accuracy of grassland biomass estimates. The MODIS productivity products regularly published by NASA are surface photosynthetic products obtained from model estimates. After processing these products, we established a relational model between the MODIS productivity products and field-measured aboveground biomass. We then estimated aboveground biomass by accepting the MODIS productivity data as input parameters to avoid or reduce saturation problems if the input parameters were too high. In the present study, based on this strategy, we used ground survey data from the Xilingol grassland for the years 2005–2012 and MODIS productivity data for the same time period to establish statistics-based models for estimating biomass. We further tested the accuracy of the models and selected the optimal model for estimating the aboveground biomass of the Xilingol grassland during the growing period.

## 2. Materials and Methods

### 2.1. Overview of the Study Area

The Xilingol grassland, located in the central part of Inner Mongolia at 41°35'~46°46'N, 111°09'~119°58'E (Figure 1), is a typical temperate grassland of northern China, with an arid and semiarid temperate continental monsoon climate. The total area of the Xilingol grassland is 192,512 km<sup>2</sup>, accounting for 95.03% of the total land area of Xilingol. The map of grassland resources in China at a 1:1,000,000 scale shows that the vegetation types in the grassland, ranging from east to west, are meadow grassland, typical grassland and desert grassland (Figure 1). The representative plants of the meadow grassland are *Leymus chinensis* and *Stipa baicalensis*, whereas those of the typical grassland are bunch grasses such as *Leymus chinensis*, *Stipa grandis* and *Stipa krylovii*, and those of the desert grassland are dwarf grasses such as *Stipa klemenzi*, *Stipa glareosa* and *Cleistogenes squarrosula* [24].

**Figure 1.** Grassland types and the distribution of sampling sites in the study area. The total number of sampling sites was 1205.



## 2.2. Field Sampling Data

The aboveground biomass data used in this study came from multi-year field survey data collected by our research group, and a large-scale field campaign organized by the Grassland Monitoring and Supervision Center Ministry of Agriculture of China (GMSC), primarily in July and August from 2005–2012 [23]. The sampling sites, each of which had an area of at least 1 km<sup>2</sup>, were chosen to represent typical vegetation communities. Each sampling site was homogeneous in terms of both its vegetation and land types. To obtain the actual aboveground biomass of herbs, all the aboveground plants in the three plots (1 m × 1 m) were harvested to measure their fresh weight. To obtain the estimated aboveground biomass of shrubs in the homogeneous grassland, one plot (10 m × 10 m) was sampled. We sorted the shrub plants (clumps) into three groups by size (*i.e.*, large, medium and small). A representative plant (clump) from each group was selected, and its green parts along with its branches were cut and weighed from the same year. For each shrub plot, we multiplied the weight of the representative plant from each group by the number of plants in that group. We then calculated the sum of the weights of the three shrub groups. Finally, we averaged the fresh weight of both the herbs and shrubs in the three plots at the sampling site to obtain the field sampling dataset.

Because the estimation accuracy for the model was significantly affected by the quality of ground survey data, the ground survey data were strictly tested and standardized before the creation of the model. Based on the grassland type and average conditions over many years, certain data showing an abnormal performance were excluded from the dataset. Ultimately, a total of 1205 sites sampled for biomass data were selected, of which 975 aboveground sites were used to establish the model and the remaining 230 sites were employed for validating the accuracy of the model. As shown in Figure 1, the number of sampling sites was sufficiently large and their distribution was sufficiently even to accurately represent the ground biomass and obtain statistically significant biomass models.

### 2.3. Remote Sensing Data and its Processing

#### 2.3.1. Introduction to MODIS Productivity Data

The MODIS satellite sensor is one of the major sensors of the U.S. Earth Observation System (EOS). Generalized MODIS data cover 44 types of products. The MODIS data used in this study came from the MOD17A2 eight-day PSNnet (net photosynthesis) product, with a resolution of 1 km, which is a global land vegetation net productivity product calculated by model [25] and has been validated and widely used in studies on productivity and biomass in different regions of the world [26–29].

The MOD17 product is a land productivity product calculated using the BIOME-BGC model and a light use efficiency model, in combination with remote sensing data. The essence of the core science in the MOD17 algorithm is the application of radiation conversion efficiency logic to predictions of daily GPP (Gross Primary Productivity) using satellite-derived FPAR (Fraction of Photosynthetically Active Radiation, from MOD15) and independent estimates of PAR (Absorption Ratio of Photosynthetically Active Radiation) and other surface meteorological fields (from the DAO). The subsequent estimates of maintenance and growth respiration terms are subtracted from GPP to arrive at NPP (Net Primary Productivity). The principles underlying daily PSNnet estimation are as follows:

$$\varepsilon = \varepsilon_{\max} \times \text{TMIN\_Scalar} \times \text{VPD\_Scalar} \quad (1)$$

where,  $\varepsilon$  is the light use efficiency,  $\varepsilon_{\max}$  is the maximum light use efficiency obtained from the Biome Parameter Look-Up Table (BPLUT) provided by NASA; TMIN\_Scalar is the daily minimum temperature scalar, and VPD\_Scalar is the daily average vapor pressure deficit scalar, which are used to calculate the scalars that attenuate  $\varepsilon_{\max}$  to produce the final  $\varepsilon$  used to predict GPP.

$$\text{IPAR} = \text{SWRad} \times 0.45 \quad (2)$$

$$\text{APAR} = \text{IPAR} \times \text{FAR} \quad (3)$$

$$\text{GPP} = \varepsilon \times \text{APAR} \quad (4)$$

where, SWRad is the incident shortwave radiation provided by the Data Assimilation Office (DAO) of NASA; IPAR is the photosynthetically active radiation incident on the vegetative surface; FAR is the absorption ratio of photosynthetically active radiation; and APAR is the photosynthetically active radiation absorbed by vegetation.

$$\text{Leaf\_Mass} = \text{LAI}/\text{SLA} \quad (5)$$

where, LAI is the leaf area index obtained from the MOD15, and SLA is the projected leaf area for a given pixel obtained from the BPLUT.

$$\text{Leaf\_MR} = \text{Leaf\_Mass} \times \text{leaf\_mr\_base} \times \text{Q10\_mr}^{[(\text{Tavg}-20.0)/10.0]} \quad (6)$$

where, leaf\_mr\_base is the maintenance respiration of leaves obtained from the BPLUT and Tavg is the average daily temperature (°C) estimated from the DAO meteorological data; Leaf MR is the leaf autotrophic respiration consumption.

$$\text{Froot\_Mass} = \text{Leaf\_Mass} \times \text{Froot\_leaf\_ratio} \quad (7)$$

$$\text{Froot\_MR} = \text{Fine\_Root\_Mass} \times \text{froot\_mr\_base} \times \text{Q10\_mr}^{[(\text{Tavg}-20.0)/10.0]} \quad (8)$$

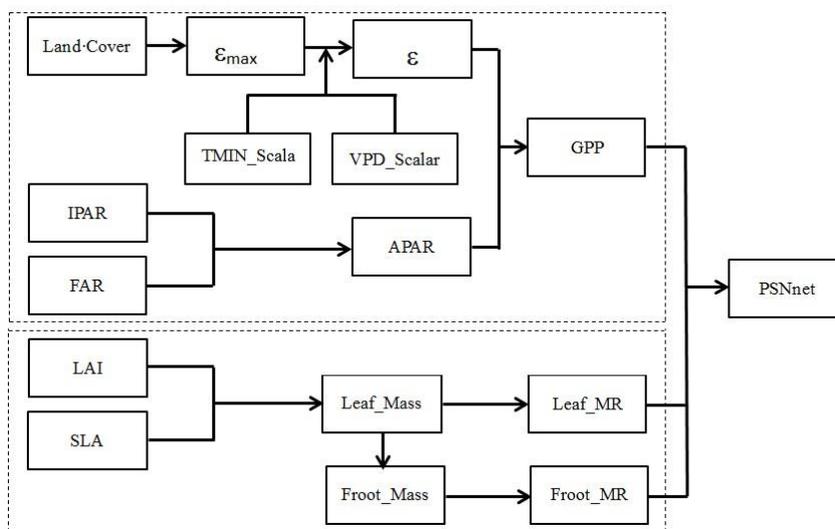
where, *Froot\_leaf\_ratio* is the ratio of fine root to leaf mass obtained from the BPLUT; *Froot\_mr\_base* is the maintenance respiration per unit of fine roots at 20 °C obtained from the BPLUT; and *Froot\_MR* is the fine root autotrophic respiration consumption.

Finally, PSNnet (kg·C·day<sup>-1</sup>) can be calculated from GPP using Equation (9):

$$\text{PSNnet} = \text{GPP} - \text{Leaf\_MR} - \text{Froot\_MR} \tag{9}$$

The process for the estimation of the MOD17A2 GPP (gross primary productivity) and PSNnet (net photosynthesis) products is illustrated in Figure 2.

**Figure 2.** Flowchart of the PSNnet estimation process.



### 2.3.2. Processing of Remote Sensing Data

Prior to conducting biomass estimates using the MOD17A2 eight-day PSNnet data, preprocessing of the data was necessary. First, a mosaic image was obtained, and geometric projection transformation was performed. The administrative boundary vector data for the Xilingol grassland were then used to extract the data for the study area. In addition, there was noise in certain areas of the PSNnet data where pixels were filled with the invalid value 32,767. In this study, the noise mostly appeared in desert grasslands, which essentially have no vegetation cover. Therefore, to ensure the accuracy of the final estimates during processing, all these invalid values were set to 0. Because there were many ground sampling sites selected for this study and they were evenly distributed, setting the values in a few areas to 0 does not have a significant effect on the statistical results. Further, to obtain a more accurate start time for the growing season and to ensure the accuracy of the accumulated productivity data, a statistical analysis of the productivity time series of the Xilingol grassland for the years 2005–2012 was conducted. The start time of PSNnet value every eight-day great than 0 was defined the start of growing season here, and the analytical results indicated a start time for the growing season of early to mid-March. Furthermore, the accumulation of PSNnet data every eight days from the beginning of the growing season to the peak of the growing season was obtained. Specifically, we can extract the mean PSNnet within a 1 km range for each sampling sites from the accumulated PSNnet data and establish a database of PSNnet and fresh grass weight data for the corresponding time. Then, based on the database, a statistical model was built for inversion of the aboveground biomass.

2.4. Establishment and Verification of the Model

In this study, the field sampling data was wet weight and needed to be converted to air-dried weight. We converted the wet weights to air-dried weights using conversion coefficients for different grassland types employing “Chinese Grassland Resources” [30] as a reference, and we further converted the air-dried weights to dry weights with a 15% water content [31], as the indicator of the grassland aboveground biomass.

Based on the database, regression models between the aboveground biomass and the corresponding accumulated PSNnet value were established. These models primarily consisted of unitary linear, logarithm function, power function and exponential function models (Table 1). Using the reserved ground sampling data (approximately 20% of the total samples), we calculated the root-mean-square error (RMSE) and mean relative estimation error (REE) between the estimated values and the ground-measured values to evaluate the accuracy of the models [32]. The equations employed for error calculation were as follows:

$$RMSE = \sqrt{\frac{\sum(Y_i - Y_i')^2}{N}} \tag{10}$$

$$REE = \sqrt{\frac{\sum[(Y_i - Y_i')/Y_i']^2}{N}} \tag{11}$$

where,  $Y_i$  is the actual aboveground biomass (fresh weight),  $Y_i'$  is the aboveground biomass estimated by the model and  $N$  is the sample size.

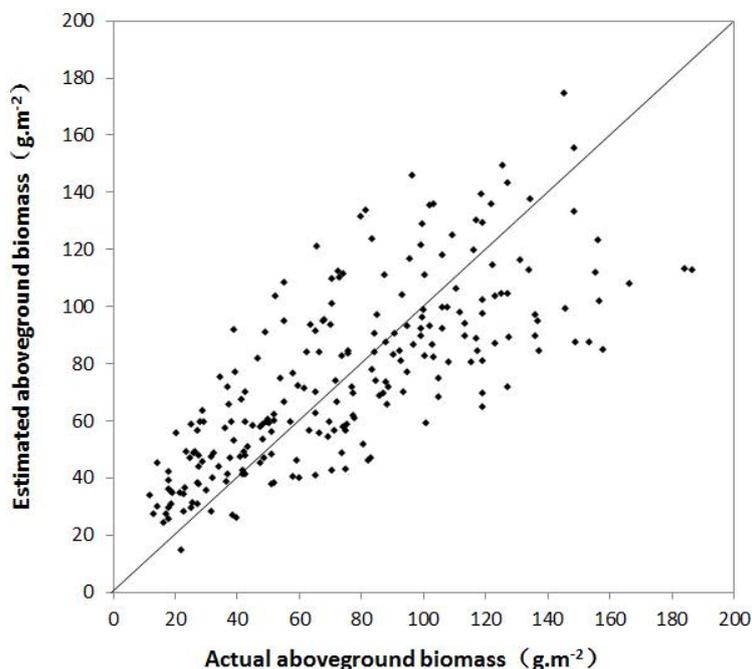
**Table 1.** Statistical models relating aboveground biomass to accumulated PSNnet.

| Model                   |                                     | $R^2$ | F Value | RMSE<br>( $g \cdot m^{-2}$ ) | REE  | Precision<br>(%) |
|-------------------------|-------------------------------------|-------|---------|------------------------------|------|------------------|
| Unitary linear function | $y = 1.097x - 4.776$                | 0.55  | 1168.75 | 26.67                        | 0.31 | 69               |
| Logarithm function      | $y = 71.308 \times \ln x - 224.634$ | 0.55  | 1154.46 | 26.77                        | 0.40 | 60               |
| Power function          | $y = 0.361x^{1.226}$                | 0.65  | 1739.58 | 27.43                        | 0.39 | 61               |
| Exponential function    | $y = 17.038 \times e^{0.0178x}$     | 0.58  | 1315.50 | 34.52                        | 0.43 | 57               |

Note: The numbers of regression and test samples were 975 and 230, respectively;  $p < 0.0001$ ; precision =  $(1 - REE) \times 100\%$ ;  $x$  is the accumulated PSNnet data ( $C_g \cdot m^{-2}$ ,  $C$  presents carbon),  $y$  is the estimated aboveground biomass ( $g \cdot m^{-2}$ ).

The results indicate that the precision of the exponential function model was relatively low, whereas the precision of the unitary linear regression model was highest, at 69%. The correlation between the grassland aboveground biomass estimated by the unitary linear model and the actual aboveground biomass was significant (Figure 3). Therefore, we selected the unitary linear regression model for estimating the aboveground biomass of the Xilingol grassland, and then analyzed the temporal and spatial characteristics of the aboveground biomass of the Xilingol grassland.

**Figure 3.** Relationship between the estimated aboveground biomass and actual aboveground biomass.



### 3. Results

#### 3.1. Total Aboveground Biomass and its Distribution in Different Banners

The unitary linear regression model was selected for estimating the aboveground biomass of the Xilingol grassland for the years 2005–2012 (Table 2). The average aboveground biomass of the Xilingol grassland for the years 2005–2012 was 14.35 Tg (1 Tg = 10<sup>12</sup> g). The annual average aboveground biomass was highest in East Ujimqin Banner (approximately 4.15 Tg), 28.92% of the total biomass of the Xilingol grassland. The annual average aboveground biomass in West Ujimqin Banner was approximately 2.58 Tg, 17.98% of the total biomass in the Xilingol grassland. As the biomass of these two Banners represented nearly one-half of the total biomass, the two banners play an important role in pasture management and in the protection of grassland resources in Xilingol. The annual average aboveground biomass in Erenhot City, Taibus Banner, Xianghuang Banner, Duolun County and Zhengxiangbai Banner were relatively low, accounting for less than 10% of the aboveground biomass of the Xilingol grassland.

An analysis of the data presented in Table 3 showed that the average density of the aboveground biomass of the Xilingol grassland for the years 2005–2012 was 71.32 g·m<sup>-2</sup>. This value is generally in agreement with the average aboveground biomass density estimated for the Xilingol grassland for the years 2001–2012 by Gao, *et al.* [23] using the NDVI. In addition, there was a substantial difference in the aboveground biomass density among the different banners. The aboveground biomass density was greater than 90 g·m<sup>-2</sup> in Duolun County, Taipusi Banner, West Ujimqin Banner, Zhenglan Banner and East Ujimqin Banner, where growth conditions for grassland vegetation are generally good. In contrast, the aboveground biomass density in Erenhot City, Sonid Right Banner and Sonid Left Banner were relatively low, as growth conditions for grassland vegetation were generally poor in these areas.

**Table 2.** Annual aboveground biomass in different banners.

| Banners              | Grassland Area (km <sup>2</sup> ) | Total Aboveground Biomass (Tg) |        |        |        |        |        |        |        |         |
|----------------------|-----------------------------------|--------------------------------|--------|--------|--------|--------|--------|--------|--------|---------|
|                      |                                   | 2005                           | 2006   | 2007   | 2008   | 2009   | 2010   | 2011   | 2012   | Average |
| East Ujimqin Banner  | 45,060                            | 5.72                           | 3.79   | 2.49   | 3.90   | 4.06   | 4.15   | 4.11   | 5.00   | 4.15    |
| Abaga Banner         | 27,492                            | 1.79                           | 1.51   | 1.38   | 2.00   | 1.73   | 1.65   | 1.56   | 2.40   | 1.75    |
| West Ujimqin Banner  | 23,726                            | 3.49                           | 2.53   | 1.83   | 2.68   | 2.08   | 2.50   | 2.46   | 3.06   | 2.58    |
| Sonid Left Banner    | 34,814                            | 1.21                           | 1.00   | 1.01   | 1.30   | 1.20   | 0.95   | 1.05   | 1.54   | 1.15    |
| Xilinhot City        | 15,753                            | 1.44                           | 1.23   | 1.00   | 1.40   | 1.10   | 1.28   | 1.09   | 1.60   | 1.27    |
| Erenhot City         | 186                               | 0.0029                         | 0.0020 | 0.0017 | 0.0026 | 0.0020 | 0.0015 | 0.0017 | 0.0025 | 0.0021  |
| Sonid Right Banner   | 25,212                            | 0.90                           | 0.70   | 0.59   | 0.80   | 0.67   | 0.55   | 0.65   | 0.88   | 0.72    |
| Zhenglan Banner      | 10,281                            | 1.09                           | 1.05   | 0.84   | 1.15   | 0.99   | 0.87   | 1.05   | 1.15   | 1.02    |
| Zhengxiangbai Banner | 6256                              | 0.48                           | 0.48   | 0.36   | 0.50   | 0.42   | 0.35   | 0.46   | 0.52   | 0.44    |
| Xianghuang Banner    | 5018                              | 0.30                           | 0.31   | 0.22   | 0.29   | 0.27   | 0.23   | 0.31   | 0.33   | 0.28    |
| Duolun County        | 3955                              | 0.59                           | 0.56   | 0.47   | 0.57   | 0.53   | 0.48   | 0.58   | 0.58   | 0.54    |
| Taipusi Banner       | 3496                              | 0.43                           | 0.47   | 0.36   | 0.50   | 0.39   | 0.38   | 0.46   | 0.48   | 0.43    |
| Total                | 201,249                           | 17.44                          | 13.62  | 10.56  | 15.07  | 13.42  | 13.39  | 13.79  | 17.54  | 14.35   |

**Table 3.** Aboveground biomass in different banners.

| Banners              | Grassland Area (km <sup>2</sup> ) | Aboveground Biomass Density (g·m <sup>-2</sup> ) |        |       |        |       |        |        |        |         |
|----------------------|-----------------------------------|--|--------|-------|--------|-------|--------|--------|--------|---------|
|                      |                                   | 2005   | 2006   | 2007  | 2008   | 2009  | 2010   | 2011   | 2012   | Average |
| East Ujimqin Banner  | 45,060                            | 127.04   | 84.20  | 55.26 | 86.54  | 90.03 | 92.13  | 91.20  | 111.03 | 92.18   |
| Abaga Banner         | 27,492                            | 65.19  | 55.07  | 50.18 | 72.66  | 62.93 | 60.09  | 56.88  | 87.03  | 63.75   |
| West Ujimqin Banner  | 23,726                            | 147.30   | 106.82 | 77.13 | 112.89 | 87.87 | 105.18 | 103.72 | 129.01 | 108.74  |
| Sonid Left Banner    | 34,814                            | 34.62  | 27.98  | 29.09 | 37.37  | 34.37 | 27.31  | 30.24  | 44.28  | 33.16   |
| Xilinhot City        | 15,753                            | 91.60  | 78.25  | 63.23 | 88.41  | 69.78 | 81.17  | 69.25  | 101.70 | 80.42   |
| Erenhot City         | 186                               | 15.54  | 10.85  | 9.40  | 14.06  | 10.80 | 7.88   | 9.13   | 13.48  | 11.39   |
| Sonid Right Banner   | 25,212                            | 35.53  | 27.50  | 23.60 | 31.80  | 26.38 | 21.84  | 25.90  | 34.89  | 28.43   |
| Zhenglan Banner      | 10,281                            | 105.60   | 102.60 | 81.82 | 111.67 | 95.95 | 84.87  | 101.70 | 111.58 | 99.47   |
| Zhengxiangbai Banner | 6256                              | 76.18  | 76.76  | 57.60 | 79.26  | 66.34 | 55.31  | 73.53  | 83.30  | 71.03   |
| Xianghuang Banner    | 5018                              | 59.94  | 61.44  | 44.64 | 57.57  | 52.96 | 45.02  | 62.25  | 65.43  | 56.16   |

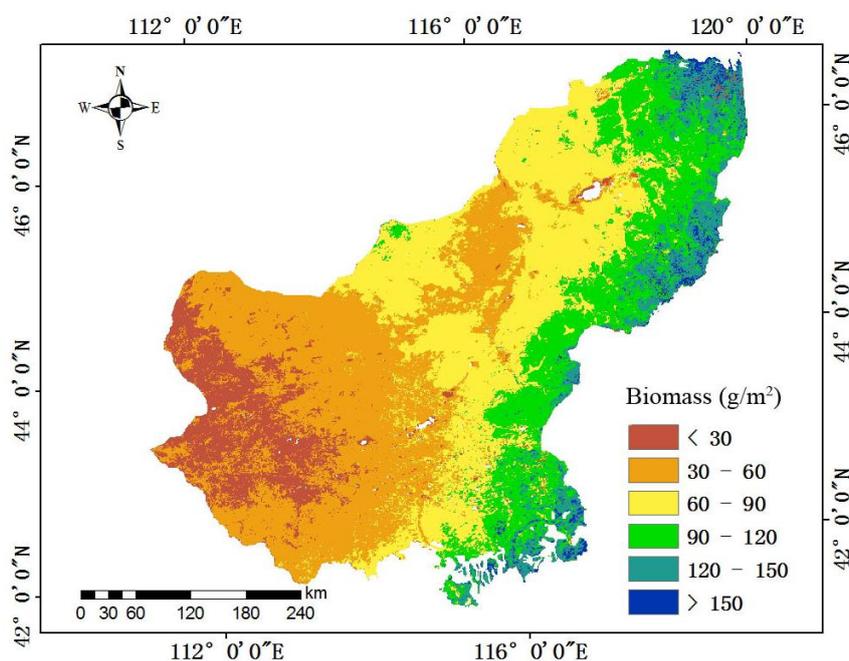
Table 3. Cont.

| Banners        | Grassland Area (km <sup>2</sup> ) | Aboveground Biomass Density (g·m <sup>-2</sup> ) |        |        |        |        |        |        |        |         |
|----------------|-----------------------------------|--|--------|--------|--------|--------|--------|--------|--------|---------|
|                |                                   | 2005   | 2006   | 2007   | 2008   | 2009   | 2010   | 2011   | 2012   | Average |
| Duolun County  | 3955                              | 149.48   | 140.62 | 118.06 | 144.66 | 133.70 | 121.23 | 146.92 | 145.73 | 137.55  |
| Taipusi Banner | 3496                              | 122.19   | 135.71 | 103.88 | 141.70 | 110.21 | 108.85 | 131.52 | 138.07 | 124.02  |
| Total          | 201,249                           | 86.66  | 67.67  | 52.48  | 74.90  | 66.66  | 66.51  | 68.53  | 87.15  | 71.32   |

3.2. Spatial Distribution of Aboveground Biomass

There were significant spatial differences in the aboveground biomass of the Xilingol grassland (Figure 4). The overall trend of aboveground biomass was as follows: the eastern meadow grassland > the central typical grassland > the western desert grassland (Table 4). The average aboveground biomass density for the years 2005–2012 was highest in the eastern montane meadow (156.07 g·m<sup>-2</sup>) and temperate meadow steppe (115.23 g·m<sup>-2</sup>), followed by the low-land meadow (86.31 g·m<sup>-2</sup>) and the temperate steppe (69.10 g·m<sup>-2</sup>). Among these four grassland types, the area of the temperate steppe was largest, accounting for 55.19% of the total area of the Xilingol grassland and 54.19% of the total aboveground biomass in the study area. The aboveground biomass density in the temperate desert-steppe and the temperate steppe-desert was markedly lower than the overall average density (70.36 g·m<sup>-2</sup>). The western temperate desert contained almost no plants. Precipitation is the major factor influencing the differences in the spatial distribution of grassland aboveground biomass [33]. The annual precipitation in most areas of the Xilingol grassland was within a range of 200–300 mm, with an overall trend involving a gradual decrease from east to west. This trend was consistent with the spatial distribution pattern of the aboveground biomass [34].

Figure 4. Spatial distribution of the aboveground biomass of the Xilingol grassland (g·m<sup>-2</sup>) from 2005–2012.



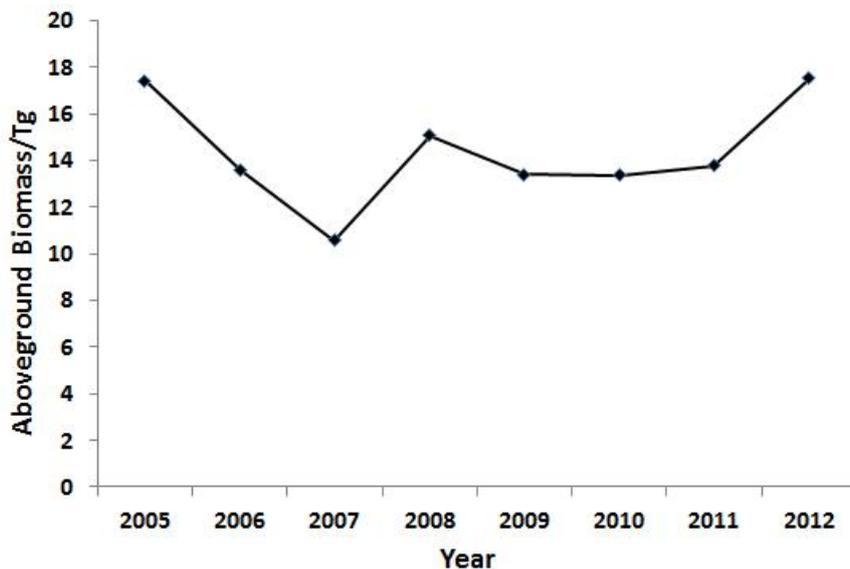
**Table 4.** Aboveground biomass density in the different grassland types.

| Grassland Types         | Grassland Area (km <sup>2</sup> ) | Aboveground Biomass Density (g·m <sup>-2</sup> ) |        |        |        |        |        |        |        |         |
|-------------------------|-----------------------------------|--|--------|--------|--------|--------|--------|--------|--------|---------|
|                         |                                   | 2005   | 2006   | 2007   | 2008   | 2009   | 2010   | 2011   | 2012   | Average |
| Low-land meadow         | 25,981                            | 109.74   | 83.87  | 62.04  | 88.41  | 79.65  | 80.92  | 84.84  | 101.02 | 86.31   |
| Improved grassland      | 472                               | 67.42  | 63.52  | 50.80  | 68.73  | 58.06  | 54.71  | 60.26  | 73.89  | 62.17   |
| Montane meadow          | 1581                              | 202.45   | 154.89 | 123.03 | 153.47 | 145.24 | 145.58 | 159.17 | 164.75 | 156.07  |
| Temperate meadow-steppe | 24,875                            | 155.81   | 110.19 | 75.29  | 109.83 | 110.44 | 111.83 | 117.66 | 130.82 | 115.23  |
| Temperate steppe-desert | 5108                              | 24.94  | 19.78  | 19.66  | 24.91  | 23.72  | 17.07  | 20.33  | 28.69  | 22.39   |
| Temperate steppe        | 108,488                           | 80.40  | 65.11  | 51.62  | 75.06  | 63.34  | 64.82  | 64.39  | 88.02  | 69.10   |
| Temperate desert-steppe | 29,598                            | 30.68  | 23.36  | 23.89  | 31.03  | 27.79  | 21.17  | 24.20  | 35.71  | 27.23   |
| Temperate desert        | 143                               | 22.93  | 16.58  | 16.91  | 22.43  | 18.53  | 14.64  | 16.94  | 23.93  | 19.11   |
| Marsh                   | 330                               | 126.45   | 71.32  | 55.81  | 83.17  | 75.64  | 84.50  | 84.36  | 97.95  | 84.90   |
| Total                   | 196,576                           | 85.88  | 66.52  | 51.54  | 73.88  | 65.71  | 65.71  | 67.39  | 86.29  | 70.36   |

3.3. Interannual Variation in Aboveground Biomass

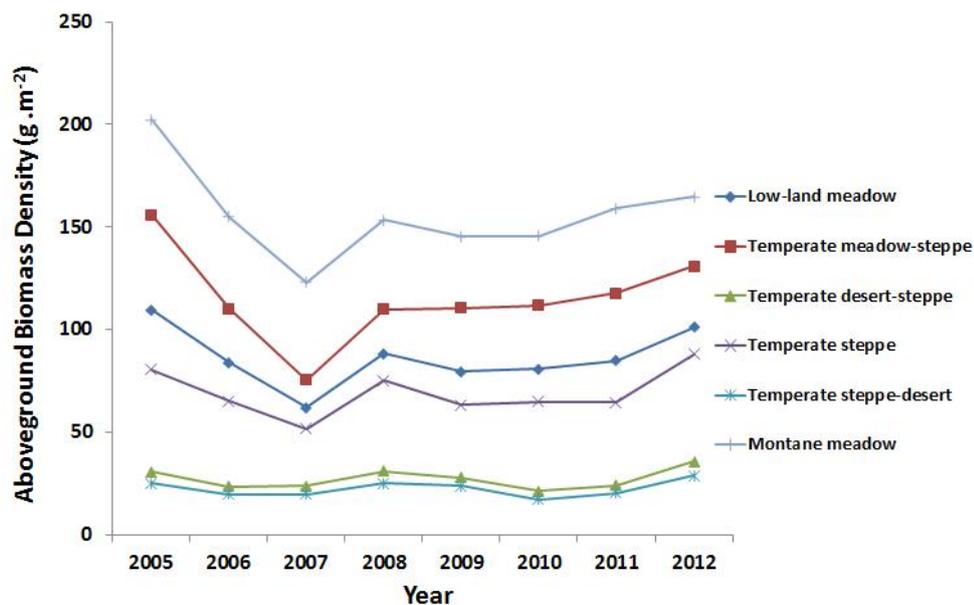
There were obvious fluctuations in the aboveground biomass of the Xilingol grassland between the years 2005 and 2012. These biomass values ranged from 10.56–17.54 Tg. As shown in Figure 5, the lowest biomass occurred in 2007, at 10.56 Tg, approximately 26% lower than the mean annual average biomass. The highest biomass was 17.54 Tg, in 2012. There were large interannual fluctuations in the aboveground biomass of the Xilingol grassland from 2005–2009, with an initial decrease being observed prior to an increasing trend. The grassland aboveground biomass from 2009–2012 exhibited an increasing trend, with biomass peaking in 2012.

**Figure 5.** Variation in the aboveground biomass of the Xilingol grassland from 2005–2012.



The aboveground biomass of different grassland types appeared to fluctuate from 2005–2012 (Figure 6). Large variations occurred in the temperate meadow-steppe and the typical grassland, which presented higher yields; whereas there was little change in the temperate desert-steppe and temperate steppe-desert, which showed lower yields. From 2005–2012, the overall grassland production in the montane meadow and the temperate meadow-steppe decreased, whereas it remained steady after showing fluctuations in the lowland and the temperate meadow and showed little change in the temperate desert-steppe and the temperate steppe-desert.

**Figure 6.** Interannual variations in the aboveground biomass in different grassland types.



## 4. Discussions

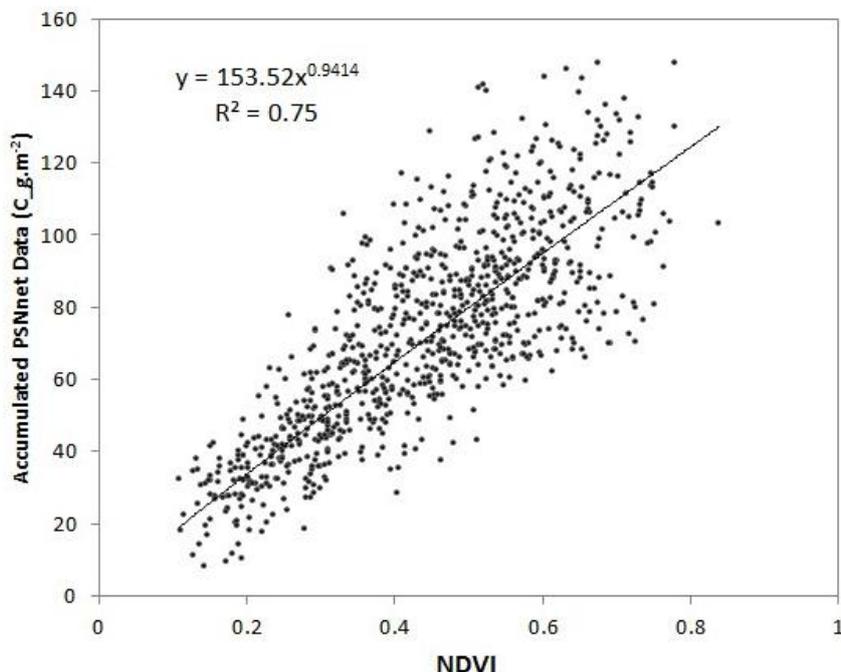
### 4.1. Potentials Analysis of Model

A significant empirical relationship was found between the accumulated PSNnet data and the aboveground biomass data, showing good prospects for the application of MODIS productivity data in combination with ground sampling data to establish models for biomass inversion. There are three reasons to support this conclusion, as follows:

(1) Through the accumulation of PSNnet data every eight days from the beginning of the growing season to the peak of the growing season, a good correlation was achieved between the obtained accumulated PSNnet data and the peak growing season NDVI data for the corresponding spatial point, showing coefficients of determination ( $R^2$ ) up to 0.75 (Figure 7). Because there have been many previous studies on statistical models of the NDVI and aboveground biomass [35–38], we can assume a good correlation between the PSNnet data and the aboveground biomass data. Furthermore, we can assume that using the PSNnet data and the aboveground biomass data to build the model and then retrieve the aboveground biomass data is a practical method. In addition, the MODIS productivity data used in this study fully considered the effects of temperature, precipitation and other climatic factors during the process of estimating vegetation productivity. Compared with the NDVI data employed in

traditional methods of biomass estimation, MODIS productivity data can better reflect the effects of environmental stresses.

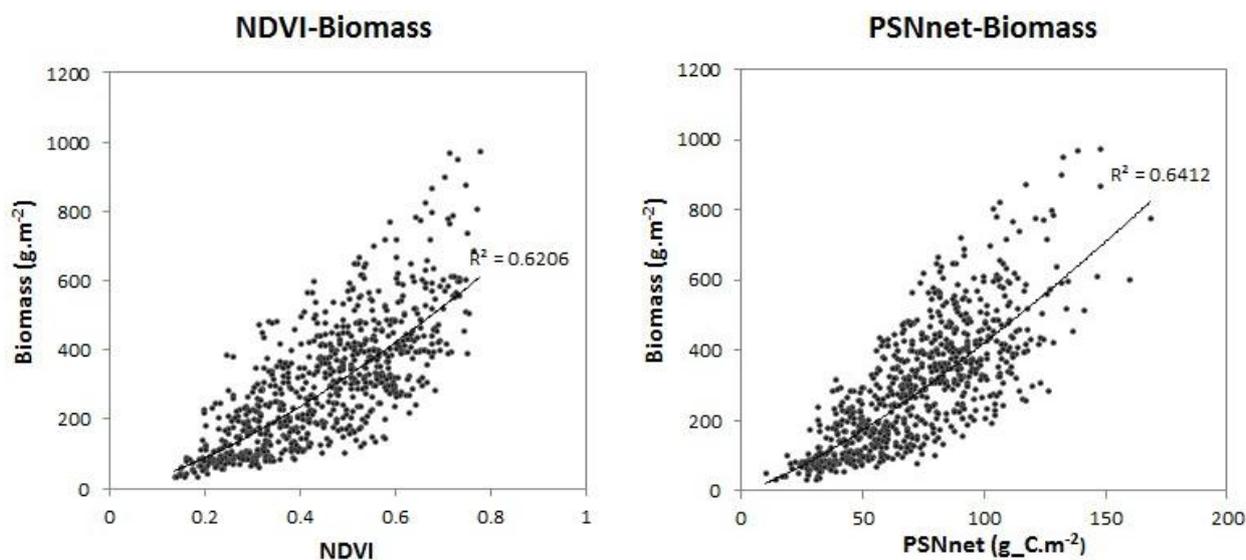
**Figure 7.** Relationship between the accumulated PSNnet data and NDVI data.



Note: x is NDVI, y is the accumulated PSNnet data (C\_g.m<sup>-2</sup>).

(2) The temporal matching between remotely sensed images and ground survey greatly affects the accuracy of remote sensing based models for grassland biomass estimation. The database of the extensive filed samples and their matching remotely sensed data is the basis of improving the model precision and stability. In this study, a sound database combining multi-year accumulated PSNnet data and ground survey biomass data with strict temporal matching was developed, which was further applied to biomass estimation models.

(3) Two methods are often used to evaluate model performances. One is based on the coefficient of determination ( $R^2$ ), another way is to assess model error. In general, a high  $R^2$  or a low error value often indicates a good fit between the model developed and the ground survey data [39]. In this study, we compared the correlation between NDVI data and biomass, as well as accumulated PSNnet data and biomass. The result showed that  $R^2$  between calculated PSNnet data and biomass was a little higher than  $R^2$  between NDVI data and biomass, as shown in Figure 8. We used ground survey data from the Xilingol grassland for the years 2005–2012 and MODIS productivity data for the same time period to establish statistics-based models for biomass estimation, with an overall accuracy of 69%, which is close to highest accuracy (74%) by Jin, *et al.* [38]. In addition, NDVI data are prone to an “oversaturation” phenomenon if the vegetation coverage is higher, which decreases the sensitivity of biomass estimation, whereas MODIS productivity data can overcome this oversaturation problem. Therefore, in high vegetation cover conditions, the biomass estimation accuracy by MODIS productivity data would be higher than the biomass estimation accuracy by NDVI data.

**Figure 8.**  $R^2$  between remotely sensed data and ground survey data.

#### 4.2. Uncertainties of Model

In this study, based on ground survey data from the Xilingol grassland for the years 2005–2012 and MODIS productivity data for the same time period, we developed statistical models based on the relationship between the PSNnet data and aboveground biomass data and then selected the optimal model to estimate the grassland aboveground biomass of the Xilingol grassland. However, our estimate still retains some uncertainties for the following reasons. First, there exists the scale effect between remotely sensed data and ground survey data. The spatial resolution of MODIS productivity data is 1 km, differing from the size of quadrat. In the course of the sampling, we have taken the average value of multiple samples to reduce the estimation error caused by the scale effect. Second, the remotely sensed data used in this study came from the MOD17A2 eight-day PSNnet product, which is a global land vegetation net productivity product calculated by model and has been validated widely, across most of the world. In the process of performing these calculations, nevertheless, since some of the maintenance respiration costs and all of the growth respiration costs have not been accounted for in the daily timestep, the daily output from this algorithm is termed PSNnet (net photosynthesis), to differentiate it from the true daily NPP. This reason has magnified the contrast between net photosynthesis and actual biomass, and becomes the important uncertainty source for biomass estimation. In addition, although the MODIS-PSNnet dataset has been processed with a series of corrections, there is still a certain level of deviation, which causes uncertainty in the estimation. Understanding and identifying the sources of uncertainties and then devoting efforts to improving them are critical in improving grassland aboveground biomass estimation performance; therefore, more research is needed in the future for reducing the uncertainties from different sources in the aboveground biomass estimation procedure.

#### 4.3. Comparison with Previous Estimates

As shown in Table 4, the average aboveground biomass densities were estimated to be  $27.23 \text{ g}\cdot\text{m}^{-2}$  for the temperate desert steppe,  $69.10 \text{ g}\cdot\text{m}^{-2}$  for the temperate steppe and  $115.23 \text{ g}\cdot\text{m}^{-2}$  for the temperate

meadow steppe. A comparison of our estimates with these previous values (Table 5 [40–43]) showed that our estimate was significantly lower than the previous estimates, and the reasons for this difference could be related to the following three aspects. First, the influence on ground survey data from different times and regions is the major reason for this difference. For example, we used field sampling data from multi-year field survey data collected by our research group and a large-scale field campaign organized by the GMSC between 2006 and 2012, whereas Ma, *et al.* [41] gathered 113 field samples from 2002–2005, Piao, *et al.* [42] used national grassland resource inventory data between 1982 and 1999. Previous studies have mainly focused on large scales, including all types of grasslands in China or Inner Mongolia, whereas this study only concerns grasslands in Xilingol. Second, the sampling method employed in these studies may have contributed to this difference. Although biomass harvest is a commonly used method of grassland quadrat survey, there are differences in actual practical operations. Different standards of sampling, including sample locations (enclosed or not) and collected objects (including standing dead biomass and litter biomass or not), dramatically influence sampling. The third aspect is the approach taken in the estimation process. For example, Ma, *et al.* [41] calculated the biomass densities for different grassland types based on field samples and further estimated the biomass according to different grassland types. Compared with this method, the method based on remote sensing can represent the spatial details of the aboveground biomass across the entire study area, thereby expanding the study area and reducing the uncertainty of the estimates.

**Table 5.** Aboveground biomass densities from different studies.

| Researchers              | Study Area     | Aboveground Biomass Densities ( $\text{g}\cdot\text{m}^{-2}$ ) |                  |                         |
|--------------------------|----------------|--|------------------|-------------------------|
|                          |                | Temperate Desert Steppe  | Temperate Steppe | Temperate Meadow Steppe |
| Ni, <i>et al.</i> [40]   | China          | 45.56  | 88.96            | 146.47                  |
| Ma, <i>et al.</i> [41]   | Inner Mongolia | 56.5   | 133.4            | 196.7                   |
| Piao, <i>et al.</i> [42] | China          | 43.57  | 91.52            | 144.9                   |
| Fan, <i>et al.</i> [43]  | China          | 111.11   | 151.11           | 182.22                  |

In addition to the studies cited above, several studies have obtained results similar to our results. Jin, *et al.* [38] estimated the average aboveground biomass density of the temperate desert steppe, the temperate steppe and the temperate meadow steppe in in the Xilingol grassland to be 23.1, 55.7 and 98.6  $\text{g}\cdot\text{m}^{-2}$ , respectively. Gao, *et al.* [23] used field-based biomass samples and MODIS time series data sets to establish two empirical models based on the relationship of the normalized difference vegetation index (NDVI) to aboveground biomass in the Xilingol grasslands of northern China. The results showed that the average aboveground biomass densities for temperate desert-steppe, temperate steppe and temperate meadow-steppe were 21.2, 59.6 and 111.3  $\text{g}\cdot\text{m}^{-2}$ , respectively. The field biomass measurements obtained at the same time and from the same regions contributed to the consistency of these results. However, several differences remained among studies as a result of the use of different remotely sensed data as input parameter and different regression models. The MODIS productivity data is a global land vegetation net productivity product calculated by model, differing from the vegetation

index data. Using a remote sensing vegetation index to estimate aboveground biomass provides accurate and rapid results, but the limitations of the vegetation index itself may affect the obtained grassland biomass estimates. Compared with NDVI data, MODIS productivity data can overcome the problem of high error rates in low-coverage grasslands and the oversaturation problem in high-coverage grasslands. In addition, MODIS productivity data can better reflect the effects of environmental stresses. However, the method using MODIS productivity data for biomass estimation should be improved to achieve more accurate estimates of grassland biomass in the future.

## 5. Conclusions

This study took the Xilingol grassland as a case study and used ground survey data and MODIS productivity data for the growing seasons of the years 2005–2012 to build a unitary linear regression model to retrieve the aboveground biomass of the Xilingol grassland, and then analyze the spatial and temporal distribution of aboveground biomass. The grassland aboveground biomass averaged 14.35 Tg in the Xilingol grassland during the years 2005–2012, and showed a gradually decreasing trend from east to the west. There were large interannual variations in the aboveground biomass, ranging from 10.56–17.54 Tg, and the aboveground biomass showed differences among different grassland types.

The study made up for the inadequacy of vegetation index to estimate the grassland biomass, and provided an improvement for grassland biomass estimation. Although the study showed promising results for remote sensing based grassland biomass estimation, there are limitations to the accuracy of biomass inversion using MODIS productivity data, therefore, further work is needed to improve the estimation accuracy. In addition, spatio-temporal patterns of aboveground biomass and their relationships with climate factors still need further research in the Xilingol grassland. Gao, *et al.* [23] have already studied the relationship between GSP (growing season total precipitation) with grassland biomass and the relationship between GST (growing season average temperature) with grassland biomass. However, the climate factors are much more than these. For example, how the growing season maximum and minimum temperatures affect grassland biomass and whether the temperature in winter affects grassland biomass need in-depth research. Therefore, the grassland response to climate change is complex, and deserves more detailed and deeper inquiry in future research.

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## Author Contributions

All authors contributed extensively to the work. Fen Zhao, Bin Xu designed and performed experiments. Bin Xu and Yunxiang Jin reviewed the manuscript and gave comments and suggestions to

the manuscript. Lang Xia performed the satellite datasets preprocessing. All authors performed the field survey. All authors participated in editing and revision of the paper.

### Conflicts of Interest

The authors declare no conflict of interest.

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