



Article Efficient Disposal of Rhodamine 6G and Acid Orange 10 Dyes from Aqueous Media Using ZrO₂/CdMn₂O₄/CdO as Novel and Facilely Synthesized Nanocomposites

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Abstract: It is essential to remove rhodamine 6G and acid orange 10 dyes from contaminated water because they can induce cancer and irritate the lungs, skin, mucous, membranes, and eyes. Hence, in the current work, the Pechini sol-gel method was used for the facile synthesis of ZrO₂/CdMn₂O₄/CdO as novel nanocomposites at 600 and 800 °C. The synthesized nanocomposites were used as novel adsorbents for the efficient removal of rhodamine 6G and acid orange 10 dyes from aqueous media. The nanocomposites, which were synthesized at 600 and 800 °C, were abbreviated as EK600 and EK800, respectively. The synthesized nanocomposites were characterized by EDS, XRD, N₂ adsorption/desorption analyzer, and FE-SEM. The patterns of XRD showed that the average crystal size of the EK600 and EK800 nanocomposites is 68.25 and 85.32 nm, respectively. Additionally, the images of FE-SEM showed that the surface of the EK600 nanocomposite consists of spherical, polyhedral, and rod shapes with an average grain size of 99.36 nm. Additionally, the surface of the EK800 nanocomposite consists of polyhedral and spherical shapes with an average grain size of 143.23 nm. In addition, the BET surface area of the EK600 and EK800 nanocomposites is 46.33 and 38.49 m²/g, respectively. The optimal conditions to achieve the highest removal of rhodamine 6G and acid orange 10 dyes are pH = 8, contact time = 24 min, and temperature = 298 kelvin. The greatest removal capacity of the EK600 and EK800 adsorbents towards rhodamine 6G dye is 311.53 and 250.63 mg/g, respectively. Additionally, the greatest removal capacity of the EK600 and EK800 adsorbents towards acid orange 10 dye is 335.57 and 270.27 mg/g, respectively. The removal of rhodamine 6G and acid orange 10 dyes using the EK600 and EK800 adsorbents is spontaneous, exothermic, follows the Langmuir adsorption isotherm, and fits well with the pseudo-first-order kinetic model.

Keywords: nanostructures; adsorption; rhodamine 6G dye; acid orange 10 dye

1. Introduction

The pollution of the aquatic environment is currently one of the most significant and critical issues around the globe [1–5]. The discharge of hazardous compounds, such as



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Copyright: © 2023 by the authors. Licensee MDPI, Basel, Switzerland. This article is an open access article distributed under the terms and conditions of the Creative Commons Attribution (CC BY) license (https:// creativecommons.org/licenses/by/ 4.0/). synthetic dyes and heavy metal ions, from industrial effluents is the primary cause of water contamination [6–9]. Dyes are complex organic substances that were once manufactured from coal ash but are now manufactured from petroleum. What distinguishes a dye from a pigment is that pigments are typically insoluble. Today, diverse synthetic dyes are manufactured to satisfy the needs of various industries, such as adhesives, paper and pulp, plastics, construction glass, ceramics, wax biomedicine, paints, beverages, art supplies, cosmetics, pharmaceuticals, leather, soap, polymers, and food [10-12]. Despite the availability of natural colorants, many manufacturers opt for synthetic dyes that pose harmful health risks. One of the issues with the majority of natural dyes is the need for mordant. Natural dyes are more expensive and scarce, and their production is more time-consuming and complicated. With natural dyes, the color palette is more restricted. According to scientists, organic dyes are carcinogenic and toxic and contribute to anxiety migraines, hyperactivity, and behavioral disorders [13–15]. It is undesirable to discharge synthetic dyes into aquatic environments due to the adverse effects they may have on aquatic organisms [16-20]. The acid orange 10 dye (orange G dye) is one of the most widely used anionic dyes in the printing and textile industries. It can induce cancer and irritate the lungs, skin, mucous, membranes, and eyes [21]. The rhodamine 6G dye is one of the most widely used cationic dyes in the food and textile industries. In addition, consuming water contaminated with rhodamine has been shown to cause subcutaneous sarcoma, which is highly neurotoxic and carcinogenic [22]. There have been several studies on dye removal techniques, such as chemical coagulation, reverse osmosis, nanofiltration, electrochemical treatment, catalysis, and adsorption [23–34]. However, adsorption is one of the most frequently used methods. Adsorption is a technique for collecting soluble components of an aqueous solution on a suitable surface without producing deleterious byproducts [35–37]. Adsorption can be employed for a wide range of substances, including liquids, gases, and dissolved ions. This versatility makes it applicable in various fields, such as environmental remediation, gas separation, water purification, and chemical processing. Many adsorbents, such as zeolites and activated carbon, have high surface areas with numerous pores and active sites [38–42]. This provides a large contact area for adsorbate molecules, allowing for efficient adsorption. The materials used as adsorbents are often inexpensive and readily available, making adsorption a cost-effective method for purification and separation processes. In many cases, adsorbents can be regenerated and reused multiple times, which can lead to cost savings and reduced waste generation compared to other separation methods. Compared to some other separation methods like distillation, adsorption processes can be energy-efficient, especially at lower temperatures and pressures. There are many adsorbents that have been used to remove dyes, such as zinc ferrite, chitosan/pyrazole/zinc oxide, cobalt oxide, carboxymethylcellulose magnetic, zeolites, and benzenesulfonyl hydrazone modified guar gum [43–48]. Gollakota et al. synthesized aluminophosphate for the removal of rhodamine 6G dye, and its maximum adsorption capacity equaled 208.11 mg/g [49]. Vanamudan et al. synthesized chitosan-g-(N-vinylpyrrolidone)/montmorillonite composite for the removal of rhodamine 6G dye, and its maximum adsorption capacity equaled 36.60 mg/g [50]. Annadurai et al. synthesized activated carbon for the removal of rhodamine 6G dye, and its maximum adsorption capacity equaled 44.70 mg/g [51]. Gollakota et al. synthesized ZSM-22 zeolite from Taiwanese coal fly ash for the removal of rhodamine 6G dye, and its maximum adsorption capacity equaled 195.30 mg/g [52]. Zheng et al. synthesized quaternary ammonium group-rich magnetic nanoparticles for the removal of acid orange 10 dye, and their maximum adsorption capacity equaled 98.70 mg/g [53]. Arulkumar et al. synthesized activated carbon for the removal of acid orange 10 dye, and its maximum adsorption capacity equaled 17.60 mg/g [54]. Banerjee et al. synthesized alumina nanoparticles for the removal of acid orange 10 dye, and their maximum adsorption capacity equaled 93.30 mg/g [55]. Imgharn et al. synthesized polyaniline/Fe-ZSM-5 composite for the removal of acid orange 10 dye, and its maximum adsorption capacity equaled 217 mg/g [56]. Due to the small crystalline size of nanomaterials and their high surface area, they are widely used in the removal of several dyes. The pechini sol gel method was

widely used to synthesize a lot of nanostructures such as Al₂O₃, PbCrO₄, Y₂O₃/MgO, and $Bi_2Ti_2O_7$ [57–60]. High prices of chemicals can limit the scalability and affordability of these adsorbents, particularly in developing countries or resource-constrained settings. As a result, researchers and scientists are actively exploring alternatives and low-cost materials that can serve as effective adsorbents while reducing overall operational expenses. In the current paper, $ZrO_2/CdMn_2O_4/CdO$ was facilely synthesized for the first time as a novel nanocomposite for the efficient removal of rhodamine 6G and acid orange 10 dyes from aqueous solutions. Three key characteristics of synthesized nanocomposites are stability, small crystal size, and high surface area, which play significant roles in enhancing their adsorption capabilities. Also, the preparation of $ZrO_2/CdMn_2O_4/CdO$ nanocomposite offers several advantages, making it a promising choice for water treatment. The combination of these three components results in a synergistic effect that enhances the overall performance of the nanocomposite. The incorporation of ZrO₂, CdO, and CdMn₂O₄ in the nanocomposite provides a larger number of active adsorption sites. This combination leads to an increased adsorption capacity, allowing for more efficient removal of rhodamine 6G and acid orange 10 dyes from aqueous media. The synergistic effect between these components enhances the overall adsorption performance of the nanocomposite. The synthesized nanocomposite was characterized by EDS, XRD, N₂ adsorption/desorption analyzer, and FE-SEM. The effects of dye solution pH, adsorbent/adsorbate contact time, temperature, and dye concentration on the removal of rhodamine 6G and acid orange 10 dyes were also studied.

2. Experimental Section

2.1. Chemicals

The utilized chemicals were cadmium(II) nitrate tetrahydrate (Cd(NO₃)₂·4H₂O), manganese(II) acetate tetrahydrate (Mn(CH₃COO)₂·4H₂O), sodium hydroxide (NaOH), zirconyl chloride octahydrate (ZrOCl₂·8H₂O), citric acid (C₆H₈O₇), ethylene glycol (C₂H₆O₂), hydrochloric acid (HCl), rhodamine 6G dye (C₂₈H₃₁N₂O₃Cl), and acid orange 10 dye (C₁₆H₁₀N₂Na₂O₇S₂). All the aforementioned chemicals were of high purity, purchased from the Sigma-Aldrich company, and used without any purification.

2.2. Synthesis of Adsorbents

Three solutions of zirconium, cadmium, and manganese ions were freshly prepared with a molar ratio of Zr/Cd/Mn equal to 1:1:1. The solution of zirconium ions was prepared via dissolving 12 g of ZrOCl₂·8H₂O in 80 mL of deionized water. In addition, the solution of cadmium ions was prepared by dissolving 11.49 g of Cd(NO₃)₂·4H₂O in 80 mL of deionized water. In addition, the solution of manganese ions was prepared by dissolving 9.13 g of Mn(CH₃COO)₂·4H₂O in 80 mL of deionized water. Moreover, a solution of citric acid was prepared by dissolving 21.46 g of citric acid in 100 mL of deionized water. After that, solutions of zirconium, cadmium, and manganese metal ions were mixed with each other and stirred at 560 rpm for 10 min. The citric acid solution is then added to the mixed metal ion solution with continuous stirring at 560 rpm for 10 min. In addition, 12 mL of ethylene glycol was added with continuous stirring at 560 rpm at 140 °C until all the solvent was evaporated. Finally, the resulting powder is collected, ground, and then calcinated at 600 and 800 °C to obtain the mixed metal oxide nanostructures. The products, which were obtained by calcination at 600 and 800 °C, were labeled as EK600 and EK800, respectively. The practical steps for synthesizing the EK600 and EK800 nanocomposites are shown in Scheme 1.



Scheme 1. The practical steps for synthesizing the EK600 and EK800 nanocomposites.

2.3. Characterization of Adsorbents

Patterns of X-ray diffraction (XRD) of EK600 and EK800 nanocomposites were obtained using a Bruker D8 Advance diffractometer with 0.15 nm CuK α radiation. Surface images and compositions of EK600 and EK800 samples were obtained using the JSM-IT800 Schottky field emission scanning electron microscope (FE-SEM) coupled with energy dispersive X-ray spectroscopy (EDS). A Quantachrome N₂ adsorption/desorption analyzer was used to determine the BET surface area in addition to the pore features of the EK600 and EK800 samples.

2.4. Removal of Rhodamine 6G and Acid Orange 10 Dyes from Aqueous Solutions

A total of 0.05 g of EK600 or EK800 adsorbent was mixed with 150 mL of a 150 mg/L aqueous solution of rhodamine 6G or acid orange 10 dyes and then stirred at 550 rpm for 180 min in order to investigate the effect of solution pH ranging from 2.5 to 8.5. The adsorbent was then removed from the mixture by centrifuging it at 2550 rpm for 4 min. After centrifugation, the produced filtrate was analyzed with a Shimadzu 1800 UV–Vis spectrophotometer at 524 and 485 nm to determine the equilibrium concentration of rhodamine 6G and acid orange 10 dyes, respectively. Furthermore, the effects of time (in the range of 4–40 min), temperature (in the range of 298–328 kelvin), and concentration (in the range of 15–250 mg/L) were also studied.

The % removal of rhodamine 6G or acid orange 10 dyes (% U) and the adsorption capacity of EK600 and EK800 adsorbents (O, mg/g) were determined by Equations (1) and (2), respectively.

$$\%U = \frac{Z_o - Z_e}{Z_o} \times 100\tag{1}$$

$$O = (Z_o - Z_e) \times \frac{V}{M}$$
⁽²⁾

 Z_o and Z_e are the original and equilibrium concentrations of rhodamine 6G or acid orange 10 dyes (mg/L), respectively. Moreover, M and V are the amount of adsorbent (g) and the solution volume of rhodamine 6G or acid orange 10 dyes (L), respectively. The practical steps for removing the rhodamine 6G and acid orange 10 dyes using the EK600 and EK800 nanocomposites are shown in Scheme 2.

Adsorption experiments



Scheme 2. The practical steps for removing the rhodamine 6G and acid orange 10 dyes using the EK600 and EK800 nanocomposites.

3. Results and Discussion

3.1. Synthesis and Characterization of the Synthesized Samples

The Pechini sol–gel method was utilized to produce a $ZrO_2/CdMn_2O_4/CdO$ nanocomposite. To achieve this, zirconyl citrate, cadmium citrate, and manganese citrate were derived from citric acid reacting with zirconyl chloride octahydrate, cadmium(II) nitrate tetrahydrate, and manganese(II) acetate tetrahydrate. In this process, ethylene glycol plays a crucial role as a control agent for polymerization and contributes to gel formation by reacting with citric acid. The resultant three-dimensional gel network serves as a template for the final nanocomposite, which is obtained through calcination at 600 and 800 °C. To obtain the desired pure $ZrO_2/CdMn_2O_4/CdO$ nanocomposite, it is necessary to fully decompose and eliminate the organic constituents. The temperatures selected for this process, 600 °C and 800 °C, appear to be appropriate based on the thermogravimetric analysis of the resulting powder prior to calcination (Figure 1).



Figure 1. The thermogravimetric analysis of the resulting powder before calcination.

3.1.1. XRD

Figure 2A,B illustrate the XRD patterns of the EK600 and KE800 nanocomposites, respectively. Also, the results showed that the EK600 and EK800 samples consist of zirconium oxide (ZrO₂), cadmium manganese oxide (CdMn₂O₄), and cadmium oxide (CdO) nanostructures as elucidated from JCPDS Nos. 00-007-0343, 00-023-0826, and 01-075-0592, respectively. The peaks at $2\theta = 17.52^{\circ}$, 24.09°, 24.42°, 34.34°, 41.02°, 45.62°, 49.21°, 50.36°, 57.29°, 57.91°, 58.66°, 59.92°, 62.02°, 62.96°, and 64.11° are due to the (100), (011), (110), (002), (-102), (202), (022), (122), (310), (131), (222), (-131), (213), (-311), and (023) miller indices of ZrO₂, respectively. Also, the peaks at $2\theta = 28.45^{\circ}$, 30.33°, 31.45°, 35.58°, 43.96°, 44.67°, 47.95°, 48.57°, 53.19°, 54.14°, 61.27°, and 68.09° are due to the (112), (200), (103), (211), (220), (213), (301), (204), (312), (303), (116), and (305) miller indices of CdMn₂O₄, respectively. In addition, the peaks at $2\theta = 33.12^{\circ}$, 38.50°, 55.30°, 65.90°, and 69.15° are due to the (111), (200), (220), (311), and (222) miller indices of CdO, respectively. Moreover, the average crystal size of the EK600 and EK800 nanocomposites is 68.25 and 85.32 nm, respectively. Hence, the compositions of EK600 and EK800 nanocomposites are the same, but the average crystallite size is different. This can be seen from the difference in intensity of the XRD peaks in the two samples, with no difference in their position. In the Pechini sol–gel method, the size of the resulting particles can increase with increasing calcination temperatures due to several underlying processes that take place during the calcination step. The calcination stage is a crucial stage in which the dried gel precursor is heated to remove the organic components and promote the formation of the desired crystalline phase. At elevated temperatures, particles start to undergo sintering, which is the process of grain boundary migration and coalescence of adjacent particles. Sintering causes the particles to fuse together, resulting in the growth of larger crystallites and an overall increase in particle size. Also, during calcination, organic solvents and other volatiles are driven off as gases from the gel. As these volatiles escape, voids or pores may be formed within the structure, leading to an increase in particle size. In addition, during calcination, particles may agglomerate or cluster together due to the increased diffusion and mobility of atoms at higher temperatures. These agglomerates can contribute to the observed increase in particle size.

3.1.2. FE-SEM and EDS

Figure 3A,B illustrate the images of FE-SEM for the EK600 and KE800 nanocomposites, respectively. Also, the results showed that the surface of the EK600 sample consists of spherical, polyhedral, and rod shapes with an average grain size of 99.36 nm. In addition, the surface of the EK800 sample consists of polyhedral and spherical shapes with an average grain size of 143.23 nm. The average crystal sizes of nanocomposites obtained from X-ray diffraction (XRD) and scanning electron microscopy (SEM) may not be the same due to the different principles and measurement techniques involved in each method. XRD measures the crystallite size in the bulk material, considering all crystallographic directions. SEM, on the other hand, provides information about the surface morphology, which is often an agglomeration of many particles and hence may not directly correlate to the crystal size obtained from XRD. Figure 4A,B illustrate the patterns of EDS for the EK600 and EK800 nanocomposites, respectively. Also, the results showed that the EK600 and EK800 nanocomposites consist of O, Mn, Cd, and Zr, as shown in Table 1. The absence of peaks for other elements confirms the purity of the synthesized nanocomposites. The mass ratios of O (oxygen), Mn (manganese), Cd (cadmium), and Zr (zirconium) from energy-dispersive X-ray spectroscopy (EDS) analysis are not consistent with the expected compositions of "ZrO₂/CdMn₂O₄/CdO due to sample inhomogeneity and different percentages of each component inside the nanocomposite. The XRD confirmed that the percentages of ZrO₂, CdMn₂O₄, and CdO inside the EK600 nanocomposite are 74.30, 24.20, and 1.5%, respectively. In addition, the XRD confirmed that the percentages of ZrO₂, CdMn₂O₄, and CdO inside the EK800 nanocomposite are 78.00, 18.50, and 3.5%, respectively. If the nanocomposite is not homogeneous, different regions within the sample might have varying elemental compositions. EDS analysis provides local elemental information, and the results may not reflect the overall average composition of the entire nanocomposite.



Figure 2. Patterns of XRD for the EK600 (A) and EK800 (B) products.





Figure 3. Images of FE-SEM for the EK600 (A) and EK800 (B) products.



Figure 4. Patterns of EDS for the EK600 (A) and EK800 (B) products.

Samples	% O	% Mn	% Cd	% Zr
EK600	14.68	21.90	25.71	37.71
EK800	13.98	21.64	27.43	36.95

Table 1. EDS analysis of the EK600 and EK800 samples.

3.1.3. Nitrogen Adsorption/Desorption

Figure 5A,B illustrate the relation between adsorbed volume and relative pressure for the EK600 and EK800 samples, respectively. In addition, Figure 6A,B illustrate the BJH pore size distribution of the EK600 and EK800 nanocomposites, respectively. Additionally, the results showed strongly that the resulting curves follow the IV types, and this confirms their mesoporous nature. The surface textures, such as BET surface area, total pore volume, and average pore size, are clarified in Table 2. The BET surface area of the EK600 nanocomposite is greater than that of the EK800 nanocomposite because of the inverse relationship between crystal size and surface area.

Sample	BET Surface Area (m²/g)	Total Pore Volume (cc/g)	Average Pore Size (nm)
EK600	46.33	0.2037	8.79
EK800	38.49	0.1398	7.27

Table 2. Surface textures of the EK600 and EK800 samples.

3.2. *Removal of Rhodamine 6G and Acid Orange 10 Dyes from Aqueous Solutions* 3.2.1. Effect of pH

Figure 7A,B represent the effect of solution pH on the % U of rhodamine 6G and acid orange 10 dyes in addition to the removal capacity of the EK600 and EK800 adsorbents, respectively. In the case of rhodamine 6G dye, it was found that the % U of rhodamine 6G dye and the removal capacity of the EK600 and EK800 adsorbents increased with increasing pH from 2.5 to 8.5. At pH 8.5, the % U of rhodamine 6G dye using EK600 and EK800 adsorbents is 66.53 and 53.85%, respectively. At pH 8.5, the removal capacity of the EK600 and EK800 adsorbents towards rhodamine 6G dye is 299.40 and 242.34 mg/g, respectively. The acidic medium produces positive hydrogen ions that surround the surface of the EK600 and EK800 adsorbents and thus repel the cationic rhodamine 6G dye as shown in Scheme 3, and thus the % U and adsorption capacity decrease [61]. The basic medium produces negative hydroxide ions that surround the surface of the EK600 and EK800 adsorbents and thus attract the cationic rhodamine 6G dye as shown in Scheme 3, and thus the % U and adsorption capacity increase. Hence, the optimum pH value for the removal of rhodamine 6G dye is 8.5. In the case of acid orange 10 dye, it was found that the % U of acid orange 10 dye and the removal capacity of the EK600 and EK800 adsorbents decreased with increasing pH from 2.5 to 8.5. At pH 2.5, the % U of acid orange 10 dye using EK600 and EK800 adsorbents is 73.19 and 57.71%, respectively. At pH 2.5, the removal capacity of the EK600 and EK800 adsorbents towards acid orange 10 dye is 329.34 and 259.71 mg/g, respectively. The acidic medium produces positive hydrogen ions that surround the surface of the EK600 and EK800 adsorbents and thus attract the anionic acid orange 10 dye as shown in Scheme 3, and thus the % U and adsorption capacity increase. The basic medium produces negative hydroxide ions that surround the surface of the EK600 and EK800 adsorbents and thus repel the anionic acid orange 10 dye as shown in Scheme 3, and thus the % U and adsorption capacity decrease [61]. Hence, the optimum pH value for the removal of acid orange 10 dye is 2.5.



Figure 5. The relation between adsorbed volume and relative pressure for the EK600 (**A**) and EK800 (**B**) samples.



Figure 6. The BJH pore size distribution of the EK600 (A) and EK800 (B) samples.



Figure 7. The effect of solution pH on the % U of rhodamine 6G and acid orange 10 dyes (**A**) in addition to the removal capacity of the EK600 and EK800 adsorbents (**B**).



Acid orange 10 dye (Anionic dye)

Scheme 3. Removal mechanism of rhodamine 6G and acid orange 10 dyes using the EK600 and EK800 adsorbents.

To confirm the adsorption mechanism, FE-SEM analysis was performed for the EK600 sample (as an illustrative example) after the adsorption of rhodamine 6G and acid orange 10 dyes, as shown in Figure 8A,B, respectively. In addition, the FE-SEM analysis of the EK600 sample is shown in Figure 3A. Figure 8A,B clearly show the disappearance of the spherical, polyhedral, and rod shapes of the EK600 adsorbent. Additionally, Figure 8A,B clearly show the aggregation of rhodamine 6G and acid orange 10 dyes on the surface of EK600 adsorbent, and this indicates that rhodamine 6G and acid orange 10 dyes were successfully adsorbed on the surface of EK600 adsorbent.

Moreover, the inductively coupled plasma (ICP) elucidated that there is no ion leaching from the synthesized EK600 and EK800 nanocomposites in the filtrate during the adsorption processes. Hence, the synthesized nanocomposites are considered safe and can be utilized in water treatment.

3.2.2. Effect of Contact Time

Figure 9A,B represent the effect of contact time on the % U of rhodamine 6G and acid orange 10 dyes in addition to the removal capacity of the EK600 and EK800 adsorbents, respectively. In addition, the results exhibited that the % U of rhodamine 6G and acid orange 10 dyes, in addition to the removal capacity of the EK600 and EK800 adsorbents, increased when increasing the contact time from 4 to 24 min. In addition, the % U of rhodamine 6G and EK800 adsorbents, were almost constant when increasing the contact time from 24 to 32 min as a result of the fullness saturation of active sites. A sudden increase in the adsorption percentage when increasing the contact time from 20 min to 25 min is due to the activation of adsorption sites of the adsorbent, which is a mixture of several metal oxides. Adsorption typically occurs on the surface of the adsorbent material, which contains specific sites where adsorption can take place. If these adsorption sites are not immediately available for interaction with the adsorbate, the adsorption rate may initially be slow. However, with

increasing contact time, these sites may become activated or exposed, leading to a sudden increase in the adsorption percentage.



Figure 8. FE-SEM analysis of the EK600 sample after the adsorption of rhodamine 6G (**A**) and acid orange 10 (**B**) dyes.



Figure 9. The effect of contact time on the % U of rhodamine 6G and acid orange 10 dyes (**A**) in addition to the removal capacity of the EK600 and EK800 adsorbents (**B**).

The % U of rhodamine 6G dye by the EK600 and EK800 adsorbents after 24 min is 67.67 and 53.85%, respectively. Additionally, the removal capacity of the EK600 and EK800 adsorbents towards rhodamine 6G dye after 24 min is 304.50 and 242.34 mg/g, respectively. In addition, the % U of acid orange 10 dye by the EK600 and EK800 adsorbents after 24 min is 73.30 and 57.95%, respectively. Additionally, the removal capacity of the EK600 and EK800 adsorbents towards acid orange 10 dye after 24 min is 329.85 and 260.76 mg/g, respectively.

The pseudo-first-order (Equation (3)), in addition to the pseudo-second-order (Equation (4)), kinetic equations [61], were used to examine the linear fitting of the obtained experimental data due to the removal of rhodamine 6G and acid orange 10 dyes by the EK600 and EK800 adsorbents, as shown in Figure 10A,B, respectively.

$$\log(O_e - O_t) = \log O_e - \frac{k_1}{2.303}t$$
(3)

$$\frac{t}{O_t} = \frac{1}{k_2 O_e^2} + \frac{1}{O_e} t$$
(4)

 k_1 represents the pseudo-first-order rate constant (1/min), O_t represents the amount of rhodamine 6G or acid orange 10 dye adsorbed at time t (mg/g), k_2 represents the pseudo-second-order rate constant (g/mg.min), and O_e represents the quantity of rhodamine 6G or acid orange 10 dye adsorbed at equilibrium (mg/g). Table 3 displays the resultant constants and correlation coefficients (R^2). Moreover, the removal of rhodamine 6G and acid orange 10 dyes by the EK600 and EK800 adsorbents follows the pseudo-first order owing to the following reasons: (1) The R^2 values of the pseudo-first order are greater than the R^2 values of pseudo-second order. (2) The calculated O_e values of the pseudo-first order are very close to the experimental O_e .

Table 3. The obtained kinetic constants for the removal of rhodamine 6G and acid orange 10 dyes by the EK600 and EK800 adsorbents.

	Pseudo-First Order			Pseudo-Second Order		
Conditions	k ₁ (1/min)	O _e (mg/g)	R ²	k ₂ (g/mg.min)	O _e (mg/g)	R ²
EK600+ Rhodamine 6G	0.1254	313.60	0.9999	0.00022	429.18	0.9955
EK800+ Rhodamine 6G	0.1081	250.93	0.9996	0.00019	371.75	0.9864
EK600+ Acid orange 10	0.1227	342.68	0.9999	0.00018	478.47	0.9927
EK800+ Acid orange 10	0.1024	265.35	0.9999	0.00018	392.16	0.9911

3.2.3. Effect of Temperature

Figure 11A,B represent the effect of temperature on the % U of rhodamine 6G and acid orange 10 dyes in addition to the removal capacity of the EK600 and EK800 adsorbents, respectively. Additionally, the results exhibited that the % U of rhodamine 6G and acid orange 10 dyes, in addition to the removal capacity of the EK600 and EK800 adsorbents, decreased when increasing the temperature from 298 to 328 kelvin. The % U of rhodamine 6G dye by the EK600 and EK800 adsorbents at 328 kelvin is 50.13 and 37.83%, respectively. Additionally, the removal capacity of the EK600 and EK800 adsorbents towards rhodamine 6G dye at 328 kelvin is 225.57 and 170.25 mg/g, respectively. In addition, the % U of acid orange 10 dye by the EK600 and EK800 adsorbents at 328 kelvin is 58.34 and 42.93%, respectively. Additionally, the removal capacity of the EK600 and EK800 adsorbents towards acid orange 10 dye at 328 kelvin is 262.53 and 193.17 mg/g, respectively.



Figure 10. Pseudo-first-order (**A**) as well as pseudo-second-order (**B**) plots for the removal of rhodamine 6G and acid orange 10 dyes by the EK600 and EK800 adsorbents.



Figure 11. The effect of dye solution temperature on the % U of rhodamine 6G and acid orange 10 dyes (**A**) in addition to the removal capacity of the EK600 and EK800 adsorbents (**B**).

The effect of solution temperature on the removal of rhodamine 6G and acid orange 10 dyes by the EK600 and EK800 adsorbents was examined using the thermodynamic constants, for example, ΔG° (change in free energy, kJ/mol), ΔH° (change in enthalpy, kJ/mol), and ΔS° (change in entropy, kJ/mol kelvin), which were estimated by Equations (5)–(7) [61].

$$lnK_T = \frac{\Delta S^{\circ}}{R} - \frac{\Delta H^{\circ}}{RT}$$
(5)

$$\Delta G^{\circ} = \Delta H^{\circ} - T \Delta S^{\circ} \tag{6}$$

$$K_T = \frac{O_e}{Z_e} \tag{7}$$

R represents the universal gas constant (kJ/mol kelvin), whereas T represents the absolute temperature (kelvin). K_T represents the distribution constant (L/g). In addition, Figure 12 depicts the linear variation of lnK_T against 1/T, from which the thermodynamic constants can be estimated based on intercept and slope. Table 4 displays the thermodynamic constants for the removal of rhodamine 6G and acid orange 10 dyes by the EK600 and EK800 adsorbents. Negative ΔG° values indicated that the EK600 and EK800 adsorbents removed the rhodamine 6G and acid orange 10 dyes spontaneously. In addition, the negative ΔH° values indicated that the EK600 adsorbents remove the rhodamine 6G and acid orange 10 dyes were removed the rhodamine 6G and acid orange 10 dyes were removed by the EK600 and EK800 adsorbents in the direction of increasing system randomization. Moreover, the removal of rhodamine 6G and acid orange 10 dyes by EK600 and EK800 adsorbents is physical in nature, as the ΔH° values are less than 40 kJ/mol [61].



Figure 12. The plot of $\ln K_T$ against 1/T for the removal of rhodamine 6G and acid orange 10 dyes by the EK600 and EK800 adsorbents.

Conditions	$\Delta \mathrm{H}^{\circ}$ (kJ/mol)	ΔS°	ΔG° (kJ/mol)			
	-	(KJ/MOI Kelvin)	298	308	318	328
EK600+ Rhodamine 6G	-19.95	0.0514	-35.26	-35.78	-36.29	-36.81
EK800+ Rhodamine 6G	-18.07	0.0501	-33.00	-33.50	-34.00	-34.51
EK600+ Acid orange 10	-18.76	0.0453	-32.26	-32.71	-33.16	-33.62
EK800+ Acid orange 10	-16.18	0.0423	-28.79	-29.22	-29.64	-30.06

Table 4. The thermodynamic constants for the removal of rhodamine 6G and acid orange 10 dyes by the EK600 and EK800 adsorbents.

3.2.4. Effect of Dye Concentration

Figure 13A,B represent the effect of dye concentration on the % U of rhodamine 6G and acid orange 10 dyes in addition to the removal capacity of the EK600 and EK800 adsorbents, respectively. Additionally, the results exhibited that the % U of rhodamine 6G and acid orange 10 dyes decreased with increasing the concentration of dyes from 15 to 250 mg/L. Also, the removal capacity of the EK600 and EK800 adsorbents improved by increasing the initial concentration of dyes from 15 to 250 mg/L.

The Langmuir (Equation (8)) and Freundlich (Equation (9)) equilibrium equations [61] were used to examine the linear fitting of the obtained experimental data due to the removal of rhodamine 6G and acid orange 10 dyes by the EK600 and EK800 adsorbents, as indicated in Figure 14A,B, respectively.

$$\frac{Z_e}{O_e} = \frac{1}{k_3 O_{max}} + \frac{Z_e}{O_{max}} \tag{8}$$

$$lnO_e = lnk_4 + \frac{1}{y}lnZ_e \tag{9}$$

where 1/y is the heterogeneity constant while k_3 is the equilibrium constant of the Langmuir equation (L/mg). Also, k_4 is the equilibrium constant of the Freundlich equation $(mg/g)(L/mg)^{1/n}$, while O_{max} is the greatest removal capacity of the Langmuir equation (mg/g). Additionally, Equation (10) can be utilized to determine the O_{max} from the Freundlich equilibrium isotherm [61].

$$O_{max} = k_4 \left(Z_o^{1/y} \right) \tag{10}$$

Table 5 displays the resultant constants and correlation coefficients (R^2). In addition, the removal of rhodamine 6G and acid orange 10 dyes by the EK600 and EK800 adsorbents follows the Langmuir isotherm because the R^2 values of the Langmuir isotherm are higher than the R^2 values of the Freundlich isotherm. Additionally, the greatest removal capacity of the EK600 and EK800 adsorbents toward rhodamine 6G dye is 311.53 and 250.63 mg/g, respectively. Moreover, the greatest removal capacity of the EK600 and EK800 adsorbents towards acid orange 10 dye is 335.57 and 270.27 mg/g, respectively.



Figure 13. The effect of concentration on the % U of rhodamine 6G and acid orange 10 dyes (**A**) in addition to the removal capacity of the EK600 and EK800 adsorbents (**B**).



Figure 14. Langmuir (**A**) and Freundlich (**B**) plots for the removal of rhodamine 6G and acid orange 10 dyes by the EK600 and EK800 adsorbents.

Langmuir Isotherm				Freundlich Isotherm		
Conditions	O _{max} (mg/g)	k ₃ (L/mg)	R ²	O _{max} (mg/g)	k ₄ (mg/g)(L/mg) ^{1/n}	R ²
EK600+ Rhodamine 6G	311.53	2.1837	0.9999	386.22	136.07	0.7504
EK800+ Rhodamine 6G	250.63	0.6196	0.9998	283.51	91.37	0.9159
EK600+ Acid orange 10	335.57	1.9477	0.9999	431.32	136.44	0.7905
EK800+ Acid orange 10	270.27	0.4302	0.9995	313.67	85.23	0.9118

Table 5. The obtained equilibrium constants for the removal of rhodamine 6G and acid orange 10 dyes by the EK600 and EK800 adsorbents.

Additionally, the removal capacity of the EK600 and EK800 synthesized nanocomposites towards rhodamine 5G and acid orange 10 dyes was superior to that of many other adsorbents, as shown in Table 6 [49–56].

Table 6. Comparison between the greatest removal capacity of the synthesized nanocomposites and that of other adsorbents in the literature.

Adsorbent	O(mg/g) towards Rhodamine 6G Dye	O(mg/g) towards Acid Orange 10 Dye	Ref.
Aluminophosphate	208.11		[49]
Chitosan-g-(N-vinylpyrrolidone)/ montmorillonite composite	36.60		[50]
Activated carbon	44.70		[51]
ZSM-22 zeolite	195.30		[52]
Quaternary ammonium group-rich magnetic nanoparticles		98.70	[53]
Activated carbon		17.60	[54]
Alumina nanoparticles		93.30	[55]
Polyaniline/Fe-ZSM-5 composite		217	[56]
EK600	311.53	335.57	This study
EK800	250.63	268.09	This study

3.2.5. Effect of Desorption and Reusability

The EK600 and EK800 adsorbents were regenerated by subjecting them to a heat treatment at 550 °C for 2 hr, which completely removed rhodamine 6G or acid orange 10 dyes from their surfaces. Subsequently, the regenerated adsorbents were employed for four consecutive cycles to remove the studied dyes, following the experimental procedure outlined earlier. Figure 15 demonstrates that the percentage removal of rhodamine 6G or acid orange 10 dyes using the synthesized adsorbents remained fairly constant, indicating that these adsorbents can be reused multiple times without a significant loss in their efficiency.

0

0



Figure 15. The effect of reusability of the EK600 and EK800 adsorbents.

2

Cycle number

4. Conclusions

1

The Pechini sol–gel method was utilized for the facile synthesis of $ZrO_2/CdMn_2O_4/CdO$ as novel nanocomposites at 600 and 800 °C. The nanocomposites, which were synthesized at 600 and 800 °C, were abbreviated as EK600 and EK800, respectively. Additionally, the XRD showed that the average crystal size of the EK600 and EK800 nanocomposites is 68.25 and 85.32 nm, respectively. Moreover, the BET surface area of the EK600 and EK800 nanocomposites is 46.33 and 38.49 m²/g, respectively. The greatest removal capacity of the EK600 and EK800 adsorbents towards acid orange 10 dye is 335.57 and 270.27 mg/g, respectively. Additionally, the greatest removal capacity of the EK600 and EK800 adsorbents toward size of the EK600 and EK800 adsorbents toward rhodamine 6G dye is 311.53 and 250.63 mg/g, respectively.

3

4

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References

- 1. Hasanpour, M.; Hatami, M. Photocatalytic Performance of Aerogels for Organic Dyes Removal from Wastewaters: Review Study. *J. Mol. Liq.* **2020**, 309, 113094. [CrossRef]
- Wazir, M.B.; Daud, M.; Ali, F.; Al-Harthi, M.A. Dendrimer Assisted Dye-Removal: A Critical Review of Adsorption and Catalytic Degradation for Wastewater Treatment. J. Mol. Liq. 2020, 315, 113775. [CrossRef]
- 3. Chowdhury, M.F.; Khandaker, S.; Sarker, F.; Islam, A.; Rahman, M.T.; Awual, M.R. Current Treatment Technologies and Mechanisms for Removal of Indigo Carmine Dyes from Wastewater: A Review. *J. Mol. Liq.* **2020**, *318*, 114061. [CrossRef]
- 4. Kadhom, M.; Albayati, N.; Alalwan, H.; Al-Furaiji, M. Removal of Dyes by Agricultural Waste. *Sustain. Chem. Pharm.* 2020, 16, 100259. [CrossRef]
- 5. Gahlot, R.; Taki, K.; Kumar, M. Efficacy of Nanoclays as the Potential Adsorbent for Dyes and Metal Removal from the Wastewater: A Review. *Environ. Nanotechn. Monit. Manag.* **2020**, *14*, 100339. [CrossRef]
- Zeng, X.; Zhang, G.; Wen, J.; Li, X.; Zhu, J.; Wu, Z. Simultaneous Removal of Aqueous Same Ionic Type Heavy Metals and Dyes by a Magnetic Chitosan/Polyethyleneimine Embedded Hydrophobic Sodium Alginate Composite: Performance, Interaction and Mechanism. *Chemosphere* 2023, 318, 137869. [CrossRef]
- Vasiraja, N.; Saravana Sathiya Prabhahar, R.; Joshua, A. Preparation and Physio–Chemical Characterisation of Activated Carbon Derived from Prosopis Juliflora Stem for the Removal of Methylene Blue Dye and Heavy Metal Containing Textile Industry Effluent. J. Clean. Prod. 2023, 397, 136579. [CrossRef]
- 8. Anastopoulos, I.; Ahmed, M.J.; Hummadi, E.H. Eucalyptus-Based Materials as Adsorbents for Heavy Metals and Dyes Removal from (Waste)Waters. J. Mol. Liq. 2022, 356, 118864. [CrossRef]
- 9. Garba, Z.N.; Lawan, I.; Zhou, W.; Zhang, M.; Wang, L.; Yuan, Z. Microcrystalline Cellulose (MCC) Based Materials as Emerging Adsorbents for the Removal of Dyes and Heavy Metals–A Review. *Sci. Total Environ.* **2020**, *717*, 135070. [CrossRef]
- 10. Ahmadian, M.; Jaymand, M. Interpenetrating Polymer Network Hydrogels for Removal of Synthetic Dyes: A Comprehensive Review. *Coord. Chem. Rev.* 2023, 486, 215152. [CrossRef]
- 11. Bilal, M.; Ihsanullah, I.; Hassan Shah, M.U.; Bhaskar Reddy, A.V.; Aminabhavi, T.M. Recent Advances in the Removal of Dyes from Wastewater Using Low-Cost Adsorbents. *J. Environ. Manag.* **2022**, *321*, 115981. [CrossRef] [PubMed]
- Lan, D.; Zhu, H.; Zhang, J.; Li, S.; Chen, Q.; Wang, C.; Wu, T.; Xu, M. Adsorptive Removal of Organic Dyes via Porous Materials for Wastewater Treatment in Recent Decades: A Review on Species, Mechanisms and Perspectives. *Chemosphere* 2022, 293, 133464. [CrossRef]
- 13. Al Naim, A.F. Mesoporous and Nanoflowers (ZnO₂) via a Hydrothermal Technique for Dye Removal and Antibacterial Applications. *Inorg. Chem. Commun.* **2023**, 151, 110575. [CrossRef]
- 14. Aramesh, N.; Bagheri, A.R.; Bilal, M. Chitosan-Based Hybrid Materials for Adsorptive Removal of Dyes and Underlying Interaction Mechanisms. *Int. J. Biol. Macromol.* **2021**, *183*, 399–422. [CrossRef] [PubMed]
- 15. Uddin, M.J.; Ampiaw, R.E.; Lee, W. Adsorptive Removal of Dyes from Wastewater Using a Metal-Organic Framework: A Review. *Chemosphere* **2021**, *284*, 131314. [CrossRef]
- Oyewo, O.A.; Elemike, E.E.; Onwudiwe, D.C.; Onyango, M.S. Metal Oxide-Cellulose Nanocomposites for the Removal of Toxic Metals and Dyes from Wastewater. *Int. J. Biol. Macromol.* 2020, 164, 2477–2496. [CrossRef]
- 17. Bhattacharjee, C.; Dutta, S.; Saxena, V.K. A Review on Biosorptive Removal of Dyes and Heavy Metals from Wastewater Using Watermelon Rind as Biosorbent. *Environ. Adv.* **2020**, *2*, 100007. [CrossRef]
- Akbari, A.; Sabouri, Z.; Hosseini, H.A.; Hashemzadeh, A.; Khatami, M.; Darroudi, M. Effect of Nickel Oxide Nanoparticles as a Photocatalyst in Dyes Degradation and Evaluation of Effective Parameters in Their Removal from Aqueous Environments. *Inorg. Chem. Commun.* 2020, 115, 107867. [CrossRef]
- Ihsanullah, I.; Jamal, A.; Ilyas, M.; Zubair, M.; Khan, G.; Atieh, M.A. Bioremediation of Dyes: Current Status and Prospects. J. Water Process Eng. 2020, 38, 101680. [CrossRef]
- Hitam, C.N.C.; Jalil, A.A. A Review on Exploration of Fe₂O₃ Photocatalyst towards Degradation of Dyes and Organic Contaminants. J. Environ. Manag. 2020, 258, 110050. [CrossRef]
- Rahimi, B.; Rahimi, N.R.; Ebrahimi, A. Catalytic Reduction of Hazardous Acid Orange 10 Dye by BiVO₄/TiO₂ Nanocrystalline Heterojunction and Influence of Aeration, FeSO₄, H₂O₂ and FeCl₃ on Removal Efficiency: A Novel and Environmentally Friendly Process. *Arab. J. Chem.* 2022, 15, 104003. [CrossRef]
- 22. Estrada-Cabrera, E.; Torres-Ferrer, L.R.; Luna-Barcenas, G.; Ramirez-Bon, R. Cellulose Dialysis Membrane Containing Raw Clinoptilolite Enhances the Removal of Rhodamine 6G from Aqueous Solutions. *Microporous Mesoporous Mater.* **2021**, 321, 111113. [CrossRef]
- 23. Merzouk, B.; Gourich, B.; Madani, K.; Vial, C.; Sekki, A. Removal of a Disperse Red Dye from Synthetic Wastewater by Chemical Coagulation and Continuous Electrocoagulation. A Comparative Study. *Desalination* **2011**, 272, 246–253. [CrossRef]
- 24. Nataraj, S.K.; Hosamani, K.M.; Aminabhavi, T.M. Nanofiltration and Reverse Osmosis Thin Film Composite Membrane Module for the Removal of Dye and Salts from the Simulated Mixtures. *Desalination* **2009**, *249*, 12–17. [CrossRef]
- Li, X.; Liu, T.; Zhang, Y.; Cai, J.; He, M.; Li, M.; Chen, Z.; Zhang, L. Growth of BiOBr/ZIF-67 Nanocomposites on Carbon Fiber Cloth as Filter-Membrane-Shaped Photocatalyst for Degrading Pollutants in Flowing Wastewater. *Adv. Fiber Mater.* 2022, 4, 1620–1631. [CrossRef]

- 26. Zhai, H.; Liu, Z.; Xu, L.; Liu, T.; Fan, Y.; Jin, L.; Dong, R.; Yi, Y.; Li, Y. Waste Textile Reutilization Via a Scalable Dyeing Technology: A Strategy to Enhance Dyestuffs Degradation Efficiency. *Adv. Fiber Mater.* **2022**, *4*, 1595–1608. [CrossRef]
- Al-Bastaki, N. Removal of Methyl Orange Dye and Na₂SO₄ Salt from Synthetic Waste Water Using Reverse Osmosis. *Chem. Eng. Process. Process Intensif.* 2004, 43, 1561–1567. [CrossRef]
- Yuan, J.; Chen, Z.; Yu, Q.; Zhu, W.; Li, S.; Han, L.; Lu, X.; Li, S.; Wu, Y.; Lv, Z.; et al. Enhanced Electrochemical Removal of Dye Wastewater by PbO₂ Anodes Using Halloysite Nanotubes with Different Surface Charge Properties. *J. Electroanal. Chem.* 2022, 923, 116816. [CrossRef]
- Bustos-Terrones, Y.A.; Hermosillo-Nevárez, J.J.; Ramírez-Pereda, B.; Vaca, M.; Rangel-Peraza, J.G.; Bustos-Terrones, V.; Rojas-Valencia, M.N. Removal of BB9 Textile Dye by Biological, Physical, Chemical, and Electrochemical Treatments. *J. Taiwan Inst. Chem. Eng.* 2021, 121, 29–37. [CrossRef]
- Abdelrahman, E.A.; Hegazey, R.M.; Ismail, S.H.; El-Feky, H.H.; Khedr, A.M.; Khairy, M.; Ammar, A.M. Facile Synthesis and Characterization of β-Cobalt Hydroxide/Hydrohausmannite/Ramsdellitee/Spertiniite and Tenorite/Cobalt Manganese Oxide/Manganese Oxide as Novel Nanocomposites for Efficient Photocatalytic Degradation of Methylene Blue Dye. *Arab. J. Chem.* 2022, 15, 104372. [CrossRef]
- Hegazey, R.M.; Abdelrahman, E.A.; Kotp, Y.H.; Hameed, A.M.; Subaihi, A. Facile Fabrication of Hematite Nanoparticles from Egyptian Insecticide Cans for Efficient Photocatalytic Degradation of Rhodamine B Dye. J. Mater. Res. Technol. 2020, 9, 1652–1661. [CrossRef]
- 32. Alharbi, A.; Abdelrahman, E.A. Efficient Photocatalytic Degradation of Malachite Green Dye Using Facilely Synthesized Hematite Nanoparticles from Egyptian Insecticide Cans. *Spectrochim. Acta-Part A Mol. Biomol. Spectrosc.* **2020**, 226, 117612. [CrossRef]
- Abdelrahman, E.A.; Hegazey, R.M.; Kotp, Y.H.; Alharbi, A. Facile Synthesis of Fe₂O₃ Nanoparticles from Egyptian Insecticide Cans for Efficient Photocatalytic Degradation of Methylene Blue and Crystal Violet Dyes. *Spectrochim. Acta-Part A Mol. Biomol. Spectrosc.* 2019, 222, 117195. [CrossRef] [PubMed]
- Abdelrahman, E.A.; Hegazey, R.M.; El-Azabawy, R.E. Efficient Removal of Methylene Blue Dye from Aqueous Media Using Fe/Si, Cr/Si, Ni/Si, and Zn/Si Amorphous Novel Adsorbents. J. Mater. Res. Technol. 2019, 8, 5301–5313. [CrossRef]
- Abdelrahman, E.A. Synthesis of Zeolite Nanostructures from Waste Aluminum Cans for Efficient Removal of Malachite Green Dye from Aqueous Media. J. Mol. Liq. 2018, 253, 72–82. [CrossRef]
- 36. Abu-Nada, A.; Abdala, A.; McKay, G. Removal of Phenols and Dyes from Aqueous Solutions Using Graphene and Graphene Composite Adsorption: A Review. J. Environ. Chem. Eng. 2021, 9, 105858. [CrossRef]
- Yagub, M.T.; Sen, T.K.; Afroze, S.; Ang, H.M. Dye and Its Removal from Aqueous Solution by Adsorption: A Review. Adv. Colloid Interface Sci. 2014, 209, 172–184. [CrossRef] [PubMed]
- Shi, L.; Wang, Q.; Zhao, X.; Che, Y.; Liu, H.; Zuo, W.; Zhang, Y. The Methyl Blue Adsorption Performance and Mechanism of NaX Zeolite Synthesized from Huadian Oil Shale Ash. J. Taiwan Inst. Chem. Eng. 2023, 147, 104904. [CrossRef]
- Al-Jubouri, S.M.; Al-Jendeel, H.A.; Rashid, S.A.; Al-Batty, S. Green Synthesis of Porous Carbon Cross-Linked Y Zeolite Nanocrystals Material and Its Performance for Adsorptive Removal of a Methyl Violet Dye from Water. *Microporous Mesoporous Mater*. 2023, 356, 112587. [CrossRef]
- Aboelfetoh, E.F.; El-Attar, H.G.; Okba, E.A. Facile Synthesis of Magnetic and Porous Zeolite/SnFe₂O₄ Nanocomposite for Cationic and Anionic Dyes Deterioration. *Microporous Mesoporous Mater.* 2023, 357, 112611. [CrossRef]
- 41. Ji, Y.; Xu, F.; Wei, W.; Gao, H.; Zhang, K.; Zhang, G.; Xu, Y.; Zhang, P. Efficient and Fast Adsorption of Methylene Blue Dye onto a Nanosheet MFI Zeolite. *J. Solid State Chem.* **2021**, 295, 121917. [CrossRef]
- He, S.; Sun, J.; Jin, X.; Chen, Q.; Wu, X.; Tian, F.; Zhang, X.; Li, P.; Sheng, H. Adsorption Enhancement of Congo Red Dye from Wastewater Based on Edamame Shell Originated Activated Carbon by the Cations: Experimental and Theoretical Studies. *Diam. Relat. Mater.* 2023, 136, 109930. [CrossRef]
- 43. Rimzim; Singh, J.; Mittal, S.; Singh, H. Robust Removal of Cationic Dyes by Zinc Ferrite Composites in Single and Ternary Dye Systems. *Inorg. Chem. Commun.* **2023**, *153*, 110756. [CrossRef]
- 44. Salama, H.E.; Abdel Aziz, M.S. Non-Toxic Chitosan-Pyrazole Adsorbent Enriched with Greenly Synthesized Zinc Oxide Nanoparticles for Dye Removal from Wastewater. *Int. J. Biol. Macromol.* **2023**, 241, 124632. [CrossRef] [PubMed]
- 45. Inamdar, A.K.; Hulsure, N.R.; Kadam, A.S.; Thabet, A.E.; Shelke, S.B.; Inamdar, S.N. Cobalt Oxide Nanoparticles by Flame Pyrolysis for Efficient Removal of Mixed Dyes. *Mater. Today Proc.* **2023**, in press. [CrossRef]
- Brandão, W.Q.; Maciel, B.G.; Lima, E.M.d.A.; Mojica-Sánchez, L.C.; da Silva, R.J.; de Melo, C.P. Carboxymethylcellulose Magnetic Composite for Adsorptive Removal of Cationic Toluidine Blue Dye. *Mater. Chem. Phys.* 2023, 303, 127782. [CrossRef]
- Oviedo, L.R.; Oviedo, V.R.; Dalla Nora, L.D.; da Silva, W.L. Adsorption of Organic Dyes onto Nanozeolites: A Machine Learning Study. Sep. Purif. Technol. 2023, 315, 123712. [CrossRef]
- 48. Wen, Y.; Xie, Z.; Xue, S.; Long, J.; Shi, W.; Liu, Y. Preparation of Benzenesulfonyl Hydrazone Modified Guar Gum and Its Adsorption Properties for Dyes and Phytotoxicity Assays. *Int. J. Biol. Macromol.* **2023**, 234, 123700. [CrossRef]
- Gollakota, A.R.K.; Munagapati, V.S.; Gautam, S.; Wen, J.C.; Shu, C.M. Hydrothermal Tuning of Morphology of Aluminophosphate (AlPO-14) Framework for the Adsorption of Rhodamine 6G Dye. *Adv. Powder Technol.* 2021, 32, 3002–3015. [CrossRef]
- 50. Vanamudan, A.; Bandwala, K.; Pamidimukkala, P. Adsorption Property of Rhodamine 6G onto Chitosan-g-(N-Vinyl Pyrrolidone)/Montmorillonite Composite. *Int. J. Biol. Macromol.* **2014**, *69*, 506–513. [CrossRef]

- 51. Annadurai, G.; Juang, R.S.; Lee, D.J. Adsorption of Rhodamine 6G from Aqueous Solutions on Activated Carbon. J. Environ. Sci. Heal.-Part A Toxic/Hazardous Subst. Environ. Eng. 2001, 36, 715–725. [CrossRef] [PubMed]
- Gollakota, A.R.K.; Volli, V.; Munagapati, V.S.; Wen, J.C.; Shu, C.M. Synthesis of Novel ZSM-22 Zeolite from Taiwanese Coal Fly Ash for the Selective Separation of Rhodamine 6G. J. Mater. Res. Technol. 2020, 9, 15381–15393. [CrossRef]
- Zheng, X.; Zheng, H.; Zhou, Y.; Sun, Y.; Zhao, R.; Liu, Y.; Zhang, S. Enhanced Adsorption of Orange G from Aqueous Solutions by Quaternary Ammonium Group-Rich Magnetic Nanoparticles. *Colloids Surfaces A Physicochem. Eng. Asp.* 2019, 580, 123746. [CrossRef]
- 54. Arulkumar, M.; Sathishkumar, P.; Palvannan, T. Optimization of Orange G Dye Adsorption by Activated Carbon of Thespesia Populnea Pods Using Response Surface Methodology. *J. Hazard. Mater.* **2011**, *186*, 827–834. [CrossRef]
- 55. Banerjee, S.; Dubey, S.; Gautam, R.K.; Chattopadhyaya, M.C.; Sharma, Y.C. Adsorption Characteristics of Alumina Nanoparticles for the Removal of Hazardous Dye, Orange G from Aqueous Solutions. *Arab. J. Chem.* **2019**, *12*, 5339–5354. [CrossRef]
- 56. Imgharn, A.; Anchoum, L.; Hsini, A.; Naciri, Y.; Laabd, M.; Mobarak, M.; Aarab, N.; Bouziani, A.; Szunerits, S.; Boukherroub, R.; et al. Effectiveness of a Novel Polyaniline@Fe-ZSM-5 Hybrid Composite for Orange G Dye Removal from Aqueous Media: Experimental Study and Advanced Statistical Physics Insights. *Chemosphere* 2022, 295, 133786. [CrossRef]
- 57. Shojaei, B.; Miri, R.; Bazyari, A.; Thompson, L.T. Asphaltene Adsorption on MgO, CaO, SiO₂, and Al₂O₃ Nanoparticles Synthesized via the Pechini-Type Sol–Gel Method. *Fuel* **2022**, *321*, 124136. [CrossRef]
- 58. Kang, J.; Gwon, Y.R.; Cho, S.K. Photoelectrochemical Water Oxidation on PbCrO₄ Thin Film Photoanode Fabricated via Pechini Method: Various Solution-Processes for PbCrO₄ Film Synthesis. *J. Electroanal. Chem.* **2020**, *878*, 114601. [CrossRef]
- Alhaji, A.; Razavi, R.S.; Ghasemi, A.; Loghman-Estarki, M.R. Modification of Pechini Sol–Gel Process for the Synthesis of MgO-Y₂O₃ Composite Nanopowder Using Sucrose-Mediated Technique. *Ceram. Int.* 2017, 43, 2541–2548. [CrossRef]
- 60. Kumar, V.; Sharma, R.; Kumar, S.; Kaur, M.; Sharma, J.D. Enhancement in the Photocatalytic Activity of Bi₂Ti₂O₇ Nanopowders Synthesised via Pechini vs Co-Precipitation Method. *Ceram. Int.* **2019**, *45*, 20386–20395. [CrossRef]
- Shah, R.K.; Naglah, A.M.; Al-Omar, M.A.; Almehizia, A.A.; AlReshaidan, S.; Subaihi, A.; Alharbi, A.; Hameed, A.M.; Alkabli, J.; Fetoh, M.E.; et al. Efficient Removal of Ni(II) Ions from Aqueous Solutions Using Analcime Modified with Dimethylglyoxime Composite. Arab. J. Chem. 2021, 14, 103197. [CrossRef]

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